# Sheboygan River and Harbor Superfund Site 

Baseline Upper and Lower River Fish Monitoring Report

Prepared for
United States Environmental Protection Agency Region 5

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### 1.0 Introduction

Monitoring of post-remedial fish tissue concentrations of polychlorinated biphenyls (PCBs) is being conducted on the Sheboygan River in accordance with the Post-Remediation Monitoring Plan (PMP). As stated in the PMP, the monitoring is being conducted in three phases consisting of the following:

- Baseline monitoring after remediation of the Upper River and prior to remediation of the Lower River reaches to determine the mean PCB concentration of each fish species of interest and establish a comparison point for future sampling, ${ }^{1}$
- Phase 1 annual monitoring following remediation of each reach to determine if the PCB concentration of each fish species is changing compared to the baseline and track the progress of the fish in meeting the remedial goals, and
- Phase 2 conformational sampling to verify the fish have reached the remedial goals.

This Baseline Upper and Lower River Fish Monitoring Report documents the post-remediation monitoring performed in 2008, specifically the collection of fish to establish baseline concentrations of several different fish species downstream of the portion of the river known as the Upper River. Baseline fish monitoring for the Upper River is considered the first annual sampling event following remediation documenting post-remedial conditions.

The data obtained during the baseline fish sampling will allow post-remedial fish tissue concentrations to be compared to baseline results to monitor remedial progress. Fish tissue results in the Upper River will be compared to the baseline fish monitoring performed in the first annual sampling event post-remediation, and the 2002 Interim Monitoring Program (IMP) Report. Fish tissue results in the Lower River reaches will be compared to the baseline fish monitoring performed prior to remediation.

In accordance with the Upper River Statement of Work (URSOW), post-remedial monitoring will occur until fish consumption advisories for the Upper River based on PCBs are lifted by the Wisconsin Department of Health, fish fillet concentrations of PCBs decrease to the target levels specified on page 32 of the Record of Decision (ROD), or for 30 years, whichever comes first.

[^0]
### 1.1 Site Description

The Sheboygan River and Harbor Superfund Site (the Site) is located on the western shore of Lake Michigan approximately fifty-five miles north of Milwaukee, Wisconsin, in Sheboygan County (Figure 1). The Site includes the former Tecumseh Manufacturing site and the lower fourteen miles of the Sheboygan River from the Sheboygan Falls Dam downstream to, and including, the Inner Harbor. This segment of the river flows west to east through the cities of Sheboygan Falls, Kohler, and Sheboygan before entering Lake Michigan.

During the Remedial Investigations (RI), the river was segmented in separate sections, known as reaches, based on physical characteristics such as average depth, width, and level of polychlorinated biphenyl (PCB) sediment contamination. The Upper River extends from the Sheboygan Falls Dam downstream four miles to the Waelderhaus Dam in Kohler. The Middle River extends seven miles from the Waelderhaus Dam to the former Chicago \& Northwestern (C\&NW) railroad bridge. The Lower River extends three miles from the C\&NW railroad bridge to the Pennsylvania Avenue Bridge in downtown Sheboygan. The Inner Harbor includes the Sheboygan River from the Pennsylvania Avenue Bridge to the river's outlet to the Outer Harbor. The Outer Harbor is defined as the area formed by the two break walls. Figure 2 provides an overview of each river reach.

Remedial Design (RD) and Remedial Action (RA) work at the Site has been phased in order to achieve proper source control prior to beginning down river work. Phase I RA work for the Upper River, which included the Tecumseh plant soils, groundwater, and adjoining riverbank soils was completed in 2004. Phase II RA work for the Upper River included addressing the Near-Shore Sediments, Armored Areas, and Soft Sediment deposits was completed in 2007. ${ }^{2}$ The Upper River floodplains have not been addressed due to access limitations. Remedial work in the Lower River has not been implemented.

## $1.2 \quad$ Site History

The following information was obtained from the ROD. The Sheboygan Harbor was constructed at the mouth of the Sheboygan River in the early 1920's. In 1954, the lower Sheboygan River, namely the channel upstream of the $8^{\text {th }}$ Street Bridge, was added as a part of the United States Army Corps of Engineers (USACE) maintenance dredging. Between 1956 and 1969, a total of 404,000 cubic yards of sediment were removed downstream of the $8^{\text {th }}$ Street bridge between 1956 and 1969. The portion of the river above the $8^{\text {th }}$ Street Bridge has not been dredged since 1956.

[^1]Prior to 1969, the USACE disposed of the sediment from the Harbor in an authorized deep water disposal area in Lake Michigan. However, there has been no dredging in the Sheboygan Harbor since the United States Environmental Protection Agency (USEPA) and Wisconsin Department of Natural Resources (WDNR) determined that the sediment was unsuitable for open-water disposal. Sediment sampling and analysis performed by the USACE in 1979 detected what was reported as moderate to high levels of lead, zinc, PCBs, and chromium. According to the ROD, the USACE last dredged the Harbor mouth in 1991 however; in 1982 a policy to discontinue maintenance dredging was promulgated due to the discovery of PCBs in the sediments.

In June 1979, the USACE collected 11 cores from the Harbor area ranging in depth from 1.5 to 9 feet. The analytical results revealed greater PCB and metal levels in the sediment of the Inner Harbor than in sediment of the Outer Harbor. In October 1979, the USACE collected a second round of 21 cores. The analytical results indicated an increase in PCB concentrations with the distance upstream from the Harbor and with the depth of sediment.

Examination of 98 sediment profile samples collected by the USACE in December, 1982 from the Harbor indicated the presence of PCBs in the surface sediment of the Harbor. The possibility that this sediment may be classified as regulated material was reason for discontinuing maintenance dredging.

Tecumseh Products Company (Tecumseh) was located adjacent to the Sheboygan River in Sheboygan Falls and operated from 1966 to 2003. Tecumseh was considered a Potentially Responsible Party (PRP) when PCBs were discovered in coolant fluids disposed to sewer lines that discharged to the Upper River reach of the Sheboygan River. The contamination level was high in the sediment adjacent to the Tecumseh Plant, but decreased in concentration downstream. Tecumseh discontinued use of PCB impregnated coolant fluids in the early 1970's.

In 1978, the WDNR conducted a survey and found numerous industries that discharged contaminants to the Sheboygan River. Some had levels of PCBs discharged to the river and others had heavy metals in their discharge. In 1985, the outfall from Thomas Industries, located along the Inner Harbor, contained PCBs when analyzed by the WDNR on two different dates. A sample collected on June 13, 1975, from the storm sewer outfall had a concentration of 125 parts per billion (ppb) PCBs. A second sample collected on August 19, 1975, had a PCB concentration of 88 ppb . The Kohler Company, downstream of Sheboygan Falls and adjacent to the Middle River, was found to have heavy metal discharges to the river above the permit limits in the 1970s. In addition, the Kohler Landfill Superfund Site is located on the banks of the river.

The USEPA placed the Sheboygan River and Harbor Site on the National Priorities List (NPL) in 1986. Remedial work performed since that time includes source removal at the former Tecumseh property and removal of $94.1 \%$ of the impacted sediment in the Upper River. This work was completed in 2007.

### 1.3 River Characteristics

### 1.3.1 Upper River

The Upper River consists of discrete Soft Sediment deposits and non-Soft Sediment areas which include a mix of Soft Sediment, rocks, cobbles, and bare river bottom. The sediment contamination in the Upper River acts as a partial source of PCB-contaminated sediment for the rest of the river system during high river conditions in addition to the other sources identified in the Middle River, Lower River, and Inner Harbor. PCB sampling results in 1989 and 1990 showed concentrations from 1.4 to $4,500 \mathrm{ppm}$. PCB-contaminated sediment was removed near the former Tecumseh facility in 1990 and 1991. Subsequent sampling of the same area showed concentrations ranging from non-detect to as high as 840 ppm . The concentrations of PCBs in the sediment vary due to the dynamic nature of this river reach.

During the 2006/2007 seasons, sediment was removed from nine (9) Armored Area Remedial Management Units (RMUs) and 122 Soft Sediment RMUs. The Soft Sediment RMUs and Armored Areas removed in 2006/2007 contained the majority of the PCB mass within the Upper River. The Upper River remedial action conducted in 2006 and 2007 removed 20,728 cubic yards of sediment and 552 pounds of PCBs for a total mass removal percentage of $94.1 \%$ exceeding the PCB mass reduction objective of $88 \%$. The Upper River SWAC was reduced from 5.2 ppm to 1.96 ppm and based on the mass removed, should reach a SWAC of 0.5 ppm over time.

### 1.3.2 Middle River

The Middle River consists of Soft and non-Soft Sediment areas similar to the Upper River, but due to the hydrodynamics of this reach, the areas of Soft Sediment are shallower and more widely scattered. The Waelderhaus dam, which marks the end of the Upper River, prevents most of the Upper River sediments from migrating downstream. As such, the Middle River sediments act as the primary source of PCB-contamination for the rest of the Lower River system. Information collected during the Remedial Investigation (RI) indicated PCB concentrations ranging from non-detect to 8.8 parts per million (ppm). WDNR sediment trap data, between 1990 and 1996, showed PCB concentrations ranging from 1.4 to 3.0 ppm . Samples obtained by the WDNR in 1997 indicated PCB concentrations ranging from 0.6 ppm to 37 ppm . Like the Upper River, sediment in the Middle River is likely to vary due to the dynamic nature of this river reach.

### 1.3.3 Lower River

The flow rate in the Lower River decreases leading to a more continuous layer of Soft Sediment throughout the reach. Based on the hydrodynamics of this reach, the Lower River is where much of the sediment released in the Middle River is deposited. During the RI, sample results showed PCB concentrations as high as 67 ppm adjacent to the WPSC Camp Marina MGP site, a site undergoing investigation and remediation under the oversight of the USEPA. WDNR sediment trap data, from 1994 to 1996 , showed PCB concentrations ranging from 1.9 to 4.2 ppm .

### 1.3.4 Inner Harbor

The Inner Harbor is generally the river reach where upstream Soft Sediment is deposited. However, while the Inner Harbor is generally depositional, deposition occurs primarily between the $8^{\text {th }}$ Street Bridge and the harbor mouth. The area between the Pennsylvania Bridge and $8^{\text {th }}$ Street Bridge has little deposition and shows evidence of scour. RI sampling indicated PCB concentrations as high as 220 ppm in the Inner Harbor; however these levels were detected in 1979 and exist many feet below the surface. Surface ( $0-6$ inches) sampling conducted in 1987 showed PCB results ranging from 0.17 to 5.8 ppm . Surface ( $0-6$ inches) sampling conducted in 1999 showed PCB results ranging from 0.38 to 5.3 ppm . As a general rule, PCB concentrations increase with depth between the $8^{\text {th }}$ Street bridge and harbor mouth. This is not the case for certain areas between the Pennsylvania Bridge and $8^{\text {th }}$ Street Bridge.

### 1.4 Summary of Previous Fish Species Evaluation

This section is provided to demonstrate how sediment cleanup goals were established. The consumption of the fish is the primary exposure route for human receptors of the PCBs in the river sediments. The PCBs in the river sediments bioaccumulate in the fish from contact with impacted sediment, surface water, or by ingesting prey that are impacted. An understanding of the process in developing the sediment PCB cleanup goals based on allowable fish PCB concentrations is important in the evaluation of long-term assessment of remedial success.

There is considerable seasonal fishing in the Middle River, Lower River, and Inner Harbor. ${ }^{3}$ Fishing is more limited in the Upper River. According to WDNR surveys, most fishing occurs during spring and fall salmon and trout runs. Resident fish taken from the Sheboygan River, between the Sheboygan Falls dam and the mouth of the river, fall into the "do not eat" consumption advisory category. Migrating trout and salmon are subject to Lake Michigan advisories as they obtain most of their PCB body burden from Lake Michigan. One objective of the sediment removal is to reduce the concentrations of PCBs in the fish over time so all the consumption advisories are lifted.

The physical setting of the Site provides several possible pathways of exposure to the contamination in the sediment: dermal contact, ingestion of contaminated surface water or sediment, and consumption of fish contaminated by sediment. The sediments are contaminated with PCBs, hydrophobic organic compounds that will strongly prefer to partition to organic material. It is assumed that the most significant exposure is from contaminated sediment, where virtually all PCBs reside, and not the surface water. In general, there is likely to be only limited direct contact with the sediment itself (i.e., dermal and/or ingestion pathway). Many studies have found that bioaccumulation of hydrophobic organic sediment contaminants is the critical and dominant fate of these compounds in the environment. As such, the human health analysis assumed that for this Site, the pathway presenting the majority of the risk and likely to yield the most protective assessment of risks is consumption of contaminated fish and not dermal contact. This does not imply that no other exposure pathways are occurring at this site, only that there is a focus on the pathway which contributes the majority of risk, the fish ingestion pathway.

[^2]Tecumseh collected fish tissue samples between 1990 and 1998 that showed smallmouth bass and white sucker PCB concentrations ranging from 1.3 ppm to 23.1 ppm . Carp had PCB levels ranging from 10.5 to 200 ppm . In general, the highest fish tissue PCB concentrations were found nearest the Tecumseh plant and tended to decrease downstream. The most recent studies by WDNR found that carp and smallmouth bass had the following mean concentrations, respectively: ${ }^{4}$

- Upper River
- Middle River
- Lower River
- Inner Harbor
16.43 and 0.44 ppm
12.5 and 2.73 ppm
2.32 and 1.35 ppm , and
1.45 and 2.0 ppm .

An Interim Monitoring Program (IMP) was performed by Blasland, Bouck, and Lee, Inc. (BBL) that consisted of the collection of smallmouth bass and white suckers at Rochester Park in the Upper River reach and between the dams in the Upper River reach. ${ }^{5}$ During the baseline and subsequent post-remedial monitoring, these areas are known as Upper River 1 and Upper River 2 Sites. These fish were also collected near Kiwanis Park or in the Lower River reach. The range of smallmouth bass PCB concentrations detected is as follows:

- Upper River 1
2.1 to 10.3 ppm
- Upper River 2
1.1 to 7.3 ppm , and
- Lower River 0.82 to 3.7 ppm .

The PCB concentration decreased between 1990 and 2002 as seen in Charts 2 and 3 of Appendix 3. The results for smallmouth bass in the Upper River Site 1 show a general decreasing trend and the regression shows a decrease with a moderate correlation. For Upper River Site 2, the decrease has a very strong correlation for the regression. The range of white sucker concentrations detected is as follows:

- Upper River $1 \quad 2.7$ to 18.3 ppm
- Upper River $2 \quad 1.9$ to 8.7 ppm , and
- Lower River
1.4 to 3.9 ppm .

These PCB concentrations also decreased between 1990 and 2002 based on a comparison of the 2002 result to the 1990 result. While a regression of all the data between this period indicates a slight increase, the correlation is very weak (Chart 4, Appendix 3).

In 1996, the United States Environmental Protection Agency (USEPA) performed a baseline risk assessment for the Site, relying on data available from WDNR on fish tissue concentrations from 1994. The USEPA assessed sport fishing and subsistence fishing. The sport fishing scenario

[^3]was developed to represent a mid-point or central tendency estimate of risk, and the subsistence fishing scenario was developed to represent an upper-bound estimate of risk. The sport fishing scenario variables were chosen to be reasonable, and not overly conservative in their assumptions. The USEPA used Great Lakes specific fish consumption information, available from an assessment of Michigan anglers. It was assumed that of the total amount of fish consumed; only half of the fish came from the Sheboygan River. This is accounted for in the fraction ingested term. For the upper-bound subsistence scenario, USEPA used a conservative estimate of all fish ingested coming from the Sheboygan River. Through this risk assessment, USEPA determined the following risks:

- Average
$1 \times 10^{-4}$ to $1 \times 10^{-5}$
- Subsistence

$$
1 \times 10^{-2} \text { to } 1 \times 10^{-4}
$$

In order to address unacceptable risks at the Site, USEPA calculated sediment cleanup goals, protective of human health. The USEPA made a conscious decision to model and be protective of the more contaminated resident fish species of smallmouth bass and carp at the Site. By selecting a cleanup goal protective of bass (or carp), the cleanup will be protective of the lesser contaminated species such as walleye, trout, salmon, and steelhead. This choice adds a layer of conservatism to allow for more fish consumption at the Site, especially of several non-resident species. Therefore, a cleanup based on resident species may allow for possibly more consumption of other types of fish that may occur as advisories are lifted.
) To calculate a sediment cleanup goal or surface goal, target fish tissue levels were placed into a Biota to Sediment Accumulation Factor (BSAF) equation to estimate the sediment concentrations that would meet these fish targets. The term "surface goal" is more appropriate for sediment at the Sheboygan Site than the usual cleanup goal because what is calculated is a surface that the fish can be exposed to that will result in the target fish tissue levels. Looking at the Site, it's necessary to calculate what the residual concentration is after dredging certain levels, or what's left after taking out everything above a certain concentration. In the case of the Sheboygan Site, it's the target Surface Weighted Average Concentration, or SWAC, of the river after remediation.

The BSAF methodology is the same as used in the Ecological Risk Assessment and is similar to what was used in the Remedial Investigation/Feasibility Study (RI/FS), except USEPA risk assessments include total organic carbon (TOC) and lipids in the calculation. Note that BSAFs were only calculated for smallmouth bass and carp and not the lesser contaminated migratory species of salmon and steelhead, to provide protection for anglers who consume several different species of fish. BSAFs were calculated for smallmouth bass because of their prevalence in the river and for carp as an indicator of concentrations in fish with higher lipid levels.

The analysis begins by calculating a site-specific BSAF using PCBs in sediment, TOC, PCBs in fish, and lipid data. The site-specific BSAFs are derived from the following values: RI/FS total river bed SWAC, and NOAA Risk Assessment TOC, and 1994 fish data (from FIELDS database). However, because the data in the RI/FS were given as summary statistics, the USEPA could not derive its own sediment surface area weighted PCB that is normalized to TOC. This
term is necessary for the BSAF model. Therefore, the SWAC derived in the RI/FS is not useable in calculating a site-specific BSAF. Because the NOAA ecological risk assessment for the site also developed BSAFs, USEPA considered the NOAA BSAFs, and found that they were quite similar to the human health based BSAFs. Using the BSAFs, the USEPA determined the sediment cleanup goals as follows:

$$
\text { Sediment Cleanup Goal = }(\text { TOC } \times \text { Conc. Fish }) /(\text { site specific BSAF x } \% \text { lipid })
$$

As can be seen, the sediment cleanup goal is entirely dependent on the accuracy of the BSAF. Therefore, the concentrations of PCBs in the fish may reach the target levels although the sediment contains more than the sediment cleanup goal. Conversely, the sediment cleanup goal may be reached before the fish actually reach the target levels. We have noted that prior to remediation; the PCB levels in the most recent fish collected in the Upper River as compared to the characterization sediment results have less PCBs than predicted by the BSAF. Therefore, the fish target levels may be reached before the sediment cleanup goals.

Target fish tissue levels corresponding to the SWAC Sediment Cleanup Goal include the following:

- Smallmouth Bass $\quad 0.31 \mathrm{ppm}$ (skin on fillet)
- Walleye $\quad 0.63 \mathrm{ppm}$ (skin on fillet)
- Trout
0.09 ppm (skin on fillet) ${ }^{6}$
- Carp
2.58 ppm (skin on fillet)
- Catfish
2.53 ppm (skin off fillet)

Using the BASF and these goals, the USEPA determined that the sediment cleanup goal SWAC is 0.5 ppm . The USEPA model predicts that once the SWAC reaches 0.5 ppm , the fish target levels will be met. ${ }^{7}$ However, as the sediment cleanup goal was determined by modeling, the fish could reach the goals before the SWAC is 0.5 ppm . Conversely, the SWAC could reach 0.5 ppm and the fish do not reach the goal.

[^4]
### 2.0 Sampling and Analysis

### 2.1 Summary of Baseline Sampling Plan

The baseline sampling and analysis of fish species was conducted consistent with the Post Remedial Monitoring Plan (PMP) and the Quality Assurance Project Plan (QAPP). These plans were conditionally approved with comment on August 13, 2008. The PMP, which was developed with assistance from WDNR and the USEPA, determined statistically the number of fish to collect in each reach as well as in two sites within both the Upper and Middle River reaches.

Smallmouth bass, carp, walleye, and catfish were selected as they have assigned target goals in the Record of Decision (ROD). According to the ROD, smallmouth bass and carp are the more contaminated resident fish species at the Site and the USEPA selected these fish to determine cleanup goals believing that if these fish met the goals, the lesser contaminated species such as walleye, trout, salmon, and steelhead would be protected. Therefore the monitoring included these fish as well as walleye and catfish. Walleye and smallmouth bass will also help evaluate risk reduction for sport fisherman while carp and catfish for sustenance fisherman.

Rock bass and longnose dace were added because catfish and walleye are rarely caught according to WDNR. Juvenile carp and white suckers were added at the suggestion of the WDNR. Initially, the draft PMP that was approved stated that "carp or white suckers" were to be caught. After realizing this may not lead to a statistically valid sample set, WDNR and Pollution Risk Services (PRS) decided that both should be collected and the final PMP was written accordingly. The following table outlines the final fish species collection requirements.

| Fish Species (size) | Number of Samples Per River Reach |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Upper <br> (Site 1) | Upper <br> (Site 2) | Middle <br> (Site 1) | Middle <br> (Site 2) | Lower | Inner <br> Harbor |
| Smallmouth Bass <br> (10-17") | 8 | 8 | 8 | 8 | 8 | 8 |
| Adult Carp <br> $\left(15-25^{\prime \prime}\right)$ | 16 | 16 | 8 | 8 | 8 | 8 |
| Juvenile Carp <br> $\left(3-8^{\prime \prime}\right)$ | 16 | 16 | 8 | 8 | 8 | 8 |
| Adult Suckers <br> $\left(8-16^{\prime \prime}\right)$ | 8 | 8 | 8 | 8 | 8 | 8 |
| Juvenile Suckers <br> $\left(3-8^{\prime \prime}\right)$ | 8 | 8 | 8 | 8 | 8 | 8 |
| Rock Bass <br> $\left(5-9^{\prime \prime}\right)$ | 8 | 8 | 8 | 8 | 9 | 9 |
| Longnose Dace <br> $\left(1-4^{\prime \prime}\right)$ | 8 | 8 | 8 | 8 | 8 | 8 |
| Walleye <br> $\left(12-22^{\prime \prime}\right)$ | 8 | 8 | 8 | 8 | 9 | 9 |
| Catfish <br> $\left(12-22^{\prime \prime}\right)$ | 8 | 8 | 8 | 8 | 8 | 8 |

The WDNR requested that the Upper and Middle River be divided into two sites per reach. The rational was stated as "Sampling stations should include the following number of sites per reach in order to represent the amount of contaminated sediment that will be removed and the variability expected. Specimens may be collected at different locations within a reach and collections sites within a reach can vary in exact location and length of river sampled (distance and location data should be reported in annual reports):"

As such, the collection included two sites in the Upper River - one from the former Tecumseh facility to River Bend reach and another from the Riverbend to Waelderhaus Dam in Kohler. For the Middle River, fish were collected from two sites within the reach: between the Waelderhaus dam and the Kohler landfill and downstream of the Kohler landfill to the C\&NW Railroad Bridge.

The fish collection would target the habitats most conducive for each species. Table 1 presents a summary of the fish species, known habitat, and range. This information was primarily obtained from Fishes in Wisconsin (1983) and is intended to provide a summary of the characteristics of the target species and their typical habitat and is not intended to describe the habitats where the target species where encountered in the Sheboygan River. The habitats where fish were collected in 2008 are shown in Figures 3 through 6.

### 2.2 Baseline Procedures

After receipt of the Scientific Collectors permit on August 19, 2008, collection began in the Upper River reach before generally proceeding in order to the Lower River, Inner Harbor, and finally, the Middle River reaches. Due to an inability to initially collect Longnose Dace and juvenile species, the Upper and Middle River reaches were revisited. The fish collection occurred between August 19, 2008, and September 17, 2008. Table 2 provides a summary of the daily fish collection. Figures 3 through 6 show the locations where fish were collected in each reach.

With one exception, all fish were collected using electro-fishing equipment. The electro-fishing equipment used to collect fish, a Smith Root, Inc. Model 2.5 GPP , was either a boat-mounted array set-up or a hand held wand, depending on the location and species to collect. Due to the inability to obtain longnose dace with this method, seining was employed for this species. Electro-fishing was performed by selecting the appropriate pulsed DC power setting to stun-fish. The appropriate DC pulse setting ( 30 or 60 ) was made based on what set-up was used ( 30 for the wand, 60 for the arrays). At that point the percentage of output power was adjusted from 0-100 to stun the fish size needed without stunning more fish than needed or killing the fish. This percentage was determined by trial and error. Current was then applied to the river water by closure of the operating switch (i.e. foot pedal) while the generator and control equipment were operative. Once fish were stunned, the fish were collected with dip nets. The fish collected in the dip nets were identified for targeted species, measured to confirm they met size requirement, and were either retained in a live well or on ice in an insulated cooler until collection was completed.

Both shore and quarter arc seining was performed to collect Longnose Dace (dace). To collect dace, a seine with dimensions and mesh size appropriate for the dace and collecting conditions was selected ( 20 ' long with $1 / 4$ " mesh). For shore seining, the seining was performed by maintaining the seine approximately perpendicular to a shoreline, with one end at or near the edge of the water and the other held out as far out from shore as practicable. The seine was pulled along the shore with both ends moving at about the same rate. At the end of the seine haul, the outer end was moved around to the shore, and the entire seine was pulled out of the water while maintaining the leadline on the bottom as practicable. The seine was pulled onto shore until the leadline was completely out of the water.

For quarter-arc seining, the seining was performed by holding one end of the seine in one place at or near the shoreline and first pulling the other end of the seine out into the water perpendicular to the shore. The water-end of the seine was moved down and back toward shore so that the outer end of the net moves approximately through a quarter of a circle. When the outer end of the net reaches shore, the entire seine was pulled out of the water while maintaining the leadline on the bottom as practicable. The seine was pulled onto shore until the leadline was completely out of the water.

All fish samples were processed and packaged in accordance with the procedures described in the WDNR's Division of Environmental Standards Field Procedures Manual in addition to the PMP. During and after collection, samples were held in a live well or on ice in an insulated cooler. Samples remained whole and ungutted. Each fish was numbered and the following recorded in field log book:

- Length,
- Species ${ }^{8}$,
- Sex (if possible),
- Age (if possible),
- Sample location,
- Other distinguishing features,
- Sampler(s), and
- Any unusual skin lesions, tumors, or other irregularities should also be noted.

The individual fish were wrapped in aluminum foil, then in freezer paper, and finally taped securely so that the package did not open during shipment. All samples were frozen as soon as possible after collection. No composite samples were created or analyzed.

For shipment to the laboratory, all fish samples were placed in a Ziploc bag or industrial grade trash bag, a label affixed and placed into second Ziploc bag, and then into a cooler with double bagged ice on the bottom of the cooler. The cooler was filled with fish samples, leaving enough room for double bagged ice on top of samples. A chain-of-custody form was placed in a sealable plastic bag and taped to the inside of cooler lid. The coolers were collected by the laboratory and as such custody seals were not used.

[^5]The laboratory prepared and analyzed the samples in accordance with the analytical method USEPA SW846-8082 Modified and Laboratory Standard Operating Procedures (SOPs) developed in accordance with method 8082 including the following:

- GB-L-001, Rev . 0 -- Tissue Preparation
- GB-L-003, Rev. 0 - Lipids
- GB-O-031, Rev. 1 - Extraction
- GB-O-034, Rev. 1 - Sulfuric Acid Cleanup
- GB-O-036, Rev. 1 - Florosil Cleanup
- GB-O-026, Rev. 2 - PCB Analysis

The analysis to be performed on fish included total PCBs (Aroclor basis), percent lipids, and gender. The PCB method detection limit was $0.019 \mathrm{mg} / \mathrm{kg}$.

QA/QC samples consisted of a matrix spike and matrix spike duplicate. A minimum of one matrix spike/matrix spike duplicate analysis was performed with every batch of fish being analyzed for PCBs. Batch size was limited to no more than 20 samples. For analysis of PCBs in tissues, the QA procedures in USEPA's Statement of Work for Organic Analysis (Feb 1988) was used, including laboratory blanks consistent with required detection limits, and initial and continuing calibration to verify recoveries.

### 2.3 Deviation from Plan

The only field deviation was not all targeted fish were collected. Table 3 provides a summary of the success of the collection process. It was anticipated that walleye or catfish would not be collected and as surrogates, rock bass and longnose dace were used. While we did not expect to catch any walleye or catfish, some were collected. Catfish were collected from the Middle, Lower and Inner Harbor River reaches. Walleye were collected from the Middle River and Inner Harbor reach.

No juvenile carp could be obtained. According to Fishes of Wisconsin (1983), carp typically spawn in late in May or early June and the incubation period is 3 to 16 days depending on temperature. Young carp grow very rapidly and by middle August have an average size of almost four inches and a range of 3 to 5 inches. Based on this growth rate, it may be difficult to catch juvenile carp in the 3 to 8 inches range specified in the PMP in late August and early September. Earlier fish collection of juvenile carp should be considered in the future.

For adult White Suckers, the target numbers were reached at both Upper River sites and one of the Middle River site. The target goal was only missed by one fish in Middle River 1 and the number collected was similar to WDNR efforts in 1999 and 2004. Failure to collect the target goal in the Lower River ( 2 of 8 ) and in the Inner Harbor ( 0 of 8 ) is attributed to lack of habitat. Very little areas'with vegetation and warm shallows of estuaries and bays, the preferred habitat of white sucker, were observed in the Lower River and none were observed in the Inner Harbor (see Figures 5 and 6). Information on habitat was obtained from Fishes of Wisconsin (1983).

WDNR has also not had much success collecting this species in the Lower River or Inner Harbor reaches. Failure to collect the target goal of juvenile White Suckers is also attributed to lack of habitat in the Lower River and Inner Harbor.

Finally, the Sheboygan River does not appear to provide an abundance of quality habitat for Longnose Dace being too deep in many areas. However, there is some suitable habitat where shallows are present (i.e. Upper River, Site 1 and Middle River, Site 1 and 2). The water is too deep in the Lower River and Inner Harbor reaches to provide suitable habitat. It is also unsuitable in Site 2 of the Upper River reach. The baseline collection obtained $61 \%$ of the expected target goal. Based on the results as compared to habitat requirements, the goal of collecting certain fish in certain locations was optimistic at best. If the completion success is based on a target goal limited to the reaches conducive to dace, a $65 \%$ completion percentage was obtained. For the adult fish in the ROD that were expected to be caught, carp, suckers and smallmouth bass, the success rate is $76 \%$.

The inability to collect the target number of fish for some of the species can increase the chances of a Type II error. That is, believing the fish tissue PCB results are less than the action level when they are not. Reducing the number of samples reduces the confidence in the decision. This is the baseline sampling event and this decision is not being made. As such, this error cannot occur.

There were no deviations from the laboratory method in order to analyze or report the fish tissue results.

### 3.0 Sampling Results

### 3.1 Fish Tissue Results

A summary of the results is provided in Appendix 1 while copies of the analytical reports are provided in Appendix 2 as a compact disc. Except for catfish, all fish samples that were analyzed were skin on fillets. Catfish samples analyzed were skin off fillets. A summary of the baseline statistics is provided in Table 4. The adult fish tissue PCB results tend to decrease moving from the Upper River to the Inner Harbor. An exception is that in almost every case, the PCB concentrations were higher in the Lower River reach than the Middle River 2 site. This would correspond to the increase in PCBs in the sediment in the Lower River and Inner Harbor due to the identified sources in these reaches. Chart 1 in Appendix 3 provides a graphical summary of the PCB concentrations of the adult fish that were most successfully collected across reaches demonstrating the decreasing trend from upstream to downstream

Adult carp tended to have the highest mean PCB concentrations of the fish species, due to being the most prevalent species collected. Although for the few caught, catfish had the highest mean concentration. These are bottom feeders and the results are not unexpected compared to the sport fish. As will be discussed in the following section, the results are higher than the most recent Interim Monitoring results. They are also higher than the older results from the Interim Monitoring Program. Adult carp had the highest mean concentration in the Upper River. However, in both sites of the Middle River, as well as the Lower River and Inner Harbor reaches, this was the only fish caught that many of the individual results were less than the ROD goal.

The age of the fish was determined by EA Engineering, Science, and Technology, Inc. who performs fish aging for the Fox River monitoring program and was recommended by Foth Infrastructure and Engineering LLC (Foth). All of the adult fish were of the age where they should have been sexually mature. None of the fish collected appeared to be of an age that exceeded the usual published longevity period. The majority of the fish collected were males.

### 3.2 Data Quality

The laboratory performs a validation of the analytical procedure using the quality control sample results, as applicable. This validation is discussed in the Narrative section of each of the 13 lab reports generated by this sampling and analysis event. The laboratory reported the following:

- All samples were extracted and analyzed within the allowable holding time,
- There were no problems with the initial or continuing calibrations,
- There were no problems with duplicate samples,
- All laboratory control spikes were within the allowable range, and
- PCBs were not detected in the method blanks.

There were problems with the surrogate recoveries in $36 \%$ of the samples. The problem was that the surrogates could not be evaluated against the control limits due to sample dilution. This should not affect the data as for the $64 \%$ that could be compared, there were no problems.

There were 9 occasions where the laboratory identified problems with the matrix spike (MS)/matrix spike duplicate (MSD) results. The purpose of MS and MSD is to identify method accuracy and precision. Matrix spikes are generated by the addition of a known amount of target analyte to a sub-sample. Unless the added target analyte is infused within a similar matrix, the ability of the matrix spike to represent method performance is limited; rather, matrix spikes often assist in the identification on chemical interferences inherent in the matrix. The efficiency of any method to dissolute an aqueous standard solution will always be significantly greater than a real world sample.

Five of the 9 samples had no recovery ( $0 \%$ ) of the matrix spike or matrix spike duplicate for PCB 1242. None of these fish samples contained PCB 1242 and as such, this lack of recovery does not affect the data. The MS/MSD results in two of the samples actually fell within the control limits. However, the laboratory had to dilute the samples heavily making it difficult to discern the spike from the actual background PCB and identified this as a possible problem. In the other two samples, the MS/MSD recovery exceeded the control limit of $130 \%$. Both samples had relatively high levels of PCBs which based on the MS/MSD results may be biased high. However, neither sample was identified as an outlier and both had PCB concentrations less than the mean for that reach. As such, it does not appear the results are biased high. None of the MS/MSD problems or potential problems appears to affect the data or conclusions drawn from the data.

Differences in the matrix between fish are more marked than in other environmental media such as soil or groundwater and could be due to the large differences in lipid content. However, according to the laboratory, the matrix spike problem is not attributed to this difference in lipid content. According to Mr. Ted Noltemeyer, Project Manager at PACE Analytical, "The analysis of fish is typically more of a challenge than waters and soils, but our methods and cleanups take care of that. The MS/MSD recoveries here are affected by the relatively high concentrations of PCBs in the samples, not by the matrix itself. Bottom line is most MS/MSD samples required dilutions which negated the ability to appropriately measure the spike recoveries."

### 4.0 Data Analysis

### 4.1 Summary Statistics

Summary statistics are provided with the data in Appendix 1 and in Table 4. The data distribution and upper $95 \%$ confidence levels ( $95 \%$ UCL) were calculated using ProUCL as requested by USEPA. ProUCL documentation is provided in Appendix 4. Consistent with historical results, the variability of the data was rather low and the majority of the data had a normal distribution. ${ }^{9}$ The distribution was calculated by ProUCL using a variety of goodness-offit methods including Shapiro-Wilk, and Kolmogorov-Smirnov tests. Knowledge of the distribution is needed to determine the proper methods for calculating $95 \%$ UCL as well as other statistical tests. Coefficient of variations ranged from 0.22 to 1.67 with an average of 0.59 . The highest coefficient of variations were observed in adult carp with the largest variation observed from Middle River site 1 and the next largest variation at Upper River site 2.

Outliers are inevitable in data sets originating from environmental applications. Outliers are defined to be an observation that does not conform to the pattern established by other observations (Gilbert, 1987). Prior to calculating the UCL, ProUCL recommends an outlier analysis. In the case of the fish tissue data from the baseline monitoring, a few of the results appeared to be outliers because the concentrations was significantly greater than the mean for the same species within the same reach. As such, ProUCL was also used to evaluate the possibility of outliers. ProUCL uses both the Dixon and Rosner outlier tests and uses the Dixon test where the data sets are less than 25 samples. Using ProUCL, a total of six outliers was detected (Appendix 4). These outliers and the significance levels at which they were identified are summarized below. ${ }^{10}$

| Location | Adult Carp | Adult Sucker | Dace | Catfish |
| :--- | :---: | :---: | :---: | :---: |
| Middle River 1 | $22.8 \mathrm{ppm} @ 0.01$ | $19.9 \mathrm{ppm} @ 0.1$ | $17.8 \mathrm{ppm} @ 0.1$ | None |
| Lower River | $44.9 \mathrm{ppm} @ 0.05$ | None | None | $28.4 \mathrm{ppm} @ 0.1$ |
| Inner Harbor | $9.14 \mathrm{ppm} @ 0.1$ | None | None | None |

The outlier analysis identified six samples that were not representative of the river reach. Reasons why these fish are not representative are discussed in the following. The two fish that represented the outliers in the Lower River reach were a carp and a catfish. They smallest fish within their species for the reach but had the highest levels of fat (lipids). As such, the length and weight variables can not explain the differences. The higher levels of lipids may be connected to the only other variable that could explain the difference, habitat. The carp outlier caught in the Lower River could be from the Upper River; its concentration of $17.8 \mathrm{mg} / \mathrm{Kg}$ is very close to the mean for the Upper River ( $25.9 \mathrm{mg} / \mathrm{kg}$ ). The catfish outlier in the Lower River could also have been from the Middle River; site 2 offers suitable habitat for catfish. The Middle River habitat, where the shoreline is much less developed than the Lower River, may have produced a more abundant food supply leading to the large fat content. According to Fishes in Wisconsin, carp range extensively and are capable of jumping dams or falls. As such, it's not

[^6]unexpected that an Upper River carp would be found downstream. Catfish are also known to move great distances and the fish caught in the Lower River could have originated from Middle River.

The PCB content of the adult carp collected in the Inner Harbor is more than $400 \%$ larger than the mean for the remainder of the species in this reach. The size and fat content were within the median of this species collected from this reach. As such, increases in PCB content cannot be attributed to these variables. The only other variable is habitat. The PCB result is very close to the mean results for this species in the Lower River.

The adult carp collected in the Middle River was older and larger than other fish of this species collected in this reach. It also had the second highest fat content. At six years old, it was $50 \%$ older than the other fish collected from the reach and near the end of its life span. This sample may not be representative adult carp in this reach because of its age. The adult sucker collected in this reach was the same age and size (weight, length, fat) as the other fish of this species collected in this reach. White suckers are also known to move about extensively. The longnose dace outlier had a PCB content ( $17.8 \mathrm{mg} / \mathrm{Kg}$ ) that was much closer to those collected in the Upper River reach (mean $13.3 \mathrm{mg} / \mathrm{Kg}$, maximum $17.6 \mathrm{mg} / \mathrm{Kg}$ ) than those in the Middle River reach (mean $7.8 \mathrm{mg} / \mathrm{Kg}$ ). While dace are not known to move much, there size would indicate the possibility of being washed over the dams from the Upper River during high river level events.

Based on this information, the outliers could be eliminated when calculating the summary statistics for the fish species within the reach. However, Region V USEPA requested that this not be done since fish from other reaches can migrate between reaches and represent possible exposure to humans via consumption. As the outliers would only be eliminated in the comparison of fish between sites, reaches, fish species and historical data but not in the covariant analysis, elimination of the outliers has no bearing on protection of human health. Elimination of the outliers allows a clearer understanding of differences between sites, reaches, fish species, and historical data. Regardless, the outliers were not eliminated from the statistical comparisons discussed.

Data analysis included an analysis of means using the $t$-test and analysis of variance (ANOVA). The $t$-test was performed based on unequal variance after an assessment indicated that was the most appropriate test. As far as the appropriateness of the test, PRS reviewed several publications such as A Guide for Selecting Statistical Techniques for Analyzing Social Science Data (The University of Michigan, 1981), Intuitive Biostatistics (Oxford University Press, 1995), Lake and Reservoir Bioassessment and Biocriteria Technical Guidance Document (USEPA, 1998) and Test Methods for Evaluating Solid Waste, Physical/Chemical Methods (USEPA, 2007). All of these indicated the $t$-test was an appropriate method for the comparisons being performed. This was also the test proposed in the approved Lower Fox River Baseline Monitoring Plan.

Both tests can evaluate if there is a significant difference between data sets. ANOVA is actually a collection of statistical methods that can evaluate the conceptual classes of data variability, fixed effect, random effect, and mixed effect. The one-way ANOVA is used to test differences
in two or more independent groups. Since the t -test can be used for two groups, the one-way ANOVA is typically used for analysis of three groups. The ANOVA was used with the $t$-test as an additional test of differences based on a different approach to add a measure of robustness to the evaluation. The tests of differences were performed to evaluate the following:

- Differences in fish species PCB concentrations between sites in the Upper and Middle River reaches,
- Differences in fish species PCB concentrations between the river reaches,
- Difference of fish species PCB concentrations compared to all fish collected, and
- Difference with historical data

No statistical analysis was performed to evaluate differences in PCB concentrations among males and females by reach. Typically, there were insufficient females collected to evaluate. In addition, differences due to age were not evaluated due to the variability of the ages. Neither sex nor adult age would appear to be a factor in decision making as anglers would not differentiate consumption patterns based on these factors.

Based on the redundancy of the t-test and the ANOVA tests, the Mann-Whitney test was used when the t-test and ANOVA results differed and box and whisker plots (boxplots) were also generated. This testing was done at the request of the USEPA. The Mann-Whitney test is a nonparametric test for assessing whether two independent samples of observations come from the same distribution. It is virtually identical to performing an ordinary parametric $t$-test on the data after ranking over the combined samples. The null hypothesis in the Mann-Whitney test is that the two samples are drawn from a single population, and therefore that their probability distributions are equal. It requires the two samples to be independent, and the observations to be ordinal or continuous measurements, i.e. one can at least say, of any two observations, which is the greater.

In descriptive statistics, a box-and-whisker is a convenient way of graphically depicting groups of numerical data through their five-number summaries (the smallest observation (sample minimum), lower quartile ( Q 1 ), median ( Q 2 ), upper quartile ( Q 3 ), and largest observation (sample maximum). Boxplots can be useful to display differences between populations without making any assumptions of the underlying statistical distribution: they are non-parametric. While the boxplots provide a convenient way of comparing data, they were not used for making decisions concerning the data.

Appendix 5 provides the results of the analysis. The t-test and ANOVA analyses were performed in Excel using equations obtained from Practical Statistics for Analytical Chemists (1987). The spreadsheets were validated using examples from the book. The analysis was only performed for the fish that were caught in sufficient quantities needed for each type of analysis. Juvenile fish were also not evaluated because of the infrequency of collection and the failure to collect these in the past.

Boxplots were generated using ProUCL then exported to Excel for formatting. The MannWhitney test was run using U-Test, a Southwestern Medical Center statistical software program.

The results were exported to Excel for formatting. Post-hoc tests were not performed. These tests are difficult to interpret and unless decisions and recommendations based on the statistical tests are accepted, unnecessary.

### 4.2 Comparison of Sites in a Reach

Fish monitoring in the Upper and Middle River reaches were divided into two sites at the request of the WDNR, in order to represent the amount of contaminated sediment removed and the variability expected. As can be seen in Table A5-1 (Appendix 5), there was no significant difference at the $95 \%$ level for fish species collected in the Upper River sites using the t-test. Table A5-2 confirmed this except for carp. As can be seen in Table A5-1, the calculated t -value for carp of 1.71 is very close to the critical value of 1.75 and as such, the ANOVA result is not surprising. The Mann-Whitney test confirmed the t -test (Table A5-3, Appendix 5) for carp indicating there was no significant difference between sites.

In the Middle River, both the $t$-test and the ANOVA indicated a significant difference for suckers and smallmouth bass. The statistical evaluation generally shows there is no difference in the PCB results for fish collected in the different sites of the Upper River reaches. For two out of the three fish species that were collected in sufficient numbers to perform the statistical comparison, there was a significant difference between sites with site 1 having much higher concentrations than site 2.

### 4.3 Comparison of Reaches

In addition to comparing the sites within the Upper and Middle River reaches, all reaches were compared. The t-test (Table A5-4) and ANOVA (Table A5-5) indicated the differences in fish PCB concentrations were significantly different between the Upper River and the Middle River reaches. Consistent with the sampling strategy of the Interim Monitoring Program which did not believe the Middle River and Lower River reaches were very different, the differences in PCB concentrations between the these two reaches were not significantly different. Consequently, the difference between the Upper River and Lower River would be significantly different between reaches. For the Lower River and Inner Harbor reaches, the t-test results indicate significant differences for smallmouth bass. The ANOVA and Whitney Mann tests did not indicate there were significant differences between the reaches for smallmouth bass. The Whitney-Mann test is documented in Table A5-6.

Two variables have been identified that would account for the differences between the Upper River and the Lower river reaches: the magnitude of sediment impact in each of these reaches and the Upper River reach was remediated while the others were not. However, the Inner Harbor has a high level of PCB sediment impact but the fish tissue concentrations are much lower than the Upper River reach (Table 4). Comparison of the fish tissue results in Section 4.4 will provide an evaluation of the differences observed between the Upper River and other reaches.

ProUCL generated boxplots comparing fish species across the reaches are provided in Appendix 5. The boxplots are consistent with Chart 1 showing a general reduction in PCB fish tissue concentration moving from upstream to downstream. The boxplots also identified outliers.

### 4.4 Comparison of Fish

The mean concentrations of each fish species was also compared to the mean concentrations of all fish, excluding the fish species under comparison. Based on the $t$-test (Table A5-7) the concentrations of white suckers, smallmouth bass, longnose dace, and walleye are not significantly different than the concentrations of all fish. However, the ANOVA (Table 5-8) test indicated there were differences for white suckers, smallmouth bass, and longnose dace. This could not be confirmed with the Mann-Whitney test (Table A5-9). The $t$-test and Mann-Whitney analyses indicate that the collection of either white suckers, smallmouth bass, or longnose dace alone could be used to evaluate the trend of fish concentrations following remediation. The data set for the walleye is not sufficiently large to be used however.

### 4.5 Comparison with Historical Data

Finally the data was compared to the historical data ${ }^{11}$, where available (Tables A5-10 and A511). A non-statistical comparison of the means shows the mean concentrations were higher than the most recent historical result. The differences were most extreme in the Upper River sites, the only areas remediated. The smallmouth bass results and Upper River 2 white sucker were higher than the oldest of the Interim Monitoring results as can be seen in Charts 2 through 5 in Appendix $3 .{ }^{12}$

The t-test evaluation indicated that 5 of the 8 adult fish species evaluated had statistically different results in the Upper River sites. The ANOVA evaluation was similar though there was some disagreement as was there with the Mann-Whitney tests (Table A5-12). Based on the weight of evidence, it appears that the remediation of the Upper River caused an increase in the PCB concentrations in the fish. Prior to the fish collection, we anticipated that this may occur due to disturbance of the sediment causing increased suspension of sediment. The increase in biota concentrations following dredging was discussed in Sediment Dredging at Superfund Megasites, Assessing the Effectiveness (National Academy of Sciences, 2007). Cadmium levels in benthic invertebrates increased compared to pre-dredging levels for the first four years following dredging at the Marathon Battery site. ${ }^{13}$ A decrease was not noted until the fifth year. At the Black River site in Ohio, an increase in cancer was noted following dredging that was "probably due to the exposure of fish and their prey to higher concentrations of PAHs in sediment and water during dredging."

While the turbidity was not measured during baseline monitoring, the results of the Lower Fox River baseline monitoring showed a strong correlation between PCB levels in the water column and the total suspended solids (TSS). This is consistent with the National Academy of Sciences findings that dredging exposes biota to more PCBs in the sediment and water column. Dredging increases TSS, which contains PCBs, and increased water column PCB levels, thus increasing exposure to the fish.

[^7]The water column had the highest PCB levels during the fall sampling period in the Lower Fox River study. This would not account for the historical differences in the fish tissue results in the Sheboygan River since the Interim Monitoring program fish collection occurred during the fall.

The repercussion of an increase in fish tissue concentrations following dredging indicates a need for further analysis. The affects of the lipid content of the fish should be evaluated during the comparison. Similar to the Waukegan Superfund site as discussed in Sediment Dredging at Superfund Megasites, Assessing the Effectiveness, the historical comparison was repeated after normalizing the PCB fish tissue results with the percent lipid concentration (Tables A5-13 and A5-14 in Appendix 5). This analysis demonstrated the pre and post-dredging fish tissue concentrations were not much different when using the lipid normalized data. Using nonnormalized data, $58 \%$ of the adult species in the reaches evaluated had statistically significant differences between pre and post-dredging PCB concentrations based on the Mann-Whitney, confirmed t-test or ANOVA test. Using lipid normalized data, $60 \%$ had a significant difference. Clearly, there was another variable besides lipid content controlling the pre and post-dredge PCB concentrations in fish tissues. This variable is apparently remediation.

### 4.6 PCB Correlation and Controlling Variables

### 4.6.1 Linear Regression

During development of the PMP, WDNR had stated that percent lipids and length could be controlling variables for fish tissue PCB content excluding external variables such as TSS, river flow rates, river temperature, etc. The Lower Fox River baseline monitoring indicated there was contradictory information concerning TSS and temperature but that there is a strong correlation between TSS and water column PCB results. Therefore, there could be a correlation between fish tissue PCB content and water column PCB concentrations although we can not evaluate this as water column testing was not performed.

Simple (one-variable) linear regression was evaluated as a data analysis tool. Charts 6 and 7 in Appendix 3 provide the results for this evaluation. Most of the adult fish in the Upper River 1 site and random adult fish from other reaches were evaluated. The regression was not performed using log transformed data since the majority of the data had a normal distribution. Generally the evaluation showed there was a positive correlation between PCB concentrations and percent lipids, for the species evaluated. The highest correlation was for catfish and white suckers, bottom feeders. While these showed good correlation, the other species did not. The evaluation also showed a generally positive but poor correlation between PCB concentrations and length. However, three of the 8 evaluated had a negative correlation and one had basically no slope (Upper River 1 smallmouth bass). These results show one-variable linear regression provides little help analyzing the data and it will not be performed for the remainder of the fish and reaches.

### 4.6.2 Co-variant Analysis

WDNR had recommended during development of the Plan that co-variant analysis be used to assess both lipid content and length to better account for co-variance between these variables. In the fish tissue PCB post remedial monitoring program we will attempt to determine if PCB
concentrations change (on average) between sequential sampling events. In its simplest form we can think of describing the process as a model, where we attempt to "explain" fish tissue concentrations by the sampling event date. For example, if concentrations fall between sampling events 1 and 2 , the sampling event date (as a factor in the model) has a decreasing effect on the fish tissue concentrations.

The variation found within a sampling event in this example is attributed to model error. If the within event model variation is large in comparison to the observed sampling event effect, we cannot conclude one way or the other that concentrations have changed. However, if we can further explain away the within event variation (thereby reducing the model error) it may still be possible to detect a concentration change. Adding covariates to the model attempts to do exactly that. By adding measurements of fish length and percent lipids as explanatory variables, we may reduce within event variation in the model so concentration changes over time are more easily detected.

This type of model is called a covariance model or analysis of covariance (ANCOVA). It is a mixture of regression analysis and analysis of variance (ANOVA) in that both qualitative and quantitative explanatory variables are utilized. The chief independent variables of interest are qualitative, with quantitative variables being introduced mainly to reduce the variance of the error terms.

This analysis will strengthen the statistical comparison of Phase 1 fish tissue results as compared to the baseline results. The analysis was performed by Foth and is documented in Appendix 7. A summary of the results is summarized in Table 5.

Foth concluded that lipids and/or length significantly affected fish tissue PCB concentrations in 17 of 27 data sets. ${ }^{14}$ Lipids had $100 \%$ more affect on PCB concentration than length. In fact, length showed an inverse affect on PCB concentration in several data sets. Both lipids and length contributed to PCB concentrations in 5 of 17 data sets. Lipid content affected PCB concentrations mostly in the two bass species and length most affected the carp. Foth concluded that inclusion of these variables into the analysis would reduce variability in the PCB concentrations. This will allow for a more powerful comparison of the Phase 1 fish monitoring results with the baseline results.

### 4.6.3 Adequacy of Fish-Tissue Samples

The number of each fish species collected during baseline monitoring was determined by using a statistical procedure based on the coefficient of variation of the most recent historical data. If the baseline coefficient of variation is much higher than the historical variation, it could be possible that insufficient fish were collected for the baseline event to detect a $50 \%$ reduction in the fish tissue PCB concentrations. The results of the coefficient of variation comparison are summarized in Table 6. It includes the number of fish to be collected as determined in the Plan compared to the number that would be required based on the coefficient of variation from the baseline event. There is excellent agreement. In 8 of $32(25 \%)$ of the comparisons, it indicates

[^8]the numbers in the plan were not sufficient. Two of these were for rock bass where only 1 additional fish was required. Based on the data available at the time the Plan was developed, a $75 \%$ agreement is excellent.

### 5.0 Phase 1 Monitoring

The number of fish to collect for annual sampling is to be calculated by the same method as used for baseline sampling. With the exception of coefficient of variation, the input variables are the same. Please note, some fish were not collected in sufficient quantities to statistically determine the number of fish necessary for the first Phase 1 monitoring event. The same number of fish collected during the baseline event will be used for these fish.

When the number of fish to be collected as determined by the statistical method is less than 8 , the number was increased to 8 . That is, a minimum of 8 fish will be collected and analyzed. In addition, annual sampling will not collect more fish of a species than was obtained during baseline monitoring. Appendix 6 provides the calculations on the number of fish to collect during the first post remedial annual monitoring event while the following summarizes the results.

| Fish Species | Number of Samples Per River Reach |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Upper <br> (Site 1) | Upper <br> (Site 2) | Middle <br> (Site 1) | Middle <br> (Site 2) | Lower | Inner <br> Harbor |
| Smallmouth Bass | 8 | 8 | 8 | 8 | 8 | 8 |
| Adult Carp | 12 | 16 | 8 | 8 | 8 | 8 |
| Juvenile Carp | 16 | 16 | 8 | 8 | 8 | 8 |
| Adult Suckers | 8 | 8 | 8 | 8 | 8 | 8 |
| Juvenile Suckers | 8 | 8 | 8 | 8 | 8 | 8 |
| Rock Bass | 8 | 8 | 8 | 8 | 8 | 9 |
| Longnose Dace | 8 | 8 | 8 | 8 | 8 | 8 |
| Walleye | 8 | 8 | 8 | 8 | 9 | 8 |
| Catfish | 8 | 8 | 8 | 8 | 8 | 8 |

Only the Upper River reach has been remediated and as such, this reach will be the only portion of the river where post remedial monitoring will occur in 2009. Recommendations to revise the annual monitoring requirements, based on the statistical analysis, are made in Section 6.0. If these recommendations are not accepted, the number of each fish species discussed in this section will be collected in the Upper River reach in 2009 and during the first post remedial event in the other reaches.

### 6.0 Assessment and Recommendations

### 6.1 Sampling Frequency

Comparison of the Upper River results to the historical data shows that the remediation will cause an increase in the PCB concentrations in the fish. Since PCBs bioaccumulate, we should not expect to see a decrease in PCB concentrations in the adult species until they die out and are replaced with fish hatched since the remediation. This indicates that collection of adult fish immediately following remediation has little value and consideration should be given to revising our approach to annual monitoring. Expected fish life spans, based on Fishes in Wisconsin are as follows:

- Adult Carp

9-15 years

- Adult White Suckers
- Smallmouth Bass
- Rock Bass
- Longnose Dace
- Walleye
- Channel Catfish

5 years
Not provided, $5-7$ years ${ }^{15}$
6 - 8 years
Not provided, 3-4 years
6-7 years
8 years

Similar to the earthworm monitoring in the floodplain where the earthworms are not collected following remediation until after the average life span of adult earthworm has passed, collection of adult fish in the years immediately following remediation should be postponed. A recommendation based on all of the assessments will be made at the end of this section.

### 6.2 Sample Locations

The data analysis indicated there was little variability between sites in the Upper River reach. However, the differences in remediation in the Upper River should be considered. A recommendation based on all of the assessments will be made at the end of this section.

### 6.3 Fish Species

The comparison of several adult fish species to all adult fish species indicated smallmouth bass, white suckers, longnose dace, and walleye could be used as indicator species when monitoring trends. White suckers, longnose dace, and walleye could not be collected in all reaches and as such, could not be used as indicators. However, smallmouth bass were successfully collected in all reaches and could be used as an indicator when monitoring trends. A recommendation based on all of the assessments will be made at the end of this section.

[^9]
### 6.4 Fish Sample Numbers

The fish sample numbers specified in the Plan is appropriate and provides statistical confidence and power for decision making. No changes to the number of fish collected in the Phase 1 sampling event or the method of calculating the number of fish is recommended.

### 6.5 Summary of Assessment Recommendations

Based on the data analysis performed, PRS believes that resources would be better utilized if the Phase 1 monitoring was revised. Based on the lack of variability between the two sampling sites established in the Upper River reach, there is no reason to collect fish from both sites. However, the dams do divide the Upper River causing each site to be physically different (depth, flow, etc.). In addition, different PCB mass exist between these sites and the amount and extent of remediation varied. As such, PRS does not propose that the site concept be dropped.

Comparison of the fish concentrations in the Upper River to historical results demonstrate that remediation will cause an increase in PCB concentrations in adult fish tissue. Since PCBs bioaccumulate, there is no reason remediation will affect adult fish that were adults when remediation was performed. As such, PRS recommends that adult fish species not be collected following remediation until such time the adults have died. According to the available data, the average life span is 6.8 years and increases to 7.3 when dace are not considered. However, we propose to begin Phase 1 monitoring of the adult fish five years following remediation, coinciding with sediment sampling. To fulfill the requirements of the ROD which requires annual monitoring but does not specify which fish require monitoring, PRS recommends that adult smallmouth bass be collected annually during the first four years following remediation. Juvenile species of carp and white suckers would also continue to be collected annually following remediation.

PRS also proposes to collect all adult fish every 5 years when the sediment sampling is performed. In the years between sediment sampling, only smallmouth bass would be collected as their concentration is representative of all fish and are easily found through out the river. This would occur until such time that it appears that the adult species, as represented by annual smallmouth bass results or 5 -year adult fish species results, indicates the PCB concentrations are reaching target levels. At that time, all adult fish species will be collected if the decision is being made on annual smallmouth data, to verify that Phase 2 confirmation monitoring can begin. If the 5 -year data indicates Phase 2 monitoring can begin, no additional Phase 1 monitoring will be needed since the decision would be made based on all fish species.

In summary, PRS proposes the following as the post remedial fish monitoring:

- Collect adult smallmouth bass, juvenile carp, and juvenile white suckers annually following remediation for the first five years following remediation,
- Collect all adult and juvenile fish species listed in the this Plan during the first 5year sediment sampling event, and
- Collect adult smallmouth bass, juvenile carp, and juvenile white suckers annually following the first 5 -year sediment sampling event and all adult and juvenile fish species listed in this plan during subsequent 5 -year sediment sampling events until Phase 1 monitoring is completed, and

Based on this recommendation, PRS proposes to sample the following during the Phase 1 annual fish monitoring event, when applicable.

| Fish Species | Number of Samples Per River Reach |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Upper <br> (Site 1) | Upper <br> (Site 2) | Middle <br> (Site 1) | Middle <br> (Site 2) | Lower | Inner <br> Harbor |
| Smallmouth Bass | 8 | 8 | 8 | 8 | 8 | 8 |
| Juvenile Carp | 16 | 16 | 8 | 8 | 8 | 8 |
| Juvenile Suckers | 8 | 8 | 8 | 8 | 8 | 8 |

Phase 1 monitoring based on juvenile fish and adult small mouth bass will require that additional efforts be made to collect juvenile carp to establish baseline conditions. To ensure collection of juvenile carp, the collection of these fish should be performed earlier in the summer when there is a greater chance of encountering this species in the required size range. This baseline monitoring would be performed prior to remediation of the Lower River reaches and in 2009 for the Upper River reach.

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PACE Analytical, Ted Noltemeyer, Project Manager communication with Keith Egan of PRS via e-mail, May 12, 2009.

## Tables

| Table 1 <br> Summary of Targeted Fish Species |  |  |  |
| :---: | :---: | :---: | :---: |
| Fish Species | Characteristics | Habitat Targeted for Collection * |  |
|  |  | Upper - Lower River | Inner Harbor |
| Smallmouth bass | Occurs in all three drainage basins in Wisconsin. A non-migratory fish, they retreat to pools, undercut banks, or fairly deep water to avoid sunlight. Spawn in May through June when the water reaches 55$75^{\circ} \mathrm{F}$. The average length of young-of year in Wisconsin is 2.7 inches by the end of September. The fish begin to reach sexual maturity at the ages of 3-4 depending on sex. The usual longevity is 5-7 years. | Area of little soft sediment. Sandy or gravel bottom best. Area of stumps or downed trees. |  |
| Carp | Occurs in all drainage basins in Wisconsin. It is found in a wide variety of habitats but prefer warm turbid water. Spawn in April to August when the water reaches $65-75^{\circ} \mathrm{F}$. The average length of young-of year in Wisconsin is 3.7 inches by the end of September. In Wisconsin, carp mature between the ages of 2 and 3 depending on the sex. The usual longevity is 9-15 years. They can have a fairly extensive range and can jump small dams. | Areas with vegetation |  |
| White suckers | Occurs in all drainage basins in Wisconsin and is probably the most widespread of all fish in Wisconsin. It is found in warm shallows of estuaries and bays and can tolerate all stream gradients and a wide range of environmental conditions and pollution. Spawn in April to May when the water reaches about $45^{\circ} \mathrm{F}$. The typical length of young-of year in Wisconsin is 2.6 inches by the end of September. The usual longevity is 5 years after maturing between the ages of 2 and 3 . They move about extensively. | Areas with vegetation |  |
| Rock Bass | Occurs in all three drainage basins in Wisconsin. It is found in clear water over a gravel or rocky bottom and is often found near breakwaters and stone-armored shorelines. Often found with other sunfish such as smallmouth bass. Spawn in spring when the water reaches $60-70^{\circ} \mathrm{F}$. The average length of young-of year in Wisconsin is 1.7 inches by the end of September. They reach maturity between ages 2 and 3. The usual longevity is $6-8$ years. They have a limited range. | Prefers clear, rocky, and vegetated stream pools. | Near structures offering protection. Bridge abutments, docks, etc. |
| Longnose Dace | Occurs in all drainage basins in Wisconsin. Occurs in riffles or torrential water over a bottom of boulder and gravel; it generally avoids pools and quiet runs. Spawn in late April to mid-June at an average water temperature of $63^{\circ} \mathrm{F}$. The average length of young-of year in Wisconsin is 1.7 inches by the end of September. The usual longevity is 3-4 years after reaching maturity at age 2 . No information on their range of migration was found. | Area of little soft sediment. Sandy, gravel or cobble bottom that have some vegetation for cover are best. |  |
| Walleye | Present throughout Wisconsin. During the day, hovers in shadows of submerged objects or in shadows of deep water. At dusk, emerge to feed over shallow weed beds or rocky shoals. Spawn in mid-April to midMay when water reaches $42-50^{\circ} \mathrm{F}$. The average length of young-of year in Wisconsin is 3 inches by the end of July. Maturity occurs between the ages of 2 to 5 for males and 5 to 7 for females. The usual longevity is 6-7 years. They have a fairly extensive range and can jump small dams. | Area of little soft sediment. Sandy or gravel bottom best. Area of rough water. |  |
| Catfish | Occurs in all three drainage basins in Wisconsin. It is found in a wide variety of habitats but prefer warm water. Spawn in May or June when the water reaches $75^{\circ} \mathrm{F}$. The average length of young-of year catfish in Wisconsin is 3.4 inches by the end of September. Sexual maturity varies by body of water but it appears both sexes begin maturing by the age of 5 . Few catfish live beyond 8 years. They can have a fairly extensive range. | Prefers some current and deep water with sand, gravel or rubble bottoms. Areas near bank overhangs or downed trees or stumps |  |

Table 2
Baseline Daily Fish Collection Summary

| Date | River <br> Reach | Adult <br> Carp | Adult <br> White <br> Suckers | Juvenile <br> White <br> Suckers | Small <br> Mouth <br> Bass | Rock <br> Bass | Longnose <br> Dace | Walleye | Channel <br> Catfish |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $8 / 18 / 2008$ | UR1 | 3 | 2 |  | 2 |  |  |  |  |
| $8 / 19 / 2008$ | UR1 | 1 | 2 | 4 | 6 | 1 |  |  |  |
| $8 / 20 / 2008$ | UR1 |  |  | 4 |  | 7 |  |  |  |
| $8 / 21 / 2008$ | UR2 | 4 | 4 | 8 | 8 | 3 |  |  |  |
| $8 / 22 / 2008$ | UR2 |  |  |  |  | 4 |  |  |  |
| $8 / 25 / 2008$ | LR | 2 | 2 | 1 | 8 |  |  |  | 4 |
| $8 / 26 / 2008$ | LR |  |  | 3 |  | 9 |  |  |  |
| $8 / 27 / 2008$ | IH | 8 |  |  | 7 |  |  | 1 |  |
| $9 / 2 / 2008$ | IH |  |  |  | 1 |  |  |  | 1 |
| $9 / 3 / 2008$ | LR | 6 |  | 1 |  |  |  |  |  |
| $9 / 5 / 2008$ | IH |  |  |  |  |  |  | 2 |  |
| $9 / 6 / 2008$ | UR1 | 12 | 4 |  |  |  |  |  |  |
| $9 / 6 / 2008$ | UR2 | 12 | 4 |  |  | 1 |  |  |  |
| $9 / 8 / 2008$ | MR2 |  | 8 | 2 | 8 | 2 |  |  | 1 |
| $9 / 10 / 2008$ | MR2 |  |  |  |  |  | 8 |  |  |
| $9 / 10 / 2008$ | MR1 |  |  |  |  |  | 4 |  |  |
| $9 / 11 / 2008$ | MR1 |  |  |  |  |  | 2 |  |  |
| $9 / 12 / 2008$ | UR1 |  |  |  |  |  | 4 |  |  |
| $9 / 15 / 2008$ | MR2 | 1 |  | 5 |  | 6 |  |  | 3 |
| $9 / 16 / 2008$ | MR1 | 8 | 3 |  | 8 |  |  | 8 |  |
| $9 / 17 / 2008$ | MR1 |  | 4 |  |  | 1 |  |  | 4 |
| $9 / 17 / 2008$ | UR1 |  |  |  |  |  | 2 |  |  |
| TOTAL | 57 | 33 | 28 | 48 | 34 | 20 | 11 | 13 |  |
| 9 |  |  |  |  |  |  |  |  |  |

UR1 - Upper River from former Tecumseh Site to Riverbend Dam
UR2 - Upper River from Riverbend Dam to Waelderhaus Dam
MR1 - Middle River from Waelderhaus Dam to Kohler Landfill (County Road A Bridge)
MR2 - Middle River from Kohler Landfill (County Road A Bridge) to C\&NW Railroad Bridge
LR - Lower River from C\&NW Railroad Bridge to Pennsylvania Avenue Bridge
IH - Inner Harbor from Pennsylvania Avenue Bridge to Coast Guard Station

| Baseline Fish Collection Summary |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | UR1 | UR1 | UR2 | UR2 | MR1 | MR1 | MR2 | MR2 | LR | LR | IH | IH |
|  | Target | Collected | Target | Collected | Target | Collected | Target | Collected | Target | Collected | Target | Collected |
| Adult Carp | 16 | 16 | 16 | 16 | 8 | 8 | 8 | 1 | 8 | 8 | 8 | 8 |
| Juvenile Carp | 16 | 0 | 16 | 0 | 8 | 0 | 8 | 0 | 8 | 0 | 8 | 0 |
| Adult White Sucker | 8 | 8 | 8 | 8 | 8 | 7 | 8 | 8 | 8 | 2 | 8 | 0 |
| Juvenile White Sucker | 8 | 8 | 8 | 8 | 8 | 0 | 8 | 7 | 8 | 5 | 8 | 0 |
| Smallmouth Bass | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 |
| Rock Bass | 8 | 8 | 8 | 8 | 8 | 1 | 8 | 8 | 9 | 9 | 9 | 0 |
| Longnose Dace | 8 | 6 | 8 | 0 | 8 | 6 | 8 | 8 | 8 | 0 | 8 | 0 |
| Walleye | 8 | 0 | 8 | 0 | 8 | 8 | 8 | 0 | 9 | 0 | 9 | 3 |
| Channel Catfish | 8 | 0 | 8 | 0 | 8 | 4 | 8 | 4 | 8 | 4 | 8 | 1 |
| Total | 88 | 54 | 88 | 48 | 72 | 42 | 72 | 44 | 74 | 36 | 74 | 20 |
| UR1 - Upper River from former Tecumseh Site to Riverbend Dam <br> UR2 - Upper River from Riverbend Dam to Waelderhaus Dam <br> MR1 - Middle River from Waelderhaus Dam to Kohler Landfill (County Road A Bridge) <br> MR2 - Middle River from Kohler Landfill (County Road A Bridge) to C\&NW Railroad Bridge <br> LR - Lower River from C\&NW Railroad Bridge to Pennsylvania Avenue Bridge <br> IH - Inner Harbor from Pennsylvania Avenue Bridge to Coast Guard Station |  |  |  |  |  |  |  |  |  |  |  |  |


| Table 4 Summary Statistics |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Statistic ${ }^{16}$ | UR1 | UR2 | MR 1 | MR 2 | LR | IH |
| Adult Carp |  |  |  |  |  |  |
| Mean | 25.9 | 14.7 | 4.44 | N/A | 11.3 | 3.16 |
| Minimum | 1.63 | 1.02 | 1.28 | 1.27 | 0.458 | 0.243 |
| Maximum | 73.1 | 47.7 | 22.8 | 1.27 | 44.9 | 9.14 |
| Count | 16 | 16 | 9 | 1 | 9 | 9 |
| Standard Deviation | 21.4 | 15.0 | 7.43 | N/A | 15.2 | 2.81 |
| Coefficient of Variation | 0.83 | 1.02 | 1.67 | N/A | 1.35 | 0.89 |
| Distribution | Normal | Gamma | Non-Par | N/A | Lognormal | Normal |
| 95\% UCL | 35.3 | 24.9 | 15.89 | N/A | 32.63 | 5.05 |
| Adult White Sucker |  |  |  |  |  |  |
| Mean | 12.4 | 8.92 | 8.77 | 3.96 | 4.31 | N/A |
| Minimum | 5.74 | 3.95 | 3.24 | 0.925 | 3.65 | N/A |
| Maximum | 20.6 | 16.6 | 19.9 | 6.98 | 4.96 | N/A |
| Count | 8 | 8 | 8 | 8 | 2 | 0 |
| Standard Deviation | 5.00 | 4.19 | 5.86 | 2.01 | 0.926 | N/A |
| Coefficient of Variation | 0.40 | 0.47 | 0.669 | 0.51 | 0.22 | N/A |
| Distribution | Normal | Normal | Normal | Normal | N/A | N/A |
| 95\% UCL | 15.8 | 11.7 | 13.07 | 5.31 | N/A | N/A |
| Juvenile White Sucker |  |  |  |  |  |  |
| Mean | 6.01 | 6.82 | N/A | 1.37 | 1.04 | N/A |
| Minimum | 1.99 | 3.73 | N/A | 0.980 | 0.587 | N/A |
| Maximum | 9.71 | 11.5 | N/A | 2.03 | 1.64 | N/A |
| Count | 8 | 8 | 0 | 7 | 5 | 0 |
| Standard Deviation | 2.85 | 2.96 | N/A | 0.389 | 0.427 | N/A |
| Coefficient of Variation | 0.47 | 0.43 | N/A | 0.28 | 0.41 | N/A |
| Distribution | Normal | Normal | Normal | Normal | Normal | N/A |
| 95\% UCL | 7.9 | 8.8 | N/A | 1.66 | 1.44 | N/A |
| Small Mouth Bass |  |  |  |  |  |  |
| Mean | 13.0 | 14.5 | 8.75 | 4.30 | 5.77 | 3.36 |
| Minimum | 4.09 | 3.12 | 4.20 | 2.64 | 1.78 | 1.44 |
| Maximum | 22.2 | 33.5 | 18.2 | 7.65 | 10.90 | 4.43 |
| Count | 8 | 8 | 8 | 8 | 8 | 8 |
| Standard Deviation | 7.28 | 11.1 | 4.94 | 1.61 | 3.05 | 1.04 |
| Coefficient of Variation | 0.56 | 0.77 | 0.56 | 0.37 | 0.53 | 0.31 |
|  | Normal | Normal | Normal | Normal | Normal | Normal |
| 95\% UCL | 17.8 | 22.0 | 12.1 | 5.38 | 7.8 | 4.06 |
| Rock Bass |  |  |  |  |  |  |
| Mean | 6.94 | 4.27 | N/A | 2.49 | 2.60 | N/A |
| Minimum | 1.22 | 0.739 | 2.79 | 1.42 | 1.40 | N/A |
| Maximum | 16.8 | 8.72 | 2.79 | 3.70 | 4.27 | N/A |
| Count | 8 | 8 | 1 | 8 | 9 | 0 |
| Standard Deviation | 5.01 | 2.94 | N/A | 0.790 | 1.11 | N/A |
| Coefficient of Variation | 0.72 | 0.69 | N/A | 0.32 | 0.43 | N/A |
| Distribution | Normal | Normal | N/A | Normal | Normal | N/A |
| 95\% UCL | 10.3 | 6.2 | N/A | 3.02 | 3.29 | N/A |

[^10]| Table 4 Summary Statistics |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Statistic ${ }^{16}$ | UR1 | UR2 | MR 1 | MR2 | LR | IH |
| Longnose Dace |  |  |  |  |  |  |
| Mean | 7.67 | N/A | 9.47 | 8.51 | N/A | N/A |
| Minimum | 1.72 | N/A | 7.08 | 4.86 | N/A | N/A |
| Maximum | 17.6 | N/A | 17.8 | 11.0 | N/A | N/A |
| Count | 6 | 0 | 7 | 8 | 0 | 0 |
| Standard Deviation | 6.85 | N/A | 4.15 | 2.25 | N/A | N/A |
| Coefficient of Variation | 0.89 | N/A | 0.44 | 0.26 | N/A | N/A |
| Distribution | Normal | N/A | Non-Par | Normal | N/A | N/A |
| 95\% UCL | 13.3 | N/A | 12.88 | 10.0 | N/A | N/A |
| Channel Catfish |  |  |  |  |  |  |
| Mean | N/A | N/A | 27.9 | 8.18 | 13.7 | N/A |
| Minimum | N/A | N/A | 15.9 | 0.532 | 6.37 | 19.4 |
| Maximum | N/A | N/A | 49.2 | 16.6 | 28.4 | 19.4 |
| Count | 0 | 0 | 4 | 4 | 5 | 1 |
| Standard Deviation | N/A | N/A | 15.6 | 6.62 | 10 | N/A |
| Coefficient of Variation | N/A | N/A | 0.56 | 0.81 | 0.73 | N/A |
| Distribution | N/A | N/A | N/A | N/A | N/A | N/A |
| 95\% UCL | N/A | N/A | 43.2 | 14.7 | 25.1 | N/A |
| Walleye |  |  |  |  |  |  |
| Mean | N/A | N/A | 11.1 | N/A | N/A | 2.03 |
| Minimum | N/A | N/A | 5.58 | N/A | N/A | 1.36 |
| Maximum | N/A | N/A | 16.8 | N/A | N/A | 3.00 |
| Count | 0 | 0 | 8 | 0 | 0 | 3 |
| Standard Deviation | N/A | N/A | 4.63 | N/A | N/A | 0.857 |
| Coefficient of Variation | N/A | N/A | 0.42 | N/A | N/A | 0.42 |
| Distribution | N/A | N/A | Normal | N/A | N/A | N/A |
| 95\% UCL | N/A | N/A | 14.2 | N/A | N/A | 3.00 |
| Mean, Minimum, Maximum, Standard Deviation and $95 \%$ UCL in $\mathrm{mg} / \mathrm{Kg}$. <br> Count is number of samples. <br> Non-Par - Non Parametric Distribution <br> N/A - Not Applicable, insufficient data |  |  |  |  |  |  |


| Table 5 Co-variable Analysis Results |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Reach | Statistic | Adult Carp |  | Adult Suckers |  | Juvenile Suckers |  | Smallmouth Bass |  | Rock Bass |  | Longnose Dace |  | Walleye |  | Catfish |  |
|  |  | Length | Lipids | Length | Lipids | Length | Lipids | Length | Lipids | Length | Lipids | Length | Lipids | Length | Lipids | Length | Lipids |
| Upper River 1 | N | 16 |  | 8 |  | 8 |  | 8 |  | 8 |  | 6 |  | 0 |  | 0 |  |
|  | $\mathrm{R}^{2}$ | 0.39 |  | 0.91 |  | 0.02 |  | 0.20 |  | 0.53 |  | 0.85 |  |  |  |  |  |
|  | Coefficient | 0.195 | 4.59 | -0.485 | 1083.66 | -0.036 | 57.48 | 3.137 | $1911.86$ | -0.876 | 1344.79 | 1.503 | 8.67 | - | - | - | - |
|  | Standard Error | 0.093 | 8.24 | 0.458 | 150.04 | 0.490 | 184.79 | 3.008 | 1690.26 | 1.548 | 870.67 | 0.507 | 19.21 | - | - | - | - |
|  | p (2-tail) | 0.056 | 0.587 | 0.338 | 0.001 | 0.944 | 0.768 | 0.345 | 0.309 | 0.596 | 0.183 | 0.059 | 0.682 | - | - | - | - |
|  | Model | Exponential |  | Linear |  | Exponential |  | Linear |  | Linear |  | Exponential |  | - |  | - |  |
| Upper River 2 | N | 16 |  | 8 |  | 8 |  | 8 |  | 8 |  | 0 |  | 0 |  | 0 |  |
|  | $\mathrm{R}^{2}$ | 0.88 |  | 0.69 |  | 0.09 |  | 0.59 |  | 0.95 |  | - |  | - |  |  |  |
|  | Coefficient | 1.925 | 341.85 | -0.224 | 191.41 | -0.072 | -136.27 | 1.033 | 2442.65 | -1.153 | 645.16 | - | - | - | - | - | - |
|  | Standard Error | 0.517 | 55.57 | 0.096 | 57.33 | 0.180 | 291.98 | 2.564 | 910.63 | 0.514 | 146.26 | - | - | - | - | - | - |
|  | p (2-tail) | 0.003 | 0.000 | 0.067 | 0.021 | 0.707 | 0.660 | 0.704 | 0.044 | 0.075 | 0.007 | - | - | - | - | - | - |
|  | Model | Linear |  | Exponential |  | Exponential |  | Linear |  | Linear |  | - |  | - |  | - |  |
| Middle River 1 | N | 8 |  | 7 |  | 0 |  | 8 |  | 1 |  | 6 |  | 8 |  | 4 |  |
|  | $\mathrm{R}^{2}$ | 0.88 |  | 0.37 |  | - |  | 0.05 |  | - |  | 0.77 |  | 0.96 |  | 0.96 |  |
|  | Coefficient | 0.445 | 20.95 | 0.159 | 33.50 | - | - | 0.655 | 50.73 | - | - | 2.716 | 140.90 | 0.017 | 635.96 | -1.850 | 436.13 |
|  | Standard Error | 0.123 | 18.07 | 0.131 | 96.16 | - | - | 1.731 | 481.56 | - | - | 1.428 | 89.31 | 0.399 | 165.11 | 3.011 | 102.35 |
|  | p (2-tail) | 0.015 | 0.299 | 0.294 | 0.745 | - | - | 0.721 | 0.920 | - | - | 0.153 | 0.213 | 0.967 | 0.012 | 0.649 | 0.147 |
|  | Model | Exponential |  | Exponential |  | - |  | Linear |  | - |  | Linear |  | Linear |  | Linear |  |
| Middle River 2 | N | 1 |  | 8 |  | 7 |  | 8 |  | 8 |  | 8 |  | 0 |  | 4 |  |
|  | $\mathrm{R}^{2}$ | - |  | 0.53 |  | 0.25 |  | 0.37 |  | 0.62 |  | 0.91 |  | - |  | 0.97 |  |
|  | Coefficient | - | - | 0.615 | 265.92 | -0.054 | 89.09 | 0.072 | 57.54 | -0.299 | 102.52 | 0.238 | 12.56 | - | - | 3.040 | -298.64 |
|  | Standard Error | - | - | 0.680 | 134.30 | 0.155 | 79.19 | 0.056 | 37.53 | 0.297 | 48.80 | 0.068 | 3.23 | - | - | 0.491 | 100.24 |
|  | p (2-tail) | - | - | 0.407 | 0.105 | 0.746 | 0.324 | 0.255 | 0.186 | 0.360 | 0.090 | 0.017 | 0.012 | - | - | 0.102 | 0.206 |
|  | Model | - |  | Linear |  | Exponential |  | Exponential |  | Linear |  | Exponential |  | - |  | Linear |  |
| Lower River | N | 8 |  | 2 |  | 5 |  | 8 |  | 9 |  | 0 |  | 0 |  | 4 |  |
|  | $\mathrm{R}^{2}$ | 0.64 |  | - |  | 0.93 |  | 0.76 |  | 0.67 |  | - |  | - |  | 0.86 |  |
|  | Coefficient | -0.639 | 425.64 | - | - | 0.414 | -145.70 | -0.710 | 506.71 | 0.022 | 311.03 | - | - | - | - | -0.508 | 499.45 |
|  | Standard Error | 1.341 | 144.76 | - | - | 0.083 | 75.92 | 0.635 | 125.95 | 0.392 | 89.06 | - | - | - | - | 4.001 | 400.02 |
|  | p (2-tail) | 0.654 | 0.032 | - | - | 0.038 | 0.195 | 0.315 | 0.010 | 0.957 | 0.013 | - | - | - | - | 0.920 | 0.430 |
|  | Model | Linear |  | - |  | Linear |  | Linear |  | Linear |  | - |  | - |  | Linear |  |
| Inner Harbor | N | 8 |  | 0 |  | 0 |  | 8 |  | 0 |  | 0 |  | 3 |  | 1 |  |
|  | $\mathrm{R}^{2}$ | 0.64 |  | - |  | - |  | 0.90 |  | - |  | - |  |  |  | - |  |
|  | Coefficient | 0.165 | 67.67 | - | - | - | - | -0.533 | 196.02 | - | - | - | - | - | - | - | - |
|  | Standard Error | 0.133 | 23.19 | - | - | - | - | 0.081 | 43.76 | - | - | - | - | - | - | - | - |
|  | p (2-tail) | 0.269 | 0.033 | - | - | - | - | 0.001 | 0.007 | - | - | - | - | - | - | - | - |
|  | Model | Exponential |  | - |  | - |  | Linear |  | - |  | - |  | - |  | - |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Table 6 <br> Analysis of Number of Fish Sampling Requirements |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Location and Species | Coefficient of Variation |  | Number of Fish Needed |  |
| Upper River 1 | Historical | Baseline | $\begin{aligned} & \text { Historical } \\ & C / V \end{aligned}$ | Baseline C/V |
| Smallmouth Bass | 0.36 | 0.56 | 8 | 8 |
| Adult Carp | 0.93 | 0.83 | 16 | 12 |
| Adult Suckers | 0.36 | 0.40 | 8 | 8 |
| Juvenile Suckers | 0.36 | 0.47 | 8 | 8 |
| Rock Bass | 0.58 | 0.72 | 8 | 9 |
| Longnose Dace | 0.08 | 0.89 | 8 | 14 |
| Upper River 2 |  |  |  |  |
| Smallmouth Bass | 0.36 | 0.77 | 8 | 11 |
| Adult Carp | 0.93 | 1.02 | 16 | 19 |
| Adult Suckers | 0.66 | 0.47 | 8 | 8 |
| Juvenile Suckers | 0.66 | 0.43 | 8 | 8 |
| Rock Bass | 0.58 | 0.69 | 8 | 9 |
| Middle River 1 |  |  |  |  |
| Smallmouth Bass | 0.36 | 0.56 | 8 | 8 |
| Adult Carp | 0.66 | 1.67 | 8 | 50 |
| Adult Suckers | 0.66 | 0.67 | 8 | 8 |
| Juvenile Suckers | 0.66 | 0.47 | 8 | 8 |
| Longnose Dace | 0.08 | 0.559 | 8 | 8 |
| Walleye | 0.48 | 0.42 | 8 | 8 |
| Catfish | 0.08 | 0.56 | 8 | 8 |
| Middle River 2 |  |  |  |  |
| Smallmouth Bass | 0.36 | 0.37 | 8 | 8 |
| Adult Suckers | 0.66 | 0.51 | 8 | 8 |
| Juvenile Suckers | 0.66 | 0.28 | 8 | 8 |
| Rock Bass | 0.25 | 0.32 | 8 | 8 |
| Lower River |  |  |  |  |
| Smallmouth Bass | 0.69 | 0.53 | 8 | 8 |
| Adult Carp | 0.44 | 1.35 | 8 | 33 |
| Adult Suckers | 0.44 | 0.22 | 8 | 8 |
| Juvenile Suckers | 0.44 | 0.41 | 8 | 8 |
| Rock Bass | 0.58 | 0.43 | 9 | 8 |
| Catfish | 0.07 | 0.73 | 8 | 10 |
| Inner Harbor |  |  |  |  |
| Smallmouth Bass | 0.69 | 0.31 | 8 | 8 |
| Adult Carp | 0.44 | 0.89 | 8 | 14 |
| Walleye | 0.69 | 0.42 | 9 | 8 |
| Catfish | 0.07 | 0.3 | 8 | 8 |
| C/V - Coefficient of variation Comparison made only for fish where sufficient were caught to determine $\mathrm{C} / \mathrm{V}$ |  |  |  |  |








## Appendix 1

Summary of Baseline Fish Tissue Results

## UPPER RIVER 1

FISH SAMPLE RESULTS - UPPER RIVER SITE 1 (UR1)

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender <br> (M/F) | Age $(Y r)^{1}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-UR1-AC1-G, 8/19/08 | Adult Carp | SO | 24.0 | 61.0 | 82.0 | 2325 | F | 7/8 | 4.60\% | 37.0 |
| BL-UR1-AC2-G, 8/18/08 |  |  | 21.0 | 53.3 | 61.0 | 1729 | M | 6 | 1.33\% | 73.1 |
| BL-UR1-AC3-G, 8/18/08 |  |  | 18.0 | 45.7 | 32.0 | 907 | M | 4 | 4.84\% | 1.63 |
| BL-UR1-AC4-G, 8/18/08 |  |  | 19.0 | 48.3 | 50.0 | 1417 | F | 4 | 4.45\% | 7.44 |
| BL-UR1-AC5-G, 9/6/08 |  |  | 15.0 | 38.1 | 30.0 | 850 | F | 4 | 2.19\% | 4.77 |
| BL-UR1-AC6-G, 9/6/08 |  |  | 16.0 | 40.6 | 30.0 | 850 | M | 3/4 | 0.625\% | 14.0 |
| BL-UR1-AC7-G, 9/6/08 |  |  | 20.0 | 50.8 | 64.0 | 1814 | M | 5 | 2.50\% | 17.6 |
| BL-UR1-AC8-G, 9/6/08 |  |  | 19.5 | 49.5 | 48.0 | 1361 | M | 4/5 | 0.340\% | 2.08 |
| BL-UR1-AC9-G, 9/6/08 |  |  | 25.0 | 63.5 | 113 | 3203 | M | 8 | 7.49\% | 53.9 |
| BL-UR1-AC10-G, 9/6/08 |  |  | 24.0 | 61.0 | 124 | 3515 | M | 7/8 | 7.55\% | 28.4 |
| BL-UR1-AC11-G, 9/6/08 |  |  | 21.0 | 53.3 | 69.0 | 1956 | F | 5/6 | 3.44\% | 9.48 |
| BL-UR1-AC12-G, 9/6/08 |  |  | 23.0 | 58.4 | 96.0 | 2722 | M | 7 | 3.02\% | 29.4 |
| BL-UR1-AC13-G, 9/6/08 |  |  | 25.0 | 63.5 | 152 | 4309 | F | 8 | 13.69\% | 33.3 |
| BL-UR1-AC14-G, 9/6/08 |  |  | 25.0 | 63.5 | 123 | 3487 | F | 8 | 1.01\% | 9.55 |
| BL-UR1-AC15-G, 9/6/08 |  |  | 22.5 | 57.2 | 96.0 | 2722 | F | 6/7 | 8.70\% | 55.5 |
| BL-UR1-AC16-G, 9/6/08 |  |  | 23.0 | 58.4 | 100 | 2835 | M | 7 | 7.03\% | 36.9 |
| Mean Result for Adult Carp |  |  | 21.3 | 54.1 | 79.4 | 2250 | NA | 6.01 | 4.55\% | 25.9 |
| Minimum Results for Adult Carp |  |  | 15.0 | 38.1 | 30.0 | 850 | NA | 3.50 | 0.340\% | 1.63 |
| Maximum Results for Adult Carp |  |  | 25.0 | 63.5 | 152.0 | 4309 | NA | 8.00 | 13.69\% | 73.1 |
| Standard Deviation for Adult Carp |  |  | 3.18 | 8.08 | 37.4 | 1059 | NA | 1.65 | 3.60\% | 21.4 |
| Coefficient of Variation for Adult Carp |  |  | 0.149 | 0.149 | 0.471 | 0.471 | NA | 0.274 | 0.791 | 0.83 |
| Distribution for Adult Carp |  |  |  |  |  |  | rmal |  |  |  |
| Upper 95\% UCL for Adult Carp |  |  | 22.9 | 58.1 | 97.7 | 2769 | NA | 6.82 | 6.31\% | 35.3 |

FISH SAMPLE RESULTS - UPPER RIVER SITE 1 (UR1)

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender <br> (M/F) | Age (Yr) ${ }^{1}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-UR1-AWS1-G, 8/18/08 | Adult White Sucker | SO | 16.0 | 40.6 | 24.0 | 680.4 | M | 4 | 1.40\% | 15.9 |
| BL-UR1-AWS2-G, 8/18/08 |  |  | 14.0 | 35.6 | 16.0 | 454 | M | 4 | 1.33\% | 16.6 |
| BL-UR1-AWS3-G, 8/19/08 |  |  | 13.0 | 33.0 | 16.0 | 454 | M | 3 | 0.555\% | 10.3 |
| BL-UR1-AWS4-G, 8/19/08 |  |  | 12.0 | 30.5 | 19.0 | 539 | M | 3 | 1.52\% | 20.6 |
| BL-UR1-AWS5-G, 9/6/08 |  |  | 14.0 | 35.6 | 18.0 | 510 | M | 4 | 0.855\% | 10.6 |
| BL-UR1-AWS6-G, 9/6/08 |  |  | 12.0 | 30.5 | 14.0 | 397 | M | 3 | 0.495\% | 5.74 |
| BL-UR1-AWS7-G, 9/6/08 |  |  | 14.0 | 35.6 | 19.0 | 539 | M | 3 | 0.330\% | 7.34 |
| BL-UR1-AWS8-G, 9/6/08 |  |  | 11.5 | 29.2 | 11.0 | 312 | M | 3 | 0.760\% | 12.3 |
| Mean Result for Adult White Sucker |  |  | 13.3 | 33.8 | 17.1 | 485 | NA | 3.38 | 0.905\% | 12.4 |
| Minimum Results for Adult White Sucker |  |  | 11.5 | 29.2 | 11.0 | 312 | NA | 3.00 | 0.330\% | 5.74 |
| Maximum Results for Adult White Sucker |  |  | 16.0 | 40.6 | 24.0 | 680 | NA | 4.00 | 1.52\% | 20.6 |
| Standard Deviation for Adult White Sucker |  |  | 1.49 | 3.77 | 3.87 | 110 | NA | 0.518 | 0.454\% | 5.00 |
| Coefficient of Variation for Adult White Sucker |  |  | 0.112 | 0.111 | 0.226 | 0.226 | NA | 0.153 | 0.502 | 0.402 |
| Distribution for Adult White Sucker |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Adult White Sucker |  |  | 14.3 | 36.4 | 19.8 | 562 | NA | 3.73 | 1.22\% | 15.8 |
| BL-UR1-JWS1-G, 8/19/08 | Juvenile White Sucker | SO | 6.00 | 15.2 | 2.00 | 56.7 | M | 1 | 0.151\% | 9.71 |
| BL-UR1-JWS2-G, 8/19/08 |  |  | 6.00 | 15.2 | 1.00 | 28.3 | M | 1 | 0.367\% | 8.93 |
| BL-UR1-JWS3-G, 8/19/08 |  |  | 5.00 | 12.7 | 1.00 | 28.3 | M | 1 | 0.462\% | 6.08 |
| BL-UR1-JWS4-G, 8/19/08 |  |  | 6.00 | 15.2 | 2.00 | 56.7 | M | 1 | 0.248\% | 4.85 |
| BL-UR1-JWS5-G, 8/20/08 |  |  | 7.00 | 17.8 | 2.00 | 56.7 | M | 1 | 0.330\% | 7.76 |
| BL-UR1-JWS6-G, 8/20/08 |  |  | 6.00 | 15.2 | 1.00 | 28.3 | M | 1 | 0.638\% | 6.51 |
| BL-UR1-JWS7-G, 8/20/08 |  |  | 6.50 | 16.5 | 2.00 | 56.7 | M | 1 | 0.281\% | 2.28 |
| BL-UR1-JWS8-G, 8/20/08 |  |  | 6.00 | 15.2 | 2.00 | 56.7 | M | 1 | 0.275\% | 1.99 |
| Mean Result for Juvenile White Sucker |  |  | 6.06 | 15.4 | 1.63 | 46.1 | NA | 1.00 | 0.344\% | 6.01 |
| Minimum Results for Juvenile White Sucker |  |  | 5.00 | 12.7 | 1.00 | 28.3 | NA | 1.00 | 0.151\% | 1.99 |
| Maximum Results for Juvenile White Sucker |  |  | 7.00 | 17.8 | 2.00 | 56.7 | NA | 1.00 | 0.638\% | 9.71 |
| Standard Deviation for Juvenile White Sucker |  |  | 0.563 | 1.43 | 0.518 | 14.7 | NA | 0.00 | 0.149\% | 2.85 |
| Coefficient of Variation for Juvenile White Sucker |  |  | 0.093 | 0.093 | 0.318 | 0.318 | NA | 0.00 | 0.434 | 0.474 |
| Distribution for Juvenile White Sucker |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Juvenile White Sucker |  |  | 6.45 | 16.4 | 1.98 | 56.2 | NA | NA | 0.448\% | 7.92 |

FISH SAMPLE RESULTS - UPPER RIVER SITE 1 (UR1)

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender <br> (M/F) | Age (Yr) ${ }^{\text {l }}$ | Fat (\%) | $\begin{gathered} P C B \\ (m g / k g) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-UR1-SB1-G, 8/18/08 | Smallmouth Bass | SO | 13.0 | 33.0 | 22.0 | 624 | F | 5 | 0.625\% | 18.6 |
| BL-UR1-SB2-G, 8/18/08 |  |  | 10.0 | 25.4 | 8.0 | 227 | M | 3 | 0.400\% | 21.5 |
| BL-UR1-SB3-G, 8/19/08 |  |  | 15.0 | 38.1 | 34.0 | 964 | F | 6 | 1.43\% | 15.2 |
| BL-UR1-SB4-G, 8/19/08 |  |  | 10.0 | 25.4 | 11.0 | 312 | M | 3/4 | 0.490\% | 22.2 |
| BL-UR1-SB5-G, 8/19/08 |  |  | 10.0 | 25.4 | 8.0 | 227 | M | 3 | 0.695\% | 7.33 |
| BL-UR1-SB6-G, 8/19/08 |  |  | 11.0 | 27.9 | 12.0 | 340 | M | 3/4 | 0.765\% | 6.14 |
| BL-UR1-SB7-G, 8/19/08 |  |  | 14.0 | 35.6 | 23.0 | 652 | F | 6 | 1.17\% | 8.59 |
| BL-UR1-SB8-G, 8/19/08 |  |  | 10.0 | 25.4 | 8.00 | 227 | M | 4 | 0.430\% | 4.09 |
| Mean Result for Smallmouth Bass |  |  | 11.6 | 29.5 | 15.8 | 447 | NA | 4.25 | 0.750\% | 13.0 |
| Minimum Results for Smallmouth Bass |  |  | 10.0 | 25.4 | 8.00 | 227 | NA | 3.00 | 0.400\% | 4.09 |
| Maximum Results for Smallmouth Bass |  |  | 15.0 | 38.1 | 34.0 | 964 | NA | 6.00 | 1.43\% | 22.2 |
| Standard Deviation for Smallmouth Bass |  |  | 2.07 | 5.25 | 9.57 | 271 | NA | 1.25 | 0.368\% | 7.28 |
| Coefficient of Variation for Smallmouth Bass |  |  | 0.178 | 0.178 | 0.608 | 0.608 | NA | 0.295 | 0.490 | 0.562 |
| Distribution for Smallmouth Bass |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Smallmouth Bass |  |  | 13.1 | 33.2 | 22.4 | 635 | NA | 5.12 | 1.00\% | 17.8 |
| BL-UR1-RB1-G, 8/19/08 | Rock Bass | SO | 8.50 | 21.6 | 8.00 | 227 | M | 5 | 0.415\% | 6.53 |
| BL-UR1-RB2-G, 8/20/08 |  |  | 8.00 | 20.3 | 7.00 | 198 | M | 4/5 | 0.590\% | 5.82 |
| BL-UR1-RB3-G, 8/20/08 |  |  | 5.50 | 14.0 | 2.00 | 57 | M | 4 | 0.775\% | 16.8 |
| BL-UR1-RB4-G, 8/20/08 |  |  | 6.00 | 15.2 | 4.00 | 113 | M | 3/4 | 1.02\% | 10.4 |
| BL-UR1-RB5-G, 8/20/08 |  |  | 6.00 | 15.2 | 4.00 | 113 | M | 4 | 0.581\% | 7.91 |
| BL-UR1-RB6-G, 8/20/08 |  |  | 7.00 | 17.8 | 4.00 | 113 | M | 4 | 0.325\% | 1.22 |
| BL-UR1-RB7-G, 8/20/08 |  |  | 8.00 | 20.3 | 6.00 | 170 | M | 4 | 0.485\% | 1.57 |
| BL-UR1-RB8-G, 8/20/08 |  |  | 5.50 | 14.0 | 3.00 | 85.0 | M | 3 | 0.619\% | 5.30 |
| Mean Result for Rock Bass |  |  | 6.81 | 17.3 | 4.75 | 135 | NA | 4.00 | 0.601\% | 6.94 |
| Minimum Results for Rock Bass |  |  | 5.50 | 14.0 | 2.00 | 56.7 | NA | 3.00 | 0.325\% | 1.22 |
| Maximum Results for Rock Bass |  |  | 8.50 | 21.6 | 8.00 | 227 | NA | 5.00 | 1.02\% | 16.8 |
| Standard Deviation for Rock Bass |  |  | 1.22 | 3.11 | 2.05 | 58.2 | NA | 0.598 | 0.217\% | 5.01 |
| Coefficient of Variation for Rock Bass |  |  | 0.180 | 0.180 | 0.432 | 0.432 | NA | 0.149 | 0.362 | 0.722 |
| Distribution for Rock Bass |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Rock Bass |  |  | 7.66 | 19.5 | 6.17 | 175 | NA | 4.41 | 0.752\% | 10.3 |

FISH SAMPLE RESULTS - UPPER RIVER SITE 1 (UR1)

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender <br> ( $M / F$ ) | Age ( Yr$)^{\prime}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-UR1-LD1-G, 9/12/08 | Longnose Dace | W | 3.00 | 7.62 | 0.260 | 7.37 | TS | NA | 2.77\% | 17.6 |
| BL-UR1-LD2-G, 9/12/08 |  |  | 2.50 | 6.35 | 0.120 | 3.40 | TS | NA | 1.24\% | 3.20 |
| BL-UR1-LD3-G, 9/12/08 |  |  | 2.00 | 5.08 | 0.070 | 1.98 | TS | NA | 1.14\% | 1.72 |
| BL-UR1-LD4-G, 9/12/08 |  |  | 2.50 | 6.35 | 0.100 | 2.83 | TS | NA | 2.30\% | 3.29 |
| BL-UR1-LD5-G, 9/17/08 |  |  | 3.50 | 8.89 | 0.260 | 7.37 | TS | NA | 4.00\% | 15.1 |
| BL-UR1-LD6-G, 9/17/08 |  |  | 2.50 | 6.35 | 0.090 | 2.55 | TS | NA | 4.40\% | 5.11 |
| Mean Result for Longnose Dace |  |  | 2.67 | 6.77 | 0.150 | 4.25 | NA | NA | 2.64\% | 7.67 |
| Minimum Results for Longnose Dace |  |  | 2.00 | 5.08 | 0.070 | 1.98 | NA | NA | 1.140\% | 1.72 |
| Maximum Results for Longnose Dace |  |  | 3.50 | 8.89 | 0.260 | 7.37 | NA | NA | 4.40\% | 17.6 |
| Standard Deviation for Longnose Dace |  |  | 0.516 | 1.31 | 0.087 | 2.46 | NA | NA | 1.363\% | 6.85 |
| Coefficient of Variation for Longnose Dace |  |  | 0.194 | 0.194 | 0.578 | 0.578 | NA | NA | 0.516 | 0.894 |
| Distribution for Longnose Dace |  |  |  |  |  |  | rmal |  |  |  |
| Upper 95\% UCL for Longnose Dace |  |  | 3.08 | 7.82 | 0.22 | 6.22 | NA | NA | 3.73\% | 13.3 |

NA - Not applicable
TS - Too small to gender/age
SO - Scale off, skin on fillet
SOF - Skin off fillet
W - Whole fish
${ }^{1}$ Where fish ages were in between ages, a half age was applied for the calculations. For example: $4 / 5$ would be 4.5 years.

UPPER RIVER 2

FISH SAMPLE RESULTS - UPPER RIVER SITE 2 (UR2)

| Sample ID, Collection Date | Sample Type | Sample Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender <br> (M/F) | Age $(Y r)^{\prime}$ | Fat (\%) | $\begin{gathered} P C B \\ (m g / k g) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-UR2-AC1-G, 8/21/08 | Adult Carp | SO | 21.0 | 53.3 | 70.0 | 1984 | M | 5/6 | 7.39\% | 34.5 |
| BL-UR2-AC2-G, 8/21/08 |  |  | 23.0 | 58.4 | 86.0 | 2438 | M | 6/7 | 2.05\% | 5.14 |
| BL-UR2-AC3-G, 8/21/08 |  |  | 18.0 | 45.7 | 32.0 | 907 | M | 4 | 3.99\% | 3.18 |
| BL-UR2-AC4-G, 8/21/08 |  |  | 15.0 | 38.1 | 31.0 | 879 | M | 4 | 4.64\% | 7.84 |
| BL-UR2-AC5-G, 9/6/08 |  |  | 18.0 | 45.7 | 35.0 | 992 | M | 4 | 1.26\% | 3.73 |
| BL-UR2-AC6-G, 9/6/08 |  |  | 23.5 | 59.7 | 94.0 | 2665 | M | 7 | 3.25\% | 30.2 |
| BL-UR2-AC7-G, 9/6/08 |  |  | 21.5 | 54.6 | 84.0 | 2381 | M | 6 | 0.975\% | 9.23 |
| BL-UR2-AC8-G, 9/6/08 |  |  | 22.5 | 57.2 | 95.0 | 2693 | M | 6 | 3.16\% | 22.7 |
| BL-UR2-AC9-G, 9/6/08 |  |  | 18.0 | 45.7 | 46.0 | 1304 | F | 4 | 0.955\% | 3.55 |
| BL-UR2-AC10-G, 9/6/08 |  |  | 15.0 | 38.1 | 22.0 | 624 | F | 4 | 0.315\% | 1.71 |
| BL-UR2-AC11-G, 9/6/08 |  |  | 25.0 | 63.5 | 122 | 3459 | F | 8 | 10.03\% | 47.7 |
| BL-UR2-AC12-G, 9/6/08 |  |  | 20.5 | 52.1 | 64.0 | 1814 | M | 7 | 1.06\% | 10.5 |
| BL-UR2-AC13-G, 9/6/08 |  |  | 20.0 | 50.8 | 47.0 | 1332 | M | 5 | 0.290\% | 1.02 |
| BL-UR2-AC14-G, 9/6/08 |  |  | 23.0 | 58.4 | 93.0 | 2637 | F | 7 | 2.06\% | 15.8 |
| BL-UR2-AC15-G, 9/6/08 |  |  | 17.5 | 44.5 | 37.0 | 1049 | M | 4/5 | 0.405\% | 1.39 |
| BL-UR2-AC16-G, 9/6/08 |  |  | 24.5 | 62.2 | 120 | 3402 | F | 7/8 | 7.55\% | 37.3 |
| Mean Result for Adult Carp |  |  | 20.4 | 51.8 | 67.4 | 1910 | NA | 5.63 | 3.08\% | 14.7 |
| Minimum Results for Adult Carp |  |  | 15.0 | 38.1 | 22.0 | 624 | NA | 4.00 | 0.290\% | 1.02 |
| Maximum Results for Adult Carp |  |  | 25.0 | 63.5 | 122 | 3459 | NA | 8.00 | 10.0\% | 47.7 |
| Standard Deviation for Adult Carp |  |  | 3.18 | 8.07 | 32.7 | 926 | NA | 1.43 | 2.96\% | 15.0 |
| Coefficient of Variation for Adult Carp |  |  | 0.156 | 0.156 | 0.485 | 0.485 | NA | 0.255 | 0.958 | 1.02 |
| Distribution for Adult Carp |  |  | Gamma |  |  |  |  |  |  |  |
| Upper 95\% UCL for Adult Carp |  |  | 21.9 | 55.7 | 83.4 | 2364 | NA | 6.33 | 4.53\% | 24.9 |


| BL-UR2-AWS1-G, 8/21/08 | Adult White Sucker | SO | 11.0 | 27.9 | 8.00 | 227 | M | 3 | 0.960\% | 10.8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-UR2-AWS2-G, 8/21/08 |  |  | 13.0 | 33.0 | 15.0 | 425 | M | 3 | 1.32\% | 12.0 |
| BL-UR2-AWS3-G, 8/21/08 |  |  | 14.0 | 35.6 | 18.0 | 510 | M | 3 | 1.14\% | 5.04 |
| BL-UR2-AWS4-G, 8/21/08 |  |  | 9.00 | 22.9 | 7.00 | 198 | M | 2 | 0.715\% | 9.44 |
| BL-UR2-AWS5-G, 9/6/08 |  |  | 10.0 | 25.4 | 9.00 | 255 | M | 3 | 0.355\% | 3.95 |
| BL-UR2-AWS6-G, 9/6/08 |  |  | 13.5 | 34.3 | 16.0 | 454 | M | 3 | 1.28\% | 16.6 |
| BL-UR2-AWS7-G, 9/6/08 |  |  | 14.0 | 35.6 | 19.0 | 539 | M | 3 | 1.12\% | 5.95 |
| BL-UR2-AWS8-G, 9/6/08 |  |  | 13.0 | 33.0 | 17.0 | 482 | M | 3 | 0.840\% | 7.52 |
| Mean Result for Adult White Sucker |  |  | 12.2 | 31.0 | 13.6 | 386 | NA | 2.88 | 0.965\% | 8.92 |
| Minimum Results for Adult White Sucker |  |  | 9.00 | 22.9 | 7.00 | 198 | NA | 2.00 | 0.355\% | 3.95 |
| Maximum Results for Adult White Sucker |  |  | 14.0 | 35.6 | 19.0 | 539 | NA | 3.00 | 1.32\% | 16.6 |
| Standard Deviation for Adult White Sucker |  |  | 1.93 | 4.89 | 4.84 | 137 | NA | 0.354 | 0.322\% | 4.19 |
| Coefficient of Variation for Adult White Sucker |  |  | 0.158 | 0.158 | 0.355 | 0.355 | NA | 0.123 | 0.334 | 0.470 |
| Distribution for Adult White Sucker |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Adult White Sucker |  |  | 13.5 | 34.3 | 17.0 | 481 | NA | 3.12 | 1.19\% | 11.7 |

FISH SAMPLE RESULTS - UPPER RIVER SITE 2 (UR2)

| Sample ID, Collection Date | Sample Type | Sample Form | Length (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender $(M / F)$ | Age (Yr) ${ }^{\text {l }}$ | Fat (\%) | $\begin{gathered} P C B \\ (m g / k g) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-UR2-JWS1-G, 8/21/08 | Juvenile <br> White <br> Sucker | SO | 6.00 | 15.2 | 5.00 | 142 | M | 1 | 0.510\% | 4.39 |
| BL-UR2-JWS2-G, 8/21/08 |  |  | 7.00 | 17.8 | 4.00 | 113 | M | 1 | 0.450\% | 11.5 |
| BL-UR2-JWS3-G, 8/21/08 |  |  | 6.00 | 15.2 | 1.00 | 28.3 | M | 1 | 0.580\% | 5.71 |
| BL-UR2-JWS4-G, 8/21/08 |  |  | 5.00 | 12.7 | 1.00 | 28.3 | M | 1 | 0.440\% | 5.96 |
| BL-UR2-JWS5-G, 8/21/08 |  |  | 5.00 | 12.7 | 1.00 | 28.3 | M | 1 | 0.490\% | 9.32 |
| BL-UR2-JWS6-G, 8/21/08 |  |  | 7.00 | 17.8 | 2.00 | 56.7 | M | 1 | 0.410\% | 4.17 |
| BL-UR2-JWS7-G, 8/21/08 |  |  | 8.00 | 20.3 | 3.00 | 85.0 | M | 2 | 0.595\% | 3.73 |
| BL-UR2-JWS8-G, 8/21/08 |  |  | 7.00 | 17.8 | 2.00 | 56.7 | M | 1 | 0.510\% | 9.78 |
| Mean Result for Juvenile White Sucker |  |  | 6.38 | 16.2 | 2.38 | 67.3 | NA | 1.13 | 0.498\% | 6.82 |
| Minimum Results for Juvenile White Sucker |  |  | 5.00 | 12.7 | 1.00 | 28.3 | NA | 1.00 | 0.410\% | 3.73 |
| Maximum Results for Juvenile White Sucker |  |  | 8.00 | 20.3 | 5.00 | 142 | NA | 2.00 | 0.595\% | 11.5 |
| Standard Deviation for Juvenile White Sucker |  |  | 1.06 | 2.69 | 1.51 | 42.7 | NA | 0.354 | 0.065\% | 2.96 |
| Coefficient of Variation for Juvenile White Sucker |  |  | 0.166 | 0.166 | 0.634 | 0.634 | NA | 0.314 | 0.131 | 0.434 |
| Distribution for Juvenile White Sucker |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Juvenile White Sucker |  |  | 7.11 | 18.1 | 3.42 | 96.9 | NA | 1.37 | 0.543\% | 8.80 |


| BL-UR2-SB1-G, 8/21/08 | Smallmouth Bass | SO | 11.0 | 27.9 | 9.00 | 255 | F | 3 | 1.78\% | 28.9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-UR2-SB2-G, 8/21/08 |  |  | 13.0 | 33.0 | 19.0 | 539 | F | 5 | 0.775\% | 5.34 |
| BL-UR2-SB3-G, 8/21/08 |  |  | 11.0 | 27.9 | 11.0 | 312 | M | 3 | 1.16\% | 14.9 |
| BL-UR2-SB4-G, 8/21/08 |  |  | 12.0 | 30.5 | 14.0 | 397 | F | 5 | 1.67\% | 33.5 |
| BL-UR2-SB5-G, 8/2108 |  |  | 13.0 | 33.0 | 19.0 | 539 | F | 5 | 1.26\% | 3.12 |
| BL-UR2-SB6-G, 8/21/08 |  |  | 10.0 | 25.4 | 10.0 | 283 | M | 3 | 0.970\% | 6.41 |
| BL-UR2-SB7-G, 8/21/08 |  |  | 10.0 | 25.4 | 11.0 | 312 | M | 3 | 1.69\% | 13.5 |
| BL-UR2-SB8-G, 8/21/08 |  |  | 10.0 | 25.4 | 8.00 | 227 | M | 3 | 1.29\% | 10.5 |
| Mean Result for Smallmouth Bass |  |  | 11.3 | 28.6 | 12.6 | 358 | NA | 3.75 | 1.32\% | 14.5 |
| Minimum Results for Smallmouth Bass |  |  | 10.0 | 25.4 | 8.00 | 227 | NA | 3.00 | 0.775\% | 3.12 |
| Maximum Results for Smallmouth Bass |  |  | 13.0 | 33.0 | 19.0 | 539 | NA | 5.00 | 1.78\% | 33.5 |
| Standard Deviation for Smallmouth Bass |  |  | 1.28 | 3.26 | 4.31 | 122 | NA | 1.04 | 0.361\% | 11.1 |
| Coefficient of Variation for Smallmouth Bass |  |  | 0.114 | 0.114 | 0.341 | 0.341 | NA | 0.276 | 0.273 | 0.765 |
| Distribution for Smallmouth Bass |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Smallmouth Bass |  |  | 12.1 | 30.8 | 15.6 | 443 | NA | 4.47 | 1.57\% | 22.0 |

FISH SAMPLE RESULTS - UPPER RIVER SITE 2 (UR2)

| Sample ID, Collection Date | Sample Type | Sample Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender (M/F) | Age (Yr) ${ }^{\text {I }}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-UR2-RB1-G, 8/21/08 | Rock Bass | SO | 9.00 | 22.9 | 9.00 | 255 | F | 5 | 0.405\% | 1.04 |
| BL-UR2-RB2-G, 8/21/08 |  |  | 8.00 | 20.3 | 8.00 | 227 | M | 4 | 0.670\% | 4.24 |
| BL-UR2-RB3-G, $8 / 21 / 08$ |  |  | 6.00 | 15.2 | 2.00 | 56.7 | M | 4 | 0.980\% | 8.25 |
| BL-UR2-RB4-G, $8 / 22 / 08$ |  |  | 7.00 | 17.8 | 4.00 | 113 | M | 4 | 1.20\% | 8.72 |
| BL-UR2-RB5-G, $8 / 22 / 08$ |  |  | 8.00 | 20.3 | 7.00 | 198 | M | 5 | 0.470\% | 4.32 |
| BL-UR2-RB6-G, 8/22/08 |  |  | 8.00 | 20.3 | 7.00 | 198 | F | 4 | 0.705\% | 3.78 |
| BL-UR2-RB7-G, 8/22/08 |  |  | 8.00 | 20.3 | 7.00 | 198 | M | 4/5 | 0.580\% | 3.04 |
| BL-UR2-RB8-G, 9/6/08 |  |  | 8.00 | 20.3 | 6.00 | 170 | M | 4 | 0.240\% | 0.739 |
| Mean Result for Rock Bass |  |  | 7.75 | 19.7 | 6.25 | 177 | NA | 4.31 | 0.656\% | 4.27 |
| Minimum Results for Rock Bass |  |  | 6.00 | 15.2 | 2.00 | 56.7 | NA | 4.00 | 0.240\% | 0.739 |
| Maximum Results for Rock Bass |  |  | 9.00 | 22.9 | 9.00 | 255 | NA | 5.00 | 1.200\% | 8.72 |
| Standard Deviation for Rock Bass |  |  | 0.886 | 2.25 | 2.25 | 63.8 | NA | 0.458 | 0.312\% | 2.94 |
| Coefficient of Variation for Rock Bass |  |  | 0.114 | 0.114 | 0.360 | 0.360 | NA | 0.106 | 0.475 | 0.688 |
| Distribution for Rock Bass |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Rock Bass |  |  | 8.36 | 21.2 | 7.81 | 221 | NA | 4.63 | 0.872\% | 6.23 |

NA - Not applicable
SO - Scale off, skin on fillet
SOF - Skin off fillet
W - Whole fish
${ }^{1}$ Where fish ages were in between ages, a half age was applied for the calculations. For example: $4 / 5$ would be 4.5 yers.

## MIDDLE RIVER 1

FISH SAMPLE RESULTS - MIDDLE RIVER SITE 1 (MR1)

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender <br> ( $M / F$ ) | $\begin{gathered} \text { Age } \\ (Y r)^{\prime} \end{gathered}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-MR1-AC1-G, 9/16/08 | Adult Carp | So | 16.0 | 40.6 | 28.0 | 794 | M | 4 | 1.22\% | 2.06 |
| BL-MR1-AC2-G, 9/16/08 |  |  | 16.0 | 40.6 | 23.0 | 652 | M | 4 | 0.770\% | 1.71 |
| BL-MR1-AC3-G, 9/16/08 |  |  | 17.0 | 43.2 | 32.0 | 907 | M | 4 | 0.390\% | 1.33 |
| BL-MR1-AC4-G, 9/16/08 |  |  | 17.0 | 43.2 | 36.0 | 1021 | F | 4 | 3.21\% | 2.51 |
| BL-MR1-AC5-G, 9/16/08 |  |  | 15.5 | 39.4 | 28.0 | 794 | M | 4 | 0.845\% | 1.62 |
| BL-MR1-AC6-G, 9/16/08 |  |  | 16.0 | 40.6 | 25.0 | 709 | M | 4 | 1.17\% | 1.28 |
| BL-MR1-AC7-G, 9/16/08 |  |  | 17.5 | 44.5 | 36.0 | 1021 | M | 4 | 1.14\% | 2.21 |
| BL-MR1-AC8-G, 9/16/08 |  |  | 20.5 | 52.1 | 74.0 | 2098 | F | 6 | 3.16\% | 22.8 |
| Mean Result for Adult Carp |  |  | 16.9 | 43.0 | 35.3 | 999 | NA | 4.25 | 1.49\% | 4.44 |
| Minimum Results for Adult Carp |  |  | 15.5 | 39.4 | 23.0 | 652 | NA | 4.00 | 0.390\% | 1.28 |
| Maximum Results for Adult Carp |  |  | 20.5 | 52.1 | 74.0 | 2098 | NA | 6.00 | 3.21\% | 22.8 |
| Standard Deviation for Adult Carp |  |  | 1.59 | 4.05 | 16.4 | 464 | NA | 0.707 | 1.08\% | 7.43 |
| Coefficient of Variation for Adult Carp |  |  | 0.094 | 0.094 | 0.464 | 0.464 | NA | 0.166 | 0.728 | 1.67 |
| Distribution for Adult Carp |  |  | Non-Parametric |  |  |  |  |  |  |  |
| Upper 95\% UCL for Adult Carp |  |  | 18.0 | 45.8 | 46.6 | 1321 | NA | 4.74 | 2.24\% | 15.89 |
| BL-MR1-AWS1-G, 9/16/08 | Adult White Sucker | SO | 16.0 | 40.6 | 31.0 | 879 | M | 4 | 0.870\% | 3.72 |
| BL-MR1-AWS2-G, 9/16/08 |  |  | 15.0 | 38.1 | 26.0 | 737 | M | 3 | 1.30\% | 11.8 |
| BL-MR1-AWS3-G, 9/16/08 |  |  | 10.0 | 25.4 | 8.0 | 227 | M | 2 | 0.740\% | 3.24 |
| BL-MR1-AWS4-G, 9/17/08 |  |  | 16.0 | 40.6 | 26.0 | 737 | M | 3/4 | 0.795\% | 19.9 |
| BL-MR1-AWS5-G, 9/17/08 |  |  | 16.0 | 40.6 | 28.0 | 794 | M | 4 | 1.50\% | 8.79 |
| BL-MR1-AWS6-G, 9/17/08 |  |  | 14.0 | 35.6 | 18.0 | 510 | M | 3 | 0.705\% | 4.68 |
| BL-MR1-AWS7-G, 9/17/08 |  |  | 16.0 | 40.6 | 27.0 | 765 | M | 4 | 1.01\% | 9.23 |
| Mean Result for Adult White Sucker |  |  | 14.7 | 37.4 | 23.4 | 664 | NA | 3.36 | 0.987\% | 8.77 |
| Minimum Results for Adult White Sucker |  |  | 10.0 | 25.4 | 8.0 | 227 | NA | 2.00 | 0.705\% | 3.24 |
| Maximum Results for Adult White Sucker |  |  | 16.0 | 40.6 | 31.0 | 879 | NA | 4.00 | 1.50\% | 19.9 |
| Standard Deviation for Adult White Sucker |  |  | 2.21 | 5.62 | 7.87 | 223 | NA | 0.748 | 0.303\% | 5.86 |
| Coefficient of Variation for Adult White Sucker |  |  | 0.151 | 0.150 | 0.336 | 0.336 | NA | 0.223 | 0.307 | 0.669 |
| Distribution for Adult White Sucker |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Adult White Sucker |  |  | 16.4 | 41.5 | 29.3 | 829 | NA | 3.91 | 1.21\% | 13.07 |

FISH SAMPLE RESULTS - MIDDLE RIVER SITE 1 (MR1)

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender $(M / F)$ | $\begin{gathered} \text { Age } \\ (\mathrm{Yr})^{\prime} \end{gathered}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-MR1-SB1-G, 9/16/08 | Smallmouth Bass | SO | 13.0 | 33.0 | 22.0 | 624 | M | 5 | 1.37\% | 14.1 |
| BL-MR1-SB2-G, 9/16/08 |  |  | 15.0 | 38.1 | 32.0 | 907 | F | 6 | 2.27\% | 6.04 |
| BL-MR1-SB3-G, 9/16/08 |  |  | 14.0 | 35.6 | 21.0 | 595 | M | 5 | 1.09\% | 5.77 |
| BL-MR1-SB4-G, 9/16/08 |  |  | 14.0 | 35.6 | 21.0 | 595 | F | 5 | 0.815\% | 4.20 |
| BL-MR1-SB5-G, 9/16/08 |  |  | 14.5 | 36.8 | 25.0 | 709 | M | 6 | 0.765\% | 7.46 |
| BL-MR1-SB6-G, 9/16/08 |  |  | 12.0 | 30.5 | 18.0 | 510 | M | 5 | 0.680\% | 9.29 |
| BL-MR1-SB7-G, 9/16/08 |  |  | 15.0 | 38.1 | 30.0 | 850 | M | 6 | 1.30\% | 18.2 |
| BL-MR1-SB8-G, 9/16/08 |  |  | 11.0 | 27.9 | 11.0 | 312 | M | 4 | 0.830\% | 4.97 |
| Mean Result for Smallmouth Bass |  |  | 13.6 | 34.4 | 22.5 | 638 | NA | 5.25 | 1.14\% | 8.75 |
| Minimum Results for Smallmouth Bass |  |  | 11.0 | 27.9 | 11.0 | 312 | NA | 4.00 | 0.680\% | 4.20 |
| Maximum Results for Smallmouth Bass |  |  | 15.0 | 38.1 | 32.0 | 907 | NA | 6.00 | 2.27\% | 18.2 |
| Standard Deviation for Smallmouth Bass |  |  | 1.45 | 3.68 | 6.65 | 189 | NA | 0.707 | 0.521\% | 4.94 |
| Coefficient of Variation for Smallmouth Bass |  |  | 0.107 | 0.107 | 0.296 | 0.296 | NA | 0.135 | 0.458 | 0.565 |
| Distribution for Smallmouth Bass |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Smallmouth Bass |  |  | 14.6 | 37.0 | 27.1 | 769 | NA | 5.74 | 1.50\% | 12.1 |
| BL-MR1-RB1-G, 9/17/08 | Rock Bass | SO | 7.00 | 17.8 | 6.00 | 170 | M | 4 | 0.810\% | 2.79 |
| Mean Result for Rock Bass |  |  | 7.00 | 17.8 | 6.00 | 170 | NA | NA | 0.810\% | 2.79 |
| Minimum Results for Rock Bass |  |  | 7.00 | 17.8 | 6.00 | 170 | NA | NA | 0.810\% | 2.79 |
| Maximum Results for Rock Bass |  |  | 7.00 | 17.8 | 6.00 | 170 | NA | NA | 0.810\% | 2.79 |
| Standard Deviation for Rock Bass |  |  | NA | NA | NA | NA | NA | NA | NA | NA |
| Coefficient of Variation for Rock Bass |  |  | NA | NA | NA | NA | NA | NA | NA | NA |
| Distribution for Rock Bass |  |  | NA | NA | NA | NA | NA | NA | NA | NA |
| Upper 95\% UCL for Rock Bass |  |  | NA | NA | NA | NA | NA | NA | NA | NA |

FISH SAMPLE RESULTS - MIDDLE RIVER SITE 1 (MR1)

| Sample ID, Collection Date | Sample <br> Type | Sample <br> Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender (M/F) | $\begin{gathered} \text { Age } \\ (\mathrm{Yr})^{\prime} \end{gathered}$ | Fat (\%) | $\begin{gathered} P C B \\ (m g / k g) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-MR1-LD1-G, 9/10/08 | Longnose Dace | W | 4.00 | 10.2 | 0.330 | 9.36 | TS | NA | 5.82\% | 17.8 |
| BL-MR1-LD2-G, 9/10/08 |  |  | 3.50 | 8.89 | 0.270 | 7.65 | TS | NA | 2.08\% | 8.35 |
| BL-MR1-LD3-G, 9/10/08 |  |  | 2.00 | 5.08 | 0.080 | 2.27 | TS | NA | 3.64\% | 8.92 |
| BL-MR1-LD4-G, 9/10/08 |  |  | 2.50 | 6.35 | 0.090 | 2.55 | TS | NA | 4.84\% | 7.08 |
| BL-MR1-LD5-G, 9/11/08 |  |  | 2.00 | 5.08 | 0.060 | 1.70 | TS | NA | 2.70\% | 7.10 |
| BL-MR1-LD6-G, 9/11/08 |  |  | 2.00 | 5.08 | 0.060 | 1.70 | TS | NA | 3.09\% | 7.56 |
| Mean Result for Longnose Dace |  |  | 2.67 | 6.78 | 0.148 | 4.21 | NA | NA | 3.70\% | 9.47 |
| Minimum Results for Longnose Dace |  |  | 2.00 | 5.08 | 0.060 | 1.70 | NA | NA | 2.08\% | 7.08 |
| Maximum Results for Longnose Dace |  |  | 4.00 | 10.2 | 0.330 | 9.36 | NA | NA | 5.82\% | 17.8 |
| Standard Deviation for Longnose Dace |  |  | 0.876 | 2.24 | 0.120 | 3.39 | NA | NA | 1.40\% | 4.15 |
| Coefficient of Variation for Longnose Dace |  |  | 0.328 | 0.330 | 0.806 | 0.806 | NA | NA | 0.379 | 0.438 |
| Distribution for Longnose Dace |  |  | Non-Parametric |  |  |  |  |  |  |  |
| Upper 95\% UCL for Longnose Dace |  |  | 3.37 | 8.57 | 0.244 | 6.92 | NA | NA | 4.81\% | 12.88 |
| BL-MR1-CC1-G, 9/17/08 | Channel Catfish | SOF | 21.0 | 53.3 | 55.0 | 1559 | M | 8 | 4.02\% | 15.9 |
| BL-MR1-CC2-G, 9/17/08 |  |  | 22.0 | 55.9 | 71.0 | 2013 | M | 8 | 12.6\% | 49.2 |
| BL-MR1-CC3-G, 9/17/08 |  |  | 19.0 | 48.3 | 42.0 | 1191 | F | 6 | 6.34\% | 29.8 |
| BL-MR1-CC4-G, 9/17/08 |  |  | 20.0 | 50.8 | 59.0 | 1673 | F | 6/7 | 5.27\% | 16.6 |
| Mean Result for Channel Catfish |  |  | 20.5 | 52.1 | 56.8 | 1609 | NA | 7.13 | 7.04\% | 27.9 |
| Minimum Results for Channel Catfish |  |  | 19.0 | 48.3 | 42.0 | 1191 | NA | 6.00 | 4.02\% | 15.9 |
| Maximum Results for Channel Catfish |  |  | 22.0 | 55.9 | 71.0 | 2013 | NA | 8.00 | 12.6\% | 49.2 |
| Standard Deviation for Channel Catfish |  |  | 1.29 | 3.28 | 12.0 | 339 | NA | 1.03 | 3.80\% | 15.6 |
| Coefficient of Variation for Channel Catfish |  |  | 0.063 | 0.063 | 0.211 | 0.211 | NA | 0.145 | 0.539 | 0.559 |
| Distribution for Channel Catfish |  |  | To few samples to determine** |  |  |  |  |  |  |  |
| Upper 95\% UCL for Channel Catfish |  |  | 21.8 | 55.3 | 68.5 | 1941 | NA | 8.135 | 10.8\% | 43.2 |

FISH SAMPLE RESULTS - MIDDLE RIVER SITE 1 (MR1)

| Sample ID, Collection Date | Sample Type | Sample Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender <br> (M/F) | $\begin{gathered} \text { Age } \\ (Y r)^{\prime} \end{gathered}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-MR1-W1-G, 9/16/08 | Walleye | SO | 21.0 | 53.3 | 58.0 | 1644 | M | 6 | 2.33\% | 16.8 |
| BL-MR1-W2-G, 9/16/08 |  |  | 19.5 | 49.5 | 54.0 | 1531 | M | 5 | 2.11\% | 16.3 |
| BL-MR1-W3-G, 9/16/08 |  |  | 12.5 | 31.8 | 12.0 | 340 | M | 2 | 0.595\% | 5.58 |
| BL-MR1-W4-G, 9/16/08 |  |  | 16.0 | 40.6 | 22.0 | 624 | M | 3 | 1.52\% | 13.7 |
| BL-MR1-W5-G, 9/16/08 |  |  | 16.0 | 40.6 | 22.0 | 624 | M | 3 | 0.695\% | 7.93 |
| BL-MR1-W6-G, 9/16/08 |  |  | 17.5 | 44.5 | 33.0 | 936 | M | 4 | 1.61\% | 14.3 |
| BL-MR1-W7-G, 9/16/08 |  |  | 13.0 | 33.0 | 12.0 | 340 | M | 2 | 0.465\% | 6.03 |
| BL-MR1-W8-G, 9/16/08 |  |  | 15.5 | 39.4 | 20.0 | 567 | M | 3 | 1.00\% | 8.41 |
| Mean Result for Walleye |  |  | 16.4 | 41.6 | 29.1 | 826 | NA | 3.50 | 1.29\% | 11.1 |
| Minimum Results for Walleye |  |  | 12.5 | 31.8 | 12.0 | 340 | NA | 2.00 | 0.465\% | 5.58 |
| Maximum Results for Walleye |  |  | 21.0 | 53.3 | 58.0 | 1644 | NA | 6.00 | 2.33\% | 16.8 |
| Standard Deviation for Walleye |  |  | 2.92 | 7.43 | 17.9 | 507 | NA | 1.41 | 0.706\% | 4.63 |
| Coefficient of Variation for Walleye |  |  | 0.179 | 0.179 | 0.614 | 0.614 | NA | 0.404 | 0.548 | 0.416 |
| Distribution for Walleye |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Walleye |  |  | 18.4 | 46.7 | 41.5 | 1177 | NA | 4.48 | 1.78\% | 14.2 |

NA - Not applicable
TS - Too small to gender/age
SO - Scale off, skin on fillet
SOF - Skin off fillet
W - Whole fish
** ProUCL could not determine. Based on the coefficient of variation being less than 1.0 and the majority of other data being normal, it was assumed to be normal and $95 \%$ UCL was determined accordingly.
${ }^{1}$ Where fish ages were in between ages, a half age was applied for the calculations. For example: $4 / 5$ would be 4.5 years.

## MIDDLE RIVER 2

FISH SAMPLE RESULTS - MIDDLE RIVER SITE 2 (MR2)

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length <br> (cm) | Weight <br> (ounces) | Weight <br> (grams) | Gender <br> $(M / F)$ | Age (Yr) ${ }^{\prime}$ | Fat (\%) | PCB <br> (mg/kg) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-MR2-AC1-G, 9/15/08 | Adult Carp | SO | 19.0 | 48.3 | 44.0 | 1247 | M | 5 | $0.730 \%$ | 1.27 |
| Mean Result for Adult Carp | 19.0 | 48.3 | 44.0 | 1247 | NA | NA | $0.730 \%$ | 1.27 |  |  |
| Minimum Results for Adult Carp | 19.0 | 48.3 | 44.0 | 1247 | NA | NA | $0.730 \%$ | 1.27 |  |  |
| Maximum Results for Adult Carp | 19.0 | 48.3 | 44.0 | 1247 | NA | NA | $0.730 \%$ | 1.27 |  |  |
| Standard Deviation for Adult Carp | NA | NA | NA | NA | NA | NA | NA | NA |  |  |
| Coefficient of Variation for Adult Carp | NA | NA | NA | NA | NA | NA | NA | NA |  |  |
| Distribution for Adult Carp | NA | NA | NA | NA | NA | NA | NA | NA |  |  |
| Upper 95\% UCL for Adult Carp | NA | NA | NA | NA | NA | NA | NA | NA |  |  |


| BL-MR2-AWS1-G, 9/8/08 | Adult White Sucker | So | 14.5 | 36.8 | 17.0 | 482 | M | 4 | 0.200\% | 3.24 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-MR2-AWS2-G, 9/8/08 |  |  | 14.5 | 36.8 | 18.0 | 510 | M | 4 | 0.170\% | 2.37 |
| BL-MR2-AWS3-G, 9/8/08 |  |  | 14.0 | 35.6 | 20.0 | 567 | M | 3 | 0.520\% | 3.51 |
| BL-MR2-AWS4-G, 9/8/08 |  |  | 16.0 | 40.6 | 26.0 | 737 | F | 4 | 0.715\% | 3.48 |
| BL-MR2-AWS5-G, 9/8/08 |  |  | 14.0 | 35.6 | 13.0 | 369 | M | 3 | 0.150\% | 0.925 |
| BL-MR2-AWS6-G, 9/8/08 |  |  | 16.0 | 40.6 | 23.0 | 652 | F | 4 | 1.23\% | 6.36 |
| BL-MR2-AWS7-G, 9/8/08 |  |  | 15.0 | 38.1 | 22.0 | 624 | F | 3 | 0.585\% | 6.98 |
| BL-MR2-AWS8-G, 9/8/08 |  |  | 13.5 | 34.3 | 16.0 | 454 | M | 3 | 1.36\% | 4.83 |
| Mean Result for Adult White Sucker |  |  | 14.7 | 37.3 | 19.4 | 549 | NA | 3.50 | 0.616\% | 3.96 |
| Minimum Results for Adult White Sucker |  |  | 13.5 | 34.3 | 13.0 | 369 | NA | 3.00 | 0.150\% | 0.925 |
| Maximum Results for Adult White Sucker |  |  | 16.0 | 40.6 | 26.0 | 737 | NA | 4.00 | 1.36\% | 6.98 |
| Standard Deviation for Adult White Sucker |  |  | 0.923 | 2.35 | 4.21 | 119 | NA | 0.535 | 0.468\% | 2.01 |
| Coefficient of Variation for Adult White Sucker |  |  | 0.063 | 0.063 | 0.217 | 0.217 | NA | 0.153 | 0.760 | 0.508 |
| Distribution for Adult White Sucker |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Adult White Sucker |  |  | 15.3 | 38.9 | 22.3 | 632 | NA | 3.87 | 0.940\% | 5.31 |

FISH SAMPLE RESULTS - MIDDLE RIVER SITE 2 (MR2)

| Sample ID, Collection Date | Sample Type | Sample Form | Length (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender (M/F) | Age (Yr) ${ }^{1}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-MR2-JWS1-G, 9/8/08 | Juvenile White Sucke | SO | 8.00 | 20.3 | 3.00 | 85.0 | M | 2 | 0.480\% | 2.03 |
| BL-MR2-JWS2-G, 9/8/08 |  |  | 8.00 | 20.3 | 3.00 | 85.0 | M | 1 | 0.400\% | 1.20 |
| BL-MR2-JWS3-G, 9/15/08 |  |  | 8.00 | 20.3 | 4.00 | 113 | M | 1 | 0.740\% | 1.76 |
| BL-MR2-JWS4-G, 9/15/08 |  |  | 8.00 | 20.3 | 3.00 | 85.0 | M | 1 | 0.575\% | 1.13 |
| BL-MR2-JWS5-G, 9/15/08 |  |  | 8.00 | 20.3 | 3.00 | 85.0 | M | 1 | 0.270\% | 0.98 |
| BL-MR2-JWS6-G, 9/15/08 |  |  | 8.00 | 20.3 | 3.00 | 85.0 | M | 1 | 0.557\% | 1.08 |
| BL-MR2-JWS7-G, 9/15/08 |  |  | 6.00 | 15.2 | 2.00 | 56.7 | M | 1 | 0.455\% | 1.40 |
| Mean Result for Juvenile White Sucker |  |  | 7.71 | 19.6 | 3.00 | 85.0 | NA | 1.14 | 0.497\% | 1.37 |
| Minimum Results for Juvenile White Sucker |  |  | 6.00 | 15.2 | 2.00 | 56.7 | NA | 1.00 | 0.270\% | 0.98 |
| Maximum Results for Juvenile White Sucker |  |  | 8.00 | 20.3 | 4.00 | 113 | NA | 2.00 | 0.740\% | 2.03 |
| Standard Deviation for Juvenile White Sucker |  |  | 0.756 | 1.92 | 0.577 | 16.4 | NA | 0.378 | 0.148\% | 0.39 |
| Coefficient of Variation for Juvenile White Sucker |  |  | 0.098 | 0.098 | 0.192 | 0.192 | NA | 0.331 | 0.298 | 0.28 |
| Distribution for Juvenile White Sucker |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Juvenile White Sucker |  |  | 8.27 | 21.0 | 3.43 | 97.2 | NA | 1.42 | 0.606\% | 1.66 |



FISH SAMPLE RESULTS - MIDDLE RIVER SITE 2 (MR2)

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender ( $M / F$ ) | Age $(\mathrm{Yr})^{\prime}$ | Fat (\%) | $\begin{gathered} P C B \\ (m g / k g) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-MR2-RB1-G, 9/8/08 | Rock Bass | SO | 7.00 | 17.8 | 4.00 | 113 | M | 4 | 0.480\% | 1.42 |
| BL-MR2-RB2-G, 9/8/08 |  |  | 7.00 | 17.8 | 4.00 | 113 | M | 4 | 0.593\% | 2.09 |
| BL-MR2-RB3-G, 9/15/08 |  |  | 7.00 | 17.8 | 6.00 | 170 | M | 3/4 | 1.24\% | 1.88 |
| BL-MR2-RB4-G, 9/15/08 |  |  | 6.50 | 16.5 | 5.00 | 142 | F | 3 | 1.80\% | 3.47 |
| BL-MR2-RB5-G, 9/15/08 |  |  | 5.50 | 14.0 | 2.00 | 56.7 | M | 3 | 1.02\% | 2.86 |
| BL-MR2-RB6-G, 9/15/08 |  |  | 6.00 | 15.2 | 2.00 | 56.7 | M | 3 | 1.30\% | 3.70 |
| BL-MR2-RB7-G, 9/15/08 |  |  | 6.00 | 15.2 | 2.00 | 56.7 | M | 3 | 0.583\% | 2.27 |
| BL-MR2-RB8-G, 9/15/08 |  |  | 8.00 | 20.3 | 5.00 | 142 | M | 4 | 0.495\% | 2.20 |
| Mean Result for Rock Bass |  |  | 6.63 | 16.8 | 3.75 | 106 | NA | 3.44 | 0.939\% | 2.49 |
| Minimum Results for Rock Bass |  |  | 5.50 | 14.0 | 2.00 | 56.7 | NA | 3.00 | 0.480\% | 1.42 |
| Maximum Results for Rock Bass |  |  | 8.00 | 20.3 | 6.00 | 170 | NA | 4.00 | 1.80\% | 3.70 |
| Standard Deviation for Rock Bass |  |  | 0.791 | 2.01 | 1.58 | 44.8 | NA | 0.496 | 0.482\% | 0.790 |
| Coefficient of Variation for Rock Bass |  |  | 0.119 | 0.119 | 0.422 | 0.422 | NA | 0.144 | 0.513 | 0.318 |
| Distribution for Rock Bass |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Rock Bass |  |  | 7.17 | 18.2 | 4.85 | 137 | NA | 3.78 | 1.27\% | 3.02 |
| BL-MR2-LD1-G, 9/10/08 | Longnose Dace | W | 3.50 | 8.89 | 0.210 | 5.95 | M | NA | 2.84\% | 6.20 |
| BL-MR2-LD2-G, 9/10/08 |  |  | 3.50 | 8.89 | 0.340 | 9.64 | TS | NA | 5.02\% | 9.60 |
| BL-MR2-LD3-G, 9/10/08 |  |  | 3.50 | 8.89 | 0.280 | 7.94 | TS | NA | 6.08\% | 10.9 |
| BL-MR2-LD4-G, 9/10/08 |  |  | 4.00 | 10.2 | 0.390 | 11.1 | TS | NA | 5.50\% | 11.0 |
| BL-MR2-LD5-G, 9/10/08 |  |  | 2.00 | 5.08 | 0.060 | 1.70 | TS | NA | 2.33\% | 4.86 |
| BL-MR2-LD6-G, 9/10/08 |  |  | 2.50 | 6.35 | 0.110 | 3.12 | TS | NA | 5.09\% | 7.17 |
| BL-MR2-LD7-G, 9/10/08 |  |  | 3.50 | 8.89 | 0.260 | 7.37 | TS | NA | 4.13\% | 9.86 |
| BL-MR2-LD8-G, 9/10/08 |  |  | 3.00 | 7.62 | 0.240 | 6.80 | TS | NA | 5.74\% | 8.47 |
| Mean Result for Longnose Dace |  |  | 3.19 | 8.10 | 0.236 | 6.70 | NA | NA | 4.59\% | 8.51 |
| Minimum Results for Longnose Dace |  |  | 2.00 | 5.08 | 0.060 | 1.70 | NA | NA | 2.33\% | 4.86 |
| Maximum Results for Longnose Dace |  |  | 4.00 | 10.2 | 0.390 | 11.1 | NA | NA | 6.08\% | 11.0 |
| Standard Deviation for Longnose Dace |  |  | 0.651 | 1.65 | 0.110 | 3.12 | NA | NA | 1.37\% | 2.25 |
| Coefficient of Variation for Longnose Dace |  |  | 0.204 | 0.204 | 0.465 | 0.465 | NA | NA | 0.299 | 0.264 |
| Distribution for Longnose Dace |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Longnose Dace |  |  | 3.64 | 9.24 | 0.312 | 8.86 | NA | NA | 5.54\% | 10.0 |

FISH SAMPLE RESULTS - MIDDLE RIVER SITE 2 (MR2)

| Sample ID, Collection Date | Sample Type | Sample Form | Length (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender <br> (M/F) | Age (Yr) ${ }^{1}$ | Fat (\%) | $\begin{gathered} P C B \\ (m g / k g) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-MR2-CC1-G, 9/8/08 | Channel Catfish | SOF | 19.0 | 48.3 | 42.0 | 1191 | F | 7 | 4.21\% | 6.90 |
| BL-MR2-CC2-G, 9/15/08 |  |  | 22.0 | 55.9 | 109 | 3090 | M | 7 | 6.01\% | 8.68 |
| BL-MR2-CC3-G, 9/15/08 |  |  | 22.0 | 55.9 | 73.0 | 2070 | M | 6 | 3.45\% | 16.6 |
| BL-MR2-CC4-G, 9/15/08 |  |  | 17.0 | 43.2 | 24.0 | 680 | F | 5 | 3.49\% | 0.532 |
| Mean Result for Channel Catfish |  |  | 20.0 | 50.8 | 62.0 | 1758 | NA | 6.25 | 4.29\% | 8.18 |
| Minimum Results for Channel Catfish |  |  | 17.0 | 43.2 | 24.0 | 680 | NA | 5.00 | 3.45\% | 0.532 |
| Maximum Results for Channel Catfish |  |  | 22.0 | 55.9 | 109 | 3090 | NA | 7.00 | 6.01\% | 16.6 |
| Standard Deviation for Channel Catfish |  |  | 2.45 | 6.22 | 37.3 | 1057 | NA | 0.957 | 1.20\% | 6.62 |
| Coefficient of Variation for Channel Catfish |  |  | 0.122 | 0.122 | 0.602 | 0.602 | NA | 0.153 | 0.280 | 0.809 |
| Distribution for Channel Catfish |  |  | To few samples to determine** |  |  |  |  |  |  |  |
| Upper 95\% UCL for Channel Catfish |  |  | 22.4 | 56.9 | 98.6 | 2794 | NA | 7.19 | 5.46\% | 14.7 |

NA - Not applicable
TS - Too small to gender/age
SO - Scale off, skin on fillet
SOF - Skin off fillet
W - Whole fish
** ProUCL could not determine. Based on the coefficient of variation being less than 1.0 and the majority of other data being normal, it was assumed to be normal and $95 \%$ UCL was determined accordingly.
${ }^{1}$ Where fish ages were in between ages, a half age was applied for the calculations. For example: $4 / 5$ vuld be 4.5 years.

## LOWER RIVER

FISH SAMPLE RESULTS - LOWER RIVER

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length <br> (cm) | Weight (ounces) | Weight (grams) | Gender $(M / F)$ | Age (Yr) ${ }^{\prime}$ | Fat (\%) | $\begin{gathered} P C B \\ (m g / k g) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-LR-AC1-G, 8/25/08 | Adult Carp | SO | 17.5 | 44.5 | 32.0 | 907 | M | 4/5 | 2.46\% | 2.52 |
| BL-LR-AC2-G, $8 / 25 / 08$ |  |  | 24.5 | 62.2 | 112 | 3175 | M | $7 / 8$ | 2.69\% | 15.7 |
| BL-LR-AC3-G, 9/3/08 |  |  | 21.0 | 53.3 | 77.0 | 2183 | F | 6 | 5.51\% | 0.458 |
| BL-LR-AC4-G, 9/3/08 |  |  | 17.5 | 44.5 | 44.0 | 1247 | M | 4/5 | 9.03\% | 44.9 |
| BL-LR-AC5-G, 9/3/08 |  |  | 24.0 | 61.0 | 115 | 3260 | M | 7 | 6.40\% | 18.4 |
| BL-LR-AC6-G, 9/3/08 |  |  | 24.0 | 61.0 | 111 | 3147 | F | 7 | 3.63\% | 4.46 |
| BL-LR-AC7-G, 9/3/08 |  |  | 18.0 | 45.7 | 46.0 | 1304 | M | 5 | 0.825\% | 1.97 |
| BL-LR-AC8-G, 9/3/08 |  |  | 19.5 | 49.5 | 60.0 | 1701 | M | 5/6 | 1.07\% | 1.89 |
| Mean Result for Adult Carp |  |  | 20.8 | 52.7 | 74.6 | 2116 | NA | 5.9 | 3.95\% | 11.3 |
| Minimum Results for Adult Carp |  |  | 17.5 | 44.5 | 32.0 | 907 | NA | 4.50 | 0.825\% | 0.458 |
| Maximum Results for Adult Carp |  |  | 24.5 | 62.2 | 115 | 3260 | NA | 7.50 | 9.03\% | 44.9 |
| Standard Deviation for Adult Carp |  |  | 3.06 | 7.76 | 34.1 | 967 | NA | 1.19 | 2.83\% | 15.2 |
| Coefficient of Variation for Adult Carp |  |  | 0.147 | 0.147 | 0.457 | 0.457 | NA | 0.202 | 0.717 | 1.35 |
| Distribution for Adult Carp |  |  | Lognormal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Adult Carp |  |  | 22.9 | 58.1 | 98 | 2786 | NA | 6.70 | 5.91\% | 32.6 |


| BL-LR-AWS $1-\mathrm{G}, 8 / 25 / 08$ | Adult White | SO | 12.5 | 31.8 | 14.0 | 397 | M | 3 | 1.03\% | 4.96 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-LR-AWS2-G, 8/25/08 | Sucker | SO | 13.5 | 34.3 | 16.0 | 454 | M | 3 | 0.705\% | 3.65 |
| Mean Result for Adult White Sucker |  |  | 13.0 | 33.0 | 15.0 | 425 | NA | 3.00 | 0.865\% | 4.31 |
| Minimum Results for Adult White Sucker |  |  | 12.5 | 31.8 | 14.0 | 397 | NA | 3.00 | 0.705\% | 3.65 |
| Maximum Results for Adult White Sucker |  |  | 13.5 | 34.3 | 16.0 | 454 | NA | 3.00 | 1.03\% | 4.96 |
| Standard Deviation for Adult White Sucker |  |  | 0.707 | 1.80 | 1.41 | 40.1 | NA | 0.00 | 0.226\% | 0.926 |
| Coefficient of Variation for Adult White Sucker |  |  | 0.054 | 0.054 | 0.094 | 0.094 | NA | 0.00 | 0.262 | 0.215 |
| Distribution for Adult White Sucker |  |  | To few samples to determine |  |  |  |  |  |  |  |
| Upper 95\% UCL for Adult White Sucker |  |  | To few samples to determine |  |  |  |  |  |  |  |

FISH SAMPLE RESULTS - LOWER RIVER

| Sample ID, Collection Date | Sample Type | Sample Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender $(M / F)$ | Age (Yr) ${ }^{\prime}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-LR-JWS1-G, 8/25/08 | Juvenile White Sucker | SO | 7.00 | 17.8 | 2.00 | 56.7 | M | 1 | 0.140\% | 1.27 |
| BL-LR-JWS2-G, 8/26/08 |  |  | 8.00 | 20.3 | 3.00 | 85.0 | M | 1 | 0.205\% | 1.64 |
| BL-LR-JWS3-G, 8/26/08 |  |  | 6.50 | 16.5 | 2.00 | 56.7 | M | 1 | 0.245\% | 0.713 |
| BL-LR-JWS4-G, $8 / 26 / 08$ |  |  | 5.00 | 12.7 | 2.00 | 56.7 | M | 1 | 0.094\% | 0.587 |
| BL-LR-JWS5-G, 9/3/08 |  |  | 7.00 | 17.8 | 2.00 | 56.7 | M | 2 | 0.405\% | 0.967 |
| Mean Result for Juvenile White Sucker |  |  | 6.70 | 17.0 | 2.20 | 62.4 | NA | 1.20 | 0.218\% | 1.04 |
| Minimum Results for Juvenile White Sucker |  |  | 5.00 | 12.7 | 2.00 | 56.7 | NA | 1.00 | 0.094\% | 0.587 |
| Maximum Results for Juvenile White Sucker |  |  | 8.00 | 20.3 | 3.00 | 85.0 | NA | 2.00 | 0.405\% | 1.64 |
| Standard Deviation for Juvenile White Sucker |  |  | 1.10 | 2.78 | 0.447 | 12.7 | NA | 0.447 | 0.120\% | 0.427 |
| Coefficient of Variation for Juvenile White Sucker |  |  | 0.163 | 0.163 | 0.203 | 0.203 | NA | 0.373 | 0.550 | 0.413 |
| Distribution for Juvenile White Sucker |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Juvenile White Sucker |  |  | 7.66 | 19.5 | 2.59 | 73.5 | NA | 1.59 | 0.323\% | 1.44 |


| BL-LR-SB1-G, 8/25/08 | SO | 10.0 | 25.4 | 8.00 | 227 | M | 3 | 1.19\% | 8.17 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-LR-SB2-G, $8 / 25 / 08$ |  | 10.5 | 26.7 | 9.00 | 255 | F | 3/4 | 0.380\% | 5.14 |
| BL-LR-SB3-G, 8/25/08 |  | 13.0 | 33.0 | 25.0 | 709 | M | 5 | 0.650\% | 2.02 |
| BL-LR-SB4-G, 8/25/08 Smallmouth |  | 10.0 | 25.4 | 9.00 | 255 | F | 3 | 0.685\% | 1.78 |
| BL-LR-SB5-G, 8/25/08 Bass |  | 12.0 | 30.5 | 15.0 | 425 | F | 3/4 | 1.50\% | 7.01 |
| BL-LR-SB6-G, 8/25/08 |  | 11.0 | 27.9 | 11.0 | 312 | M | 4 | 0.915\% | 4.84 |
| BL-LR-SB7-G, 8/25/08 |  | 12.0 | 30.5 | 17.0 | 482 | M | 5 | 2.13\% | 10.9 |
| BL-LR-SB8-G, 8/25/08 |  | 10.5 | 26.7 | 9.00 | 255 | M | 3 | 1.05\% | 6.30 |
| Mean Result for Smallmouth Bass |  | 11.1 | 28.3 | 12.9 | 365 | NA | 3.75 | 1.06\% | 5.77 |
| Minimum Results for Smallmouth Bass |  | 10.0 | 25.4 | 8.00 | 227 | NA | 3.00 | 0.380\% | 1.78 |
| Maximum Results for Smallmouth Bass |  | 13.0 | 33.0 | 25.0 | 709 | NA | 5.00 | 2.13\% | 10.9 |
| Standard Deviation for Smallmouth Bass |  | 1.09 | 2.78 | 5.87 | 166 | NA | 0.845 | 0.552\% | 3.05 |
| Coefficient of Variation for Smallmouth Bass |  | 0.098 | 0.098 | 0.456 | 0.456 | NA | 0.225 | 0.520 | 0.529 |
| Distribution for Smallmouth Bass |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Smallmouth Bass |  | 11.9 | 30.2 | 16.9 | 480 | NA | 4.34 | 1.44\% | 7.81 |

FISH SAMPLE RESULTS - LOWER RIVER

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender $(M / F)$ | Age (Yr) ${ }^{\prime}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-LR-RB1-G, 8/26/08 | Rock Bass | SO | 7.00 | 17.8 | 4.00 | 113 | M | 3/4 | 0.510\% | 1.76 |
| BL-LR-RB2-G, 8/26/08 |  |  | 6.50 | 16.5 | 4.00 | 113 | M | 3 | 0.410\% | 1.95 |
| BL-LR-RB3-G, 8/26/08 |  |  | 5.50 | 14.0 | 3.00 | 85.0 | M | 3 | 0.283\% | 1.40 |
| BL-LR-RB4-G, 8/26/08 |  |  | 5.00 | 12.7 | 2.00 | 56.7 | M | 2 | 0.982\% | 4.11 |
| BL-LR-RB5-G, 8/26/08 |  |  | 6.50 | 16.5 | 4.00 | 113 | M | 3 | 0.980\% | 3.33 |
| BL-LR-RB6-G, $8 / 26 / 08$ |  |  | 6.50 | 16.5 | 4.00 | 113 | M | 3 | 0.445\% | 1.84 |
| BL-LR-RB7-G, 8/26/08 |  |  | 6.00 | 15.2 | 3.00 | 85.0 | M | 3 | 0.393\% | 1.63 |
| BL-LR-RB8-G, 8/26/08 |  |  | 7.00 | 17.8 | 4.00 | 113 | M | 3 | 0.915\% | 4.27 |
| BL-LR-RB9-G, 8/26/08 |  |  | 6.50 | 16.5 | 4.00 | 113 | M | 3 | 0.300\% | 3.07 |
| Mean Result for Rock Bass |  |  | 6.28 | 15.9 | 3.56 | 101 | NA | 2.94 | 0.580\% | 2.60 |
| Minimum Results for Rock Bass |  |  | 5.00 | 12.7 | 2.00 | 56.7 | NA | 2.00 | 0.283\% | 1.40 |
| Maximum Results for Rock Bass |  |  | 7.00 | 17.8 | 4.00 | 113 | NA | 3.50 | 0.982\% | 4.27 |
| Standard Deviation for Rock Bass |  |  | 0.667 | 1.69 | 0.726 | 20.6 | NA | 0.391 | 0.293\% | 1.11 |
| Coefficient of Variation for Rock Bass |  |  | 0.106 | 0.106 | 0.204 | 0.204 | NA | 0.133 | 0.506 | 0.429 |
| Distribution for Rock Bass |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Rock Bass |  |  | 6.71 | 17.1 | 4.03 | 114 | NA | 3.20 | 0.771\% | 3.29 |

FISH SAMPLE RESULTS - LOWER RIVER

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length (cm) | Weight (ounces) | Weight (grams) | Gender <br> (M/F) | Age (Yr) ${ }^{\prime}$ | Fat (\%) | $\begin{gathered} P C B \\ (m g / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-LR-CC1-G, 8/25/08 | Channel <br> Catfish | SOF | 19.0 | 48.3 | 44.0 | 1247 | M | 6 | 4.11\% | 8.49 |
| BL-LR-CC2-G, 8/25/08 |  |  | 21.0 | 53.3 | 55.0 | 1559 | M | 7 | 4.34\% | 11.7 |
| BL-LR-CC3-G, 8/25/08 |  |  | 20.0 | 50.8 | 58.0 | 1644 | M | 6 | 4.98\% | 6.37 |
| BL-LR-CC4-G, 8/25/08 |  |  | 17.0 | 43.2 | 34.0 | 964 | M | 6 | 7.81\% | 28.4 |
| Mean Result for Channel Catfish |  |  | 19.3 | 48.9 | 47.8 | 1354 | NA | 6.25 | 5.31\% | 13.7 |
| Minimum Results for Channel Catfish |  |  | 17.0 | 43.2 | 34.0 | 964 | NA | 6.00 | 4.11\% | 6.37 |
| Maximum Results for Channel Catfish |  |  | 21.0 | 53.3 | 58.0 | 1644 | NA | 7.00 | 7.81\% | 28.4 |
| Standard Deviation for Channel Catfish |  |  | 1.71 | 4.33 | 11.0 | 311 | NA | 0.500 | 1.71\% | 10.0 |
| Coefficient of Variation for Channel Catfish |  |  | 0.089 | 0.089 | 0.230 | 0.230 | NA | 0.080 | 0.322 | 0.729 |
| Distribution for Channel Catfish |  |  | To few samples to determine** |  |  |  |  |  |  |  |
| Upper 95\% UCL for Channel Catfish |  |  | 20.9 | 53.1 | 58.5 | 1658 | NA | 6.74 | 6.98\% | 25.1 |

## NA - Not applicable

SO - Scale off, skin on fillet
SOF - Skin off fillet
W - Whole fish
** ProUCL could not determine. Based on the coefficient of variation being less than 1.0 and the majority of other data being normal, it was assumed to be normal and $95 \%$ UCL was determined accordingly.
${ }^{1}$ Where fish ages were in between ages, a half age was applied for the calculations. For example: $4 / 5$ would be 4.5 years.

INNER HARBOR

FISH SAMPLE RESULTS - INNER HARBOR

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length <br> (cm) | Weight (ounces) | Weight (grams) | Gender $(M / F)$ | Age (Yr) ${ }^{\text {l }}$ | Fat (\%) | $\begin{gathered} P C B \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-IH-ACl-G, 8/27/08 | Adult Carp | SO | 21.0 | 53.3 | 69.0 | 1956 | M | 6 | 3.83\% | 9.14 |
| BL-IH-AC2-G, 8/27/08 |  |  | 23.0 | 58.4 | 112 | 3175 | M | 7 | 1.91\% | 3.21 |
| BL-LH-AC3-G, 8/27/08 |  |  | 16.5 | 41.9 | 36.0 | 1021 | F | 5 | 2.52\% | 2.46 |
| BL-IH-AC4-G, 8/27/08 |  |  | 17.0 | 43.2 | 37.0 | 1049 | F | 4 | 3.03\% | 5.02 |
| BL-IH-AC5-G, 8/27/08 |  |  | 18.5 | 47.0 | 58.0 | 1644 | M | 5 | 4.04\% | 2.30 |
| BL-IH-AC6-G, 8/27/08 |  |  | 16.5 | 41.9 | 36.0 | 1021 | F | 4/5 | 4.06\% | 2.05 |
| BL-IH-AC7-G, 8/27/08 |  |  | 18.5 | 47.0 | 47.0 | 1332 | M | 5 | 1.29\% | 0.890 |
| BL-IH-AC8-G, 8/27/08 |  |  | 19.0 | 48.3 | 53.0 | 1503 | F | 5 | 0.630\% | 0.243 |
| Mean Result for Adult Carp |  |  | 18.8 | 47.6 | 56.0 | 1588 | NA | 5.19 | 2.66\% | 3.16 |
| Minimum Results for Adult Carp |  |  | 16.5 | 41.9 | 36.0 | 1021 | NA | 4.00 | 0.630\% | 0.243 |
| Maximum Results for Adult Carp |  |  | 23.0 | 58.4 | 112 | 3175 | NA | 7.00 | 4.06\% | 9.14 |
| Standard Deviation for Adult Carp |  |  | 2.28 | 5.79 | 25.5 | 724 | NA | 0.923 | 1.31\% | 2.81 |
| Coefficient of Variation for Adult Carp |  |  | 0.122 | 0.122 | 0.456 | 0.456 | NA | 0.178 | 0.491 | 0.889 |
| Distribution for Adult Carp |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Adult Carp |  |  | 20.33 | 51.64 | 73.69 | 2089 | NA | 5.83 | 3.57\% | 5.05 |
| BL-IH-SB1-G, 8/27/08 | Smallmouth Bass | SO | 15.0 | 38.1 | 31.0 | 879 | M | 6 | 0.680\% | 1.44 |
| BL-IH-SB2-G, 8/27/08 |  |  | 14.0 | 35.6 | 26.0 | 737 | F | 5/6 | 0.855\% | 2.70 |
| BL-IH-SB3-G, 8/27/08 |  |  | 12.0 | 30.5 | 16.0 | 454 | M | 4 | 0.935\% | 4.43 |
| BL-IH-SB4-G, 8/27/08 |  |  | 13.0 | 33.0 | 18.0 | 510 | F | 4 | 1.00\% | 3.10 |
| BL-IH-SB5-G, $8 / 27 / 08$ |  |  | 11.5 | 29.2 | 14.0 | 397 | F | 3 | 0.980\% | 4.18 |
| BL-IH-SB6-G, 8/27/08 |  |  | 11.0 | 27.9 | 13.0 | 369 | F | 3 | 1.13\% | 4.31 |
| BL-IH-SB7-G, 8/27/08 |  |  | 14.0 | 35.6 | 25.0 | 709 | F | 5 | 1.58\% | 3.91 |
| BL-IH-SB8-G, 9/2/08 |  |  | 17.0 | 43.2 | 46.0 | 1304 | M | 7/8 | 1.77\% | 2.83 |
| Mean Result for Smallmouth Bass |  |  | 13.4 | 34.1 | 23.6 | 670 | NA | 4.75 | 1.12\% | 3.36 |
| Minimum Results for Smallmouth Bass |  |  | 11.0 | 27.9 | 13.0 | 369 | NA | 3.00 | 0.680\% | 1.44 |
| Maximum Results for Smallmouth Bass |  |  | 17.0 | 43.2 | 46.0 | 1304 | NA | 7.50 | 1.77\% | 4.43 |
| Standard Deviation for Smallmouth Bass |  |  | 1.99 | 5.05 | 11.1 | 314 | NA | 1.56 | 0.369\% | 1.04 |
| Coefficient of Variation for Smallmouth Bass |  |  | 0.148 | 0.148 | 0.469 | 0.469 | NA | 0.328 | 0.331 | 0.308 |
| Distribution for Smallmouth Bass |  |  | Normal |  |  |  |  |  |  |  |
| Upper 95\% UCL for Smallmouth Bass |  |  | 14.82 | 37.63 | 31.30 | 887 | NA | 5.83 | 1.37\% | 4.06 |

FISH SAMPLE RESULTS - INNER HARBOR

| Sample ID, Collection Date | Sample Type | Sample <br> Form | Length <br> (in) | Length <br> $(\mathrm{cm})$ | Weight <br> (ounces) | Weight <br> $(\mathrm{grams})$ | Gender <br> $(M / F)$ | Age (Yr) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |\(\left(\left.$$
\begin{array}{c}\text { Fat (\%) }\end{array}
$$ \begin{array}{c}PCB <br>

(\mathrm{mg} / \mathrm{kg})\end{array} \right\rvert\,\right.\)

| BL-IH-W1-G, 8/27/08 | So | 21.0 | 53.3 | 79.0 | 2240 | M | 6 | 3.71\% | 3.00 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BL-IH-W2-G, 9/5/08 Walleye |  | 21.0 | 53.3 | 72.0 | 2041 | M | 5/6 | 2.71\% | 1.36 |
| BL-IH-W3-G, 9/5/08 |  | 22.0 | 55.9 | 81.0 | 2296 | M | 6 | 1.72\% | 1.74 |
| Mean Result for Walleye |  | 21.3 | 54.2 | 77.3 | 2192 | NA | 5.83 | 2.71\% | 2.03 |
| Minimum Results for Walleye |  | 21.0 | 53.3 | 72.0 | 2041 | NA | 5.50 | 1.72\% | 1.36 |
| Maximum Results for Walleye |  | 22.0 | 55.9 | 81.0 | 2296 | NA | 6.00 | 3.71\% | 3.00 |
| Standard Deviation for Walleye |  | 0.577 | 1.47 | 4.73 | 134 | NA | 0.289 | 1.00\% | 0.857 |
| Coefficient of Variation for Walleye |  | 0.027 | 0.027 | 0.061 | 0.061 | NA | 0.049 | 0.367 | 0.422 |
| Distribution for Walleye |  | To few samples to determine** |  |  |  |  |  |  |  |
| Upper 95\% UCL for Walleye |  | 21.99 | 55.85 | 82.68 | 2343.96 | NA | 6.16 | 3.84\% | 3.00 |

## NA - Not applicable

SO - Scale off, skin on fillet
SOF - Skin off fillet
W - Whole fish
** ProUCL could not determine. Based on the coefficient of variation being less than 1.0 and the majority of other data being normal, it was assumed to be normal and $95 \%$ UCL was determined accordingly.
${ }^{1}$ Where fish ages were in between ages, a half age was applied for the calculations. For example: $4 / 5$ would be 4.5 years.

Baseline Upper and Lower River Fish Monitoring Report
Appendix 2
Laboratory Analytical Reports
CD Contents

| Report Date | Lab Report \# <br> (CD Link) | Sample Locations |
| :---: | :---: | :--- |
| $08 / 26 / 08$ | 408210 | Upper River 1 |
| $08 / 26 / 08$ | 408211 | Upper River 1 \& 2 |
| $08 / 28 / 08$ | 408330 | Upper River 2 |
| $08 / 28 / 08$ | 408328 | Upper River 2, Lower River |
| $09 / 03 / 08$ | 408460 | Lower River, Inner Harbor |
| $09 / 05 / 08$ | 408619 | Lower River, Inner Harbor |
| $09 / 09 / 08$ | 408719 | Upper River 1 \& 2, Inner Harbor |
| $09 / 09 / 08$ | 408721 | Upper River 2, Middle River 2 |
| $09 / 11 / 08$ | 408870 | Middle River 2 |
| $09 / 18 / 08$ | 409156 | Upper River 1, Middle River 1 \& 2 |
| $09 / 18 / 08$ | 409155 | Middle River 1 \& 2 |
| $09 / 19 / 08$ | 409244 | Middle River 1 |
| $09 / 19 / 08$ | 409245 | Upper River 1, Middle River 1 |

Note: Click on blue CD link to access the report in a new window.

## Appendix 3

## Charts

## Chart 1

2008 Baseline Fish Monitoring PCB Results - Sheboygan River


Chart 2
Upper River Site 1 PCB Results - Small Mouth Bass


## Chart 3

Upper River Site 2 PCB Results - Small Mouth Bass


## Chart 4

Upper River PCB Results - White Sucker


Chart 5
Upper River 2 PCB Results - White Suckers


Chart 6
Upper River 1 Linear Regression Analysis of Selected Fish Species


## Chart 7

Other Reaches Linear Regression of Selected Fish Species


Appendix 4

## ProUCL Documentation

UPPER RIVER 1









## UPPER RIVER 2








## MIDDLE RIVER 1






Warning: A sample size of ' $n$ ' $=\mathbf{6}$ may not adequate enough to compute meaningful and reliable test statistics and estimates!

It is suggested to collect at least 8 to 10 observations using these statistical methods! If possible compute and collect Data Quality Objectives (DQO) based sample size and analytical results.

Warning: There are only 6 Values in this data
Note: It should be noted that even though bootstrap methods may be performed on this data set, the resulting calculations may not be reliable enough to draw conclusions

The literature suggests to use bootstrap methods on data sets having more than $10-15$ observations.



MIDDLE RIVER 2



## Warning: A sample size of ' $n$ ' $=\mathbf{7}$ may not adequate enough to compute meaningful and reliable test statistics and estimates! <br> It is suggested to collect at least 8 to 10 observations using these statistical methods! <br> If possible compute and collect Data Quality Objectives (DQO) based sample size and analytical results. <br> Waming: There are only 7 Values in this data

Note: It should be noted that even though bootstrap methods may be performed on this data set, the resulting calculations may not be reliable enough to draw conclusions

The literature suggests to use bootstrap methods on data sets having more than 10-15 observations.

| Relevant UCL Statistics |  |  |  |
| :---: | :---: | :---: | :---: |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Shapiro Wilk Test Statistic | 0.886 | Shapiro Wilk Test Statistic | 0.921 |
| Shapiro Wilk Critical Value | 0.803 | Shapiro Wilk Critical Value | 0.803 |
| Data appear Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |

## Assuming Lognormal Distribution

| 95\% H-UCL | 1.736 |
| :---: | :---: |
| 95\% Chebyshev (MVUE) UCL | 1.973 |
| 97.5\% Chebyshev (MVUE) UCL | 2.236 |
| 99\% Chebyshev (MVUE) UCL | 2.751 |






## General Statistics

Number of Valid Observations 4
Number of Distinct Observations 4

# Warning: This data set only has 4 observations! <br> Data set is too small to compute reliable and meaningful statistics and estimates! 

The data set for variable Catfish was not processed!

It is suggested to collect at least 8 to 10 observations before using these statistical methods! If possible, compute and collect Data Quality Objectives (DQO) based sample size and analytical results.

## LOWER RIVER





Relevant UCL Statistics




INNER HARBOR




## Appendix 5

Fish Tissue Statistical Analysis
t-test, ANOVA, Mann Whitney Analysis Box and Whisker Plots
t-test, ANOVA, Mann Whitney Analysis

Table A5-1
River Reach Sites Analysis of Means

| Statistic | Carp |  | Sucker |  | Small Mouth Bass |  | Rock Bass |  | Dace |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | UR1 | UR2 | UR1 | UR2 | UR1 | UR2 | UR1 | UR2 | UR1 | UR2 |
| Upper River |  |  |  |  |  |  |  |  |  |  |
| Mean | 25.9 | 14.7 | 12.4 | 8.92 | 13 | 14.5 | 6.94 | 4.27 | 7.67 | No Fish |
| Standard Deviation | 21.4 | 15 | 5 | 4.19 | 7.28 | 11.1 | 5.01 | 2.94 | 6.85 |  |
| Count | 16 | 16 | 8 | 8 | 8 | 8 | 8 | 8 | 6 |  |
| t | 1.71 |  | 1.51 |  | 0.32 |  | 1.30 |  | Not Applicable |  |
| Critcal Value at $\mathrm{t}_{0.1 / 2}$ | 1.75 |  | 1.86 |  | 1.86 |  | 1.86 |  |  |  |  |
| Significant Difference | No |  | No |  | No |  | No |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Statistic | MR1 | MR2 | MR1 | MR2 | MR1 | MR2 | MR1 | MR2 | MR1 | MR2 |
| Middle River |  |  |  |  |  |  |  |  |  |  |
| Mean | 4.44 |  | 8.77 | 3.96 | 8.75 | 4.3 |  | 2.49 | 9.47 | 8.51 |
| Standard Deviation | 7.43 |  | 5.86 | 2.01 | 4.94 | 1.61 |  | 0.78 | 4.15 | 2.25 |
| Count | 8 |  | 7 | 8 | 8 | 8 |  | 8 | 6 | 8 |
| t | Not Applicable |  | 2.07 |  | 2.42 |  | Not Applicable |  | 0.51 |  |
| Critcal Value at $t_{0.1 / 2}$ |  |  | 1.86 |  | 1.86 |  |  |  | 1.86 |  |
| Significant Difference |  |  | Yes |  | Yes |  |  |  | No |  |

Mean and Standard Deviation in $\mathrm{mg} / \mathrm{Kg}$
Values in Red exceed the Critical Value and the means for the data sets are significantly different.

Table A5-2

## River Reach Sites Analysis of Variance

| Statistic | White Sucker UR 1/UR 2 |  |  | White Sucker MR 1/MR 2 |  |  | Carp UR 1/UR 2 |  |  | Carp MR 1/MR 2 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square |
| SSc | 49.2 | 1 | 49.16 | 86.2 | 1 | 86.15 | 996.7 | 1 | 996.68 | 8.9 | 1 | 8.93 |
| SSw | 297.8 | 14 | 21.27 | 234.4 | 13 | 18.03 | 6247.9 | 30 | 208.26 | 386.5 | 7 | 55.22 |
| SSt | 346.9 | 15 |  | 320.6 | 14 |  | 7244.6 | 31 |  | 395.4 | 8 |  |
| F statistic | 2.3 |  |  | 4.8 |  |  | 4.8 |  |  | 0.2 |  |  |
| $\mathrm{F}_{0.05,1, \mathrm{SSwDF}}$ | 4.6 |  |  | 4.67 |  |  | 4.17 |  |  | 5.59 |  |  |
| Significant Difference |  | No |  |  | Yes |  |  | Yes |  |  | No |  |

Values in Red exceed
the F Value and the data
sets are significantly
different.

Table A5-2
River Reach Sites Analysis of Variance

| Statistic | SM Bass UR 1/UR 2 |  |  | SM Bass MR 1/MR 2 |  |  | Rock Bass UR 1/UR 2 |  |  | Longnose Dace MR 1/MR 2 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square |
| SSc | 9.8 | 1 | 9.80 | 79.3 | 1 | 79.28 | 28.7 | 1 | 28.68 | 3.2 | 1 | 3.16 |
| SSw | 1235.5 | 14 | 88.25 | 189.1 | 14 | 13.51 | 236.1 | 14 | 16.86 | 121.3 | 12 | 10.11 |
| SSt | 1245.3 | 15 |  | 268.4 | 15 |  | 264.8 | 15 |  | 124.5 | 13 |  |
| F statistic | 0.1 |  |  | 5.9 |  |  | 1.7 |  |  | 0.3 |  |  |
| $\mathrm{F}_{0.05,1, \mathrm{SSWDF}}$ | 4.6 |  |  | 4.6 |  |  | 4.6 |  |  | 4.67 |  |  |
| Significant Difference | No |  |  | Yes |  |  | No |  |  | No |  |  |

Values in Red exceed
the F Value and the data
sets are significantly
different.

Table A5-3
River Reach Analysis of Population
River Reach Sites Mann Whitney Test

| UR1 | UR2 |
| :---: | :---: |
| 37.0 | 34.5 |
| 73.1 | 5.14 |
| 1.63 | 3.18 |
| 7.44 | 7.84 |
| 4.77 | 3.73 |
| 14.0 | 30.2 |
| 17.6 | 9.23 |
| 2.08 | 22.7 |
| 53.9 | 3.55 |
| 28.4 | 1.71 |
| 9.48 | 47.7 |
| 29.4 | 10.5 |
| 33.3 | 1.02 |
| 9.55 | 15.8 |
| 55.5 | 1.39 |
| 36.9 | 37.3 |

Results in $\mathrm{mg} / \mathrm{Kg}$

| $\mathbf{n}_{\mathbf{1}}$ | $\mathbf{n}_{\mathbf{2}}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| :---: | :---: | :---: | :---: | :---: |
| 16 | 16 | 170 | $0.117926^{*}$ | $0.058963^{*}$ |
| normal approx <br> $\mathrm{z}=1.58293$ |  |  | $0.1134364^{*}$ | $0.0567182^{*}$ |

*These values are approximate.
The two samples are not significantly different ( $\mathbf{P}>=\mathbf{0 . 0 5}$, two-tailed test).

Table A5-4

## River Reach Analysis of Means

| Statistic | Carp |  | Sucker |  | Small Mouth Bass |  | Rock Bass |  | Dace |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | UR | MR | UR | MR | UR | MR | UR | MR | UR | MR |
| Upper to Middle River |  |  |  |  |  |  |  |  |  |  |
| Mean | 20.3 | 4.1 | 10.7 | 6.2 | 13.7 | 6.5 | 5.6 | 2.5 | 7.7 | 8.9 |
| Standard Deviation | 19.1 | 7.0 | 4.8 | 4.8 | 9.1 | 4.2 | 4.2 | 0.7 | 6.9 | 3.1 |
| Count | 32 | 9 | 16 | 15 | 16 | 16 | 16 | 9 | 6 | 14 |
| t | 3.95 |  | 2.59 |  | 2.87 |  | 2.86 |  | 0.43 |  |
| Critcal Value at $\mathrm{t}_{0.1 / 2}$ | 1.70 |  | 1.75 |  | 1.75 |  | 1.75 |  | 1.76 |  |
| Significant Difference | Yes |  | Yes |  | Yes |  | Yes |  | No |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  | Carp |  | Sucker |  | Small Mouth Bass |  | Rock Bass |  | Catfish |  |
| Statistic | MR | LR | MR | LR | MR | LR | MR | LR | MR | LR |
| Middle River to Lower River |  |  |  |  |  |  |  |  |  |  |
| Mean | 4.1 | 11.3 | 6.2 | 4.3 | 6.5 | 5.8 | 2.5 | 2.6 | 18.0 | 13.7 |
| Standard Deviation | 7.0 | 15.2 | 4.8 | 0.9 | 4.2 | 3.1 | 0.7 | 1.1 | 15.3 | 10.0 |
| Count | 9 | 8 | 15 | 2 | 16 | 8 | 9 | 9 | 8 | 4 |
| t | 1.23 |  | 1.36 |  | 0.50 |  | 0.17 |  | 0.58 |  |
| Critcal Value at $\mathrm{t}_{0.1 / 2}$ | 1.83 |  | 1.75 |  | 1.75 |  | 1.83 |  | 1.86 |  |
| Significant Difference | No |  | No |  | No |  | No |  | No |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  | Carp |  | Sucker |  | Small Mouth Bass |  | Rock Bass |  | Catfish |  |
| Statistic | LR | IH | LR | IH | LR | IH | LR | IH | LR | IH |
| Lower River to Inner Harbor |  |  |  |  |  |  |  |  |  |  |
| Mean | 11.3 | 3.2 | Too Few Fish |  | 5.8 | 3.4 | 2.6 | Too Few Fish | 13.7 | 19.4 |
| Standard Deviation | 15.2 | 2.8 |  |  | 3.1 | 1.0 | 1.1 |  | 10.0 | 0 |
| Count | 8 | 8 |  |  | 8 | 8 | 9 |  | 4 | 1 |
|  | 1.49 |  | Not Applicable |  | 2.11 |  | Not Applicable |  | 1.13 |  |
| Critcal Value at $\mathrm{t}_{0.1 / 2}$ | 1.86 |  |  |  | 1.86 |  |  |  | 2.13 |  |
| Significant Difference | No |  |  |  | Yes |  |  |  | No |  |

## Mean and Standard Deviation in $\mathrm{mg} / \mathrm{Kg}$

Values in Red exceed the Critical Value and the means for the data sets are significantly different.

Table A5-5
River Reach Analysis of Variance

| Statistic | White Sucker UR/MR |  |  | Carp UR/MR |  |  | Smallmouth Bass UR/MR |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean Square |
| SSc | 154.4 | 1 | 154.41 | 1846.0 | 1 | 1846.03 | 415.9 | 1 | 415.92 |
| SSw | 524.1 | 29 | 18.07 | 9859.0 | 39 | 252.80 | 1347.5 | 30 | 44.92 |
| SSt | 678.5 | 30 |  | 11705.1 | 40 |  | 1763.5 | 31 |  |
| F statistic | 8.5 |  |  | 7.3 |  |  | 9.3 |  |  |
| $\mathrm{F}_{0.05,1, \text { SSwDF }}$ | 4.18 |  |  | 4.1 |  |  | 4.17 |  |  |
| Significant Difference | Yes |  |  | Yes |  |  | Yes |  |  |

Values in Red exceed
the F Value and the data
sets are significantly

Table A5-5

## River Reach Analysis of Variance

| Statistic | Rock Bass UR/MR |  |  | Dace UR/MR |  |  | White Sucker MR/LR |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean Square |
| SSc | 54.8 | 1 | 54.81 | 6.6 | 1 | 6.56 | 6.4 | 1 | 6.36 |
| SSw | 264.4 | 23 | 11.50 | 124.4 | 18 | 6.91 | 321.4 | 15 | 21.43 |
| SSt | 319.2 | 24 |  | 117.8 | 19 |  | 327.8 | 16 |  |
| F statistic | 4.8 |  |  | 0.9 |  |  | 0.3 |  |  |
| $\mathrm{F}_{0.05,1, \mathrm{SSwDF}}$ | 4.28 |  |  | 4.41 |  |  | 4.54 |  |  |
| Significant Difference | Yes |  |  | No |  |  | No |  |  |

Values in Red exceed the F Value and the data sets are significantly

Table A5-5

## River Reach Analysis of Variance

| Statistic | Carp MR/LR |  |  | Smallmouth Bass MR/LR |  |  | Rock Bass Bass MR/LR |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean Square |
| SSc | 219.5 | 1 | 219.53 | 3.1 | 1 | 3.07 | 0.0 | 1 | 0.03 |
| SSw | 2011.2 | 15 | 134.08 | 333.6 | 22 | 15.16 | 4.9 | 16 | 0.31 |
| SSt | 2230.7 | 16 |  | 336.6 | 23 |  | 5.0 | 17 |  |
| F statistic | 1.6 |  |  | 0.2 |  |  | 0.1 |  |  |
| $\mathrm{F}_{0.05,1, \mathrm{SSW} \text { WF }}$ | 4.54 |  |  | 4.3 |  |  | 4.49 |  |  |
| Significant Difference | No |  |  | No |  |  | No |  |  |

Values in Red exceed the F Value and the data sets are significantly

Table A5-5
River Reach Analysis of Variance

| Statistic | Catfish MR/LR |  |  | White Sucker LR/IH |  |  | Carp LR/IH |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square | SS | DF | Mean Square |
| SSc | 49.0 | 1 | 49.00 | Species not collected in a reach; can not calculate. |  |  | 263.9 | 1 | 263.94 |
| SSw | 1937.2 | 10 | 193.72 |  |  |  | 1671.1 | 14 | 119.36 |
| SSt | 1986.2 | 11 |  |  |  |  | 1935.0 | 15 |  |
| F statistic | 0.3 |  |  |  |  |  | 2.2 |  |  |
| $\mathrm{F}_{0.05,1, \text { SSwDF }}$ | 4.96 |  |  |  |  |  | 4.6 |  |  |
| Significant Difference | No |  |  |  |  |  |  | No |  |

Values in Red exceed
the F Value and the data
sets are significantly

Table A5-5

## River Reach Analysis of Variance

| Statistic | Smallmouth Bass LR/IH |  |  | Rock Bass LR/IH |  |  | Catfish LR/IH |  |  | Longnose Dace |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square |
| SSc | 23.2 | 1 | 23.17 | Species not collected in a reach; can not calculate. |  |  | 25.6 | 1 | 25.63 | Species not collected in a reach; can not calculate. |  |  |
| SSw | 72.6 | 14 | 5.19 |  |  |  | 301.0 | 3 | 100.32 |  |  |  |
| SSt | 95.8 | 15 |  |  |  |  | 326.6 | 4 |  |  |  |  |
| F statistic | 4.5 |  |  |  |  |  | 0.3 |  |  |  |  |  |
| $\mathrm{F}_{0.05,1, \mathrm{SSwDF}}$ | 4.6 |  |  |  |  |  | 10.1 |  |  |  |  |  |
| Significant Difference | No |  |  |  |  |  |  | No |  |  |  |  |

Values in Red exceed
the F Value and the data
sets are significantly

Table A5-6
River Reach Analysis of Population
River Reaches Mann Whitney Test
LR IH
Smallmouth Bass

| 8.17 | 1.44 |
| :---: | :---: |
| 5.14 | 2.70 |
| 2.02 | 4.43 |
| 1.78 | 3.10 |
| 7.01 | 4.18 |
| 4.84 | 4.31 |
| 10.9 | 3.91 |
| 6.30 | 2.83 |

Results in $\mathrm{mg} / \mathrm{Kg}$

| $\mathbf{n}_{\mathbf{1}}$ | $\mathbf{n}_{\mathbf{2}}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| :---: | :---: | :---: | :---: | :---: |
| 8 | 8 | 50 | 0.0649572 | 0.0324786 |
| normal approx <br> $\mathrm{z}=1.89038$ |  | $0.0587074^{*}$ | $0.0293537 *$ |  |

*These values are approximate.
The two samples are not significantly different ( $P>=0.05$, two-tailed test).

| LR |  |
| :---: | :---: |
| Catfish |  |
| 8.49 19.4 <br> 11.7  <br> 6.37  <br> 28.4  |  |

Results in $\mathrm{mg} / \mathrm{Kg}$

| $\mathbf{n}_{\mathbf{1}}$ | $\mathbf{n}_{\mathbf{2}}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| :---: | :---: | :---: | :---: | :---: |
| 4 | 1 | 3 | 0.8 | 0.4 |
| normal approx <br> $\mathrm{z}=0.707107$ |  |  | $0.4795^{*}$ | $0.23975^{*}$ |

The two samples are not significantly different $(P>=0.05$, two-tailed test).

Table A5-7
Fish Species Analysis of Means

| Statistic | Carp |  | Sucker |  | Small Mouth Bass |  | Rock Bass |  | Dace |  | Catfish |  | Walleye |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Carp | All* | Sucker | All ${ }^{\text {* }}$ | SM Bass | All ${ }^{\text {* }}$ | Rock Bass | All $^{*}$ | Dace | All ${ }^{\text {* }}$ | Catfish | All ${ }^{*}$ | Walleye | All* |
| Mean | 13.43 | 7.94 | 8.25 | 9.58 | 8.28 | 9.69 | 3.99 | 10.40 | 8.54 | 9.46 | 15.85 | 8.98 | 8.65 | 9.41 |
| Standard Deviation | 17.07 | 6.87 | 5.17 | 11.40 | 7.08 | 11.52 | 3.29 | 11.28 | 4.39 | 11.13 | 12.96 | 10.43 | 5.76 | 10.89 |
| Count | 54 | 158 | 33 | 179 | 48 | 164 | 34 | 178 | 20 | 192 | 12 | 200 | 11 | 189 |
| t | 2.30 |  | 1.07 |  | 1.04 |  | 6.30 |  | 0.72 |  | 1.80 |  | 0.40 |  |
| Significant Difference | Yes |  | No |  | No |  | Yes |  | No |  | Yes |  | No |  |

Mean and Standard Deviation in $\mathrm{mg} / \mathrm{Kg}$
Critcal Value at $t_{0.12}=1.64$
Values in Red exceed the Critical Value and the means for the data sets are significantly different.

*     - Excluding the fish species being compared.

Table A5-8
Fish Species Analysis of Variance

| Source | Adult Carp/All Fish |  |  | Adult Sucker/All Fish |  |  | SM Bass/All Fish |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean <br> Square |
| SSc | 1488.5 | 1 | 1488.5 | 79.8 | 1 | 79.8 | 122.2 | 1 | 122.2 |
| SSw | 10056.8 | 214 | 47.0 | 452.7 | 214 | 2.1 | 1293.7 | 214 | 6.0 |
| SSt | 11156.9 | 215 |  | 144.2 | 215 |  | 1559.8 | 215 |  |
| F | 31.7 |  |  | 37.7 |  |  | 20.2 |  |  |
| $\mathrm{F}_{0.05,1, \text { SSWDF }}$ | 3.89 |  |  | 3.89 |  |  | 3.89 |  |  |
| Significant Difference | Yes |  |  | Yes |  |  | Yes |  |  |

Values in Red exceed the $F$
Value and the data sets are
significantly different.

Table A5-8
Fish Species Analysis of Variance

| Source | Rock Bass/All Fish |  |  | Long Nose Dace/All Fish |  |  | Adult Catfish/All Fish |  |  | Adult Walleye/All Fish |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean <br> Square | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean Square |
| SSc | 1308.0 | 1 | 1308.0 | 28.7 | 1 | 28.7 | 702.8 | 1 | 702.8 | 12.4 | 1 | 12.4 |
| SSw | 557.8 | 214 | 2.6 | 1225.1 | 214 | 5.7 | 3517.1 | 214 | 16.4 | 1715.9 | 214 | 8.0 |
| SSt | 1477.4 | 215 |  | 865.4 | 215 |  | 3202.7 | 215 |  | 1728.3 | 215 |  |
| F | 501.9 |  |  | 5.0 |  |  | 42.8 |  |  | 1.5 |  |  |
| $\mathrm{F}_{0.05,1, \mathrm{SSWDF}}$ | 3.89 |  |  | 3.89 |  |  | 3.89 |  |  | 3.89 |  |  |
| Significant Difference | Yes |  |  | Yes |  |  | Yes |  |  | No |  |  |

Values in Red exceed the $F$
Value and the data sets are significantly different.

Table A5-9
Fish Species Analysis of Population
Fish Species Mann Whitney Test
Smallmouth Bass

| 18.6 | 21.5 | 15.2 | 22.2 | 7.3 | 6.1 | 8.6 | 4.1 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 28.9 | 5.3 | 14.9 | 33.5 | 3.1 | 6.4 | 13.5 | 10.5 |
| 14.1 | 6.0 | 5.8 | 4.2 | 7.5 | 9.3 | 18.2 | 5.0 |
| 3.5 | 7.7 | 5.5 | 2.6 | 3.7 | 3.1 | 4.3 | 4.1 |
| 8.2 | 5.1 | 2.0 | 1.8 | 7.0 | 4.8 | 10.9 | 6.3 |
| 1.4 | 2.7 | 4.4 | 3.1 | 4.2 | 4.3 | 3.9 | 2.8 |

## All Other Species

| 37.0 | 73.1 | 1.6 | 7.4 | 4.8 | 14.0 | 17.6 | 2.1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 53.9 | 28.4 | 9.5 | 29.4 | 33.3 | 9.6 | 55.5 | 36.9 |
| 34.5 | 5.1 | 3.2 | 7.8 | 3.7 | 30.2 | 9.2 | 22.7 |
| 3.6 | 1.7 | 47.7 | 10.5 | 1.0 | 15.8 | 1.4 | 37.3 |
| 2.1 | 1.7 | 1.3 | 2.5 | 22.8 | 1.6 | 1.3 | 2.2 |
| 1.3 | 2.5 | 15.7 | 0.5 | 44.9 | 18.4 | 4.5 | 2.0 |
| 1.9 | 3.2 | 2.5 | 5.0 | 9.1 | 2.3 | 2.1 | 0.9 |
| 0.2 | 15.9 | 16.6 | 10.3 | 20.6 | 10.6 | 5.7 | 7.3 |
| 12.3 | 10.8 | 12.0 | 5.0 | 9.4 | 4.0 | 16.6 | 6.0 |
| 7.5 | 3.7 | 11.8 | 3.2 | 19.9 | 8.8 | 4.7 | 9.2 |
| 3.2 | 2.4 | 3.5 | 3.5 | 0.9 | 6.4 | 7.0 | 4.8 |
| 5.0 | 3.7 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
| Smallmouth Bass |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
|  |  | 6.5 | 5.8 | 16.8 | 10.4 | 7.9 | 1.2 |
| 1.6 | 5.3 | 1.0 | 4.2 | 8.3 | 8.7 | 4.3 | 3.8 |
| 3.0 | 0.7 | 2.8 | 1.4 | 2.1 | 1.9 | 3.5 | 2.9 |
| 3.7 | 2.3 | 2.2 | 1.8 | 2.0 | 1.4 | 4.1 | 3.3 |
| 1.8 | 1.6 | 4.3 | 3.1 | 17.6 | 3.2 | 1.7 | 3.3 |
| 15.1 | 5.1 | 17.8 | 8.4 | 8.9 | 7.1 | 7.1 | 7.6 |
| 6.2 | 9.6 | 10.9 | 11.0 | 4.9 | 7.2 | 9.9 | 8.5 |
| 15.9 | 49.2 | 29.8 | 16.6 | 6.9 | 8.7 | 16.6 | 0.5 |
| 8.5 | 11.7 | 6.4 | 28.4 | 19.4 | 16.8 | 16.3 | 5.6 |
| 13.7 | 7.9 | 14.3 | 6.0 | 8.4 | 3.0 | 1.4 | 1.7 |

Results in $\mathrm{mg} / \mathrm{Kg}$

| $\mathbf{n}_{\mathbf{1}}$ | $\mathbf{n}_{\mathbf{2}}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| :---: | :---: | :---: | :---: | :---: |
| 168 | 48 | 4140 | $0.778912^{*}$ | $0.389456^{*}$ |
| normal approx <br> $\mathrm{z}=0.282819$ |  |  | $0.777316^{*}$ | $0.388658^{*}$ |

*These values are approximate.
The two samples are not significantly different $(P>=0.05$, two-tailed test).

Table A5-9
Fish Species Analysis of Population
Fish Species Mann Whitney Test
Dace

| 17.6 | 3.2 | 1.7 | 3.3 | 15.1 | 5.1 | 17.8 | 8.4 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 8.9 | 7.1 | 7.1 | 7.6 | 6.2 | 9.6 | 10.9 | 11.0 |
| 4.9 | 7.2 | 9.9 | 8.5 |  |  |  |  |

All Other Species

| 37.0 | 73.1 | 1.6 | 7.4 | 4.8 | 14.0 | 17.6 | 2.1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 53.9 | 28.4 | 9.5 | 29.4 | 33.3 | 9.6 | 55.5 | 36.9 |
| 34.5 | 5.1 | 3.2 | 7.8 | 3.7 | 30.2 | 9.2 | 22.7 |
| 3.6 | 1.7 | 47.7 | 10.5 | 1.0 | 15.8 | 1.4 | 37.3 |
| 2.1 | 1.7 | 1.3 | 2.5 | 22.8 | 1.6 | 1.3 | 2.2 |
| 1.3 | 2.5 | 15.7 | 0.5 | 44.9 | 18.4 | 4.5 | 2.0 |
| 1.9 | 3.2 | 2.5 | 5.0 | 9.1 | 2.3 | 2.1 | 0.9 |
| 0.2 | 15.9 | 16.6 | 10.3 | 20.6 | 10.6 | 5.7 | 7.3 |
| 12.3 | 10.8 | 12.0 | 5.0 | 9.4 | 4.0 | 16.6 | 6.0 |
| 7.5 | 3.7 | 11.8 | 3.2 | 19.9 | 8.8 | 4.7 | 9.2 |
| 3.2 | 2.4 | 3.5 | 3.5 | 0.9 | 6.4 | 7.0 | 4.8 |
| 5.0 | 3.7 | 18.6 | 21.5 | 15.2 | 22.2 | 7.3 | 6.1 |
| 8.6 | 4.1 | 28.9 | 5.3 | 14.9 | 33.5 | 3.1 | 6.4 |
| 13.5 | 10.5 | 14.1 | 6.0 | 5.8 | 4.2 | 7.5 | 9.3 |
| 18.2 | 5.0 | 3.5 | 7.7 | 5.5 | 2.6 | 3.7 | 3.1 |
| 4.3 | 4.1 | 8.2 | 5.1 | 2.0 | 1.8 | 7.0 | 4.8 |
| 10.9 | 6.3 | 1.4 | 2.7 | 4.4 | 3.1 | 4.2 | 4.3 |
| 3.9 | 2.8 | 6.5 | 5.8 | 16.8 | 10.4 | 7.9 | 1.2 |
| 1.6 | 5.3 | 1.0 | 4.2 | 8.3 | 8.7 | 4.3 | 3.8 |
| 3.0 | 0.7 | 2.8 | 1.4 | 2.1 | 1.9 | 3.5 | 2.9 |
| 3.7 | 2.3 | 2.2 | 1.8 | 2.0 | 1.4 | 4.1 | 3.3 |
| 1.8 | 1.6 | 4.3 | 3.1 |  |  |  |  |
| Dace |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
| 15.9 | 49.2 | 29.8 | 16.6 | 6.9 | 8.7 | 16.6 | 0.5 |
| 8.5 | 11.7 | 6.4 | 28.4 | 19.4 | 16.8 | 16.3 | 5.6 |
| 13.7 | 7.9 | 14.3 | 6.0 | 8.4 | 3.0 | 1.4 | 1.7 |

Results in $\mathrm{mg} / \mathrm{Kg}$

| $\mathbf{n}_{\mathbf{1}}$ | $\mathbf{n}_{\mathbf{2}}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| :---: | :---: | :---: | :---: | :---: |
| 196 | 20 | 2325 | $0.17171^{*}$ | $0.085855^{*}$ |
| normal approx <br> $z=1.37092$ |  |  | $0.1704014^{*}$ | $0.0852007^{*}$ |

*These values are approximate.
The two samples are not significantly different $(P>=0.05$, two-tailed test).

Table A5-9
Fish Species Analysis of Population
Fish Species Mann Whitney Test
Suckers

| 15.9 | 16.6 | 10.3 | 20.6 | 10.6 | 5.7 | 7.3 | 12.3 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 10.8 | 12.0 | 5.0 | 9.4 | 4.0 | 16.6 | 6.0 | 7.5 |
| 3.7 | 11.8 | 3.2 | 19.9 | 8.8 | 4.7 | 9.2 | 3.2 |
| 2.4 | 3.5 | 3.5 | 0.9 | 6.4 | 7.0 | 4.8 | 5.0 |
| 3.7 |  |  |  |  |  |  |  |

All Other Species

| 37.0 | 73.1 | 1.6 | 7.4 | 4.8 | 14.0 | 17.6 | 2.1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 53.9 | 28.4 | 9.5 | 29.4 | 33.3 | 9.6 | 55.5 | 36.9 |
| 34.5 | 5.1 | 3.2 | 7.8 | 3.7 | 30.2 | 9.2 | 22.7 |
| 3.6 | 1.7 | 47.7 | 10.5 | 1.0 | 15.8 | 1.4 | 37.3 |
| 2.1 | 1.7 | 1.3 | 2.5 | 22.8 | 1.6 | 1.3 | 2.2 |
| 1.3 | 2.5 | 15.7 | 0.5 | 44.9 | 18.4 | 4.5 | 2.0 |
| 1.9 | 3.2 | 2.5 | 5.0 | 9.1 | 2.3 | 2.1 | 0.9 |
| 0.2 |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
| White Suckers |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
|  |  | 18.6 | 21.5 | 15.2 | 22.2 | 7.3 | 6.1 |
| 8.6 | 4.1 | 28.9 | 5.3 | 14.9 | 33.5 | 3.1 | 6.4 |
| 13.5 | 10.5 | 14.1 | 6.0 | 5.8 | 4.2 | 7.5 | 9.3 |
| 18.2 | 5.0 | 3.5 | 7.7 | 5.5 | 2.6 | 3.7 | 3.1 |
| 4.3 | 4.1 | 8.2 | 5.1 | 2.0 | 1.8 | 7.0 | 4.8 |
| 10.9 | 6.3 | 1.4 | 2.7 | 4.4 | 3.1 | 4.2 | 4.3 |
| 3.9 | 2.8 | 6.5 | 5.8 | 16.8 | 10.4 | 7.9 | 1.2 |
| 1.6 | 5.3 | 1.0 | 4.2 | 8.3 | 8.7 | 4.3 | 3.8 |
| 3.0 | 0.7 | 2.8 | 1.4 | 2.1 | 1.9 | 3.5 | 2.9 |
| 3.7 | 2.3 | 2.2 | 1.8 | 2.0 | 1.4 | 4.1 | 3.3 |
| 1.8 | 1.6 | 4.3 | 3.1 | 17.6 | 3.2 | 1.7 | 3.3 |
| 15.1 | 5.1 | 17.8 | 8.4 | 8.9 | 7.1 | 7.1 | 7.6 |
| 6.2 | 9.6 | 10.9 | 11.0 | 4.9 | 7.2 | 9.9 | 8.5 |
| 15.9 | 49.2 | 29.8 | 16.6 | 6.9 | 8.7 | 16.6 | 0.5 |
| 8.5 | 11.7 | 6.4 | 28.4 | 19.4 | 16.8 | 16.3 | 5.6 |
| 13.7 | 7.9 | 14.3 | 6.0 | 8.4 | 3.0 | 1.4 | 1.7 |

Results in $\mathrm{mg} / \mathrm{Kg}$

| $\mathbf{n}_{\mathbf{1}}$ | $\mathbf{n}_{\mathbf{2}}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| :---: | :---: | :---: | :---: | :---: |
| 183 | 33 | 3358.5 | $0.306272^{*}$ | $0.153136^{*}$ |
| normal approx <br> $\mathrm{z}=1.02584$ |  |  | $0.304968^{*}$ | $0.152484^{*}$ |

*These values are approximate.
The two samples are not significantly different ( $\mathbf{P}>=0.05$, two-tailed test).

Table A5-10
Baseline and Historical Data Analysis of Means

|  | Carp |  | Sucker |  | Small Mouth Bass |  | Rock Bass |  | Catfish |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2000 | 2008 | 2002 | 2008 | 2002 | 2008 | 2004 | 2008 | 2004 | 2008 |
| Upper River 1 |  |  |  |  |  |  |  |  |  |  |
| Mean | 16.4 | 25.9 | 2.7 | 12.4 | 2.14 | 13 | 3.5 | 6.94 | 7.5 | 号 |
| Standard Deviation | 15.32 | 21.4 | 0.98 | 5 | 0.76 | 7.28 | 2.02 | 5.01 | 0.566 |  |
| Count | 6 | 16 | 25 | 8 | 11 | 8 | 3 | 8 | 2 |  |
| t | 1.15 |  | 5.45 |  | 4.20 |  | 1.62 |  | Not Applicable |  |
| Critcal Value at $\mathrm{t}_{0.1 / 2}$ | 1.75 |  | 1.71 |  | 1.80 |  | 1.86 |  |  |  |  |
| Significant Difference | No |  | Yes |  | Yes |  | No |  |  |  |  |
| - |  |  |  |  |  |  |  |  |  |  |
|  | 2004 | 2008 | 2004 | 2008 | 2004 | 2008 | 2004 | 2008 | 2004 | 2008 |
| Upper River 2 |  |  |  |  |  |  |  |  |  |  |
| Mean | 12.5 | 14.7 | 4.6 | 8.92 | 2.7 | 14.5 | 0.906 | 4.27 | 7.5 | 号 |
| Standard Deviation | 6.36 | 15 | 3 | 4.19 | 1.31 | 11.1 | 0.231 | 2.94 | 0.57 |  |
| Count | 2 | 16 | 5 | 8 | 6 | 8 | 5 | 8 | 2 |  |
| $t$ | 0.38 |  | 2.16 |  | 2.98 |  | 3.22 |  | Not Applicable |  |
| Critcal Value at $\mathrm{t}_{0.1 / 2}$ | 1.75 |  | 1.86 |  | 1.86 |  | 1.86 |  |  |  |  |
| Significant Difference | No |  | Yes |  | Yes |  | Yes |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  | 2004 | 2008 | 2004 | 2008 | 2004 | 2008 | 2004 | 2008 | 2004 | 2008 |
| Lower River |  |  |  |  |  |  |  |  |  |  |
| Mean | 2.32 | 11.3 | 2.5 | 4.31 | 1.3 | 5.77 | $\begin{aligned} & \text { D } \\ & \text { Z } \end{aligned}$ | 2.6 | 3.25 | 13.7 |
| Standard Deviation | 1.03 | 15.2 | 1.11 | 0.926 | 0.93 | 3.05 |  | 1.11 | 0.21 | 10 |
| Count | 5 | 9 | 5 | 2 | 5 | 8 |  | 8 | 2 | 4 |
| t | 5.09 |  | 0.82 |  | 1.16 |  | Not Applicable |  | 5.00 |  |
| Critcal Value at $\mathrm{t}_{0.1 / 2}$ | 1.86 |  | 2.01 |  | 1.86 |  |  |  | 2.35 |  |
| Significant Difference | Yes |  | No |  | No |  |  |  | Yes |  |

Mean and Standard Deviation in $\mathrm{mg} / \mathrm{Kg}$
Values in Red exceed the Critical Value and the means for the data sets are significantly different.

Table A5-11

## Baseline and Historical Data Analysis of Variance

| Statistic | Upper River 1 Suckers |  |  | Upper River 2 Suckers |  |  | Lower River Suckers |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean <br> Square | SS | DF | Mean Square | SS | DF | Mean <br> Square |
| SSc | 573.4 | 1 | 573.36 | 57.9 | 1 | 57.87 | 4.5 | 1 | 4.45 |
| SSw | 197.9 | 31 | 6.38 | 159.1 | 11 | 14.47 | 5.8 | 5 | 1.16 |
| SSt | 771.3 | 32 |  | 217.0 | 12 |  | 10.2 | 6 |  |
| F Statistic | 89.8 |  |  | 4.0 |  |  | 3.8 |  |  |
| $\mathrm{F}_{0.05,1, \mathrm{SSwDF}}$ | 4.16 |  |  | 4.84 |  |  | 6.61 |  |  |
| Significant Diffference | Yes |  |  | No |  |  | No |  |  |

Values in Red exceed the
F Value and the data sets
are significantly
different.

Table A5-11
Baseline and Historical Data Analysis of Variance

| Statistic | Upper River 1 Carp |  |  | Upper River 2 Carp |  |  | Lower River Carp |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean <br> Square | SS | DF | Mean <br> Square | SS | DF | Mean Square |
| SSc | 389.4 | 1 | 389.40 | 8.7 | 1 | 8.75 | No individual historical results. Can not calculate. |  |  |
| SSw | 2235.1 | 20 | 111.76 | 610.8 | 16 | 38.17 |  |  |  |
| SSt | 1845.7 | 21 |  | 602.0 | 17 |  |  |  |  |
| F Statistic | 3.5 |  |  | 0.2 |  |  |  |  |  |
| $\mathrm{F}_{0.05,1, \text { SSwDF }}$ | 4.35 |  |  | 4.49 |  |  |  |  |  |
| Significant Diffference | No |  |  | No |  |  |  |  |  |

Values in Red exceed the
F Value and the data sets
are significantly
different.

Table A5-11

## Baseline and Historical Data Analysis of Variance

| Statistic | Upper River 1 Small Mouth Bass |  |  | Upper River 2 Small Mouth Bass |  |  | Lower River Small Mouth Bass |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean Square |
| SSc | 542.2 | 1 | 542.22 | 476.8 | 1 | 476.82 | 60.2 | 1 | 60.17 |
| SSw | 377.0 | 17 | 22.17 | 873.0 | 12 | 72.75 | 68.6 | 11 | 6.24 |
| SSt | 919.2 | 18 |  | 1349.8 | 13 |  | 128.8 | 12 |  |
| F Statistic | 24.5 |  |  | 6.6 |  |  | 9.6 |  |  |
| $\mathrm{F}_{0.05,5, \text { SSwDF }}$ | 4.45 |  |  | 4.75 |  |  | 4.84 |  |  |
| Significant Diffference | Yes |  |  | Yes |  |  | Yes |  |  |

Values in Red exceed the
F Value and the data sets
are significantly
different.

Table A5-11
Baseline and Historical Data Analysis of Variance

| Statistic | Upper River 1 Rock Bass |  |  | Upper River 2 Rock Bass |  |  | Lower River Catfish |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean Square |
| SSc | 25.9 | 1 | 25.88 | No individual historical results. Can not calculate. |  |  | 146.7 | 1 | 146.72 |
| SSw | 184.0 | 9 | 20.44 |  |  |  | 301.0 | 4 | 75.25 |
| SSt | 209.9 | 10 |  |  |  |  | 447.7 | 5 |  |
| F Statistic | 1.3 |  |  |  |  |  | 1.9 |  |  |
| $\mathrm{F}_{0.05,1, \text { SSwDF }}$ | 5.12 |  |  |  |  |  | 7.71 |  |  |
| Significant Diffference | No |  |  |  |  |  |  | No |  |

Values in Red exceed the
F Value and the data sets
are significantly
different.

Table A5-12
Baseline and Historical Data Analysis of Population
Baseline \& Historical Mann Whitney Test
Upper River 2 Suckers

| Historical | Baseline |
| :---: | :---: |
| 2.2 | 10.8 |
| 4.3 | 12.03 |
| 2.6 | 5.04 |
| 4.1 | 9.44 |
| 9.7 | 3.95 |
|  | 16.6 |
|  | 5.95 |
|  | 7.52 |

Results in $\mathrm{mg} / \mathrm{Kg}$

| $\mathbf{n}_{1}$ | $\mathbf{n}_{\mathbf{2}}$ | U | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| :---: | :---: | :---: | :---: | :---: |
| $\mathbf{n}_{1}$ | $\mathbf{n}_{\mathbf{2}}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| 8 | 5 | 38 | 0.006216 | 0.003108 |
| normal approx <br> $\mathrm{z}=2.63493$ |  |  | $0.00841546^{*}$ | $0.00420773^{*}$ |

The two samples are not significantly different ( $P>=0.05$, two-tailed test).

| Lower River Smallmouth Bass |
| :--- |
| Historical |
| 1.7 Baseline |
| 2.8 |
| 0.17 |
| 0.86 |
| .14 |
| 0.45 |

Results in $\mathrm{mg} / \mathrm{Kg}$

| $\mathbf{n}_{1}$ | $\mathbf{n}_{2}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| :---: | :---: | :---: | :---: | :---: |
| $\mathbf{n}_{1}$ | $\mathbf{n}_{2}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| 8 | 5 | 33 | 0.065268 | 0.032634 |
| normal approx <br> $\mathrm{z}=1.90301$ |  |  | $0.0570398^{*}$ | $0.0285199^{*}$ |

The two samples are significantly different ( $\mathrm{P}<0.01$, two-tailed test).

Table A5-12
Baseline and Historical Data Analysis of Population

| Lower River Catfish <br> Historical |  |
| :--- | :---: |
| 3.4 8.49 <br> 3.1 11.7 <br>  6.37 <br>  28.4 |  |

Results in $\mathrm{mg} / \mathrm{Kg}$

| $\mathbf{n}_{\mathbf{1}}$ | $\mathbf{n}_{2}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| :---: | :---: | :---: | :---: | :---: |
| $\mathbf{n}_{1}$ | $\mathbf{n}_{2}$ | $\mathbf{U}$ | $\mathbf{P}$ (two- <br> tailed) | $\mathbf{P}$ (one- <br> tailed) |
| 4 | 2 | 8 | 0.1333334 | 0.0666667 |
| normal approx <br> $\mathrm{z}=1.85164$ |  |  | $0.0640776^{*}$ | $0.0320388^{*}$ |

The two samples are significantly different ( $\mathbf{P}<\mathbf{0 . 0 1}$, two-tailed test).

Table A5-13

## Baseline and Historical Data Analysis of Lipid Normalized Means



Mean and Standard Deviation in $\mathrm{mg} / \mathrm{Kg}$
Values in Red exceed the Critical Value and the means for the data sets are significantly different.

Table A5-14
Baseline and Historical Data Analysis of Lipid Normalized Variance

| Statistic | Upper River 1 Suckers |  |  | Upper River 2 Suckers |  |  | Lower River Suckers |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean <br> Square |
| SSc | 9777363.5 | 1 | 9777363 | 1364960.9 | 1 | 1364961 | 69848.3 | 1 | 69848.33 |
| SSw | 1269848.7 | 31 | 40963 | 867611.0 | 11 | 78874 | 925.0 | 5 | 185.00 |
| SSt | 11047212.2 | 32 |  | 2232571.9 | 12 |  | 70773.3 | 6 |  |
| F Statistic | 238.7 |  |  | 17.3 |  |  | 377.6 |  |  |
| $\mathrm{F}_{0.05,1, S \mathrm{SwDF}}$ | 4.16 |  |  | 4.84 |  |  | 6.61 |  |  |
| Significant Diffference | Yes |  |  | Yes |  |  | Yes |  |  |

Values in Red exceed the
F Value and the data sets
are significantly
different.

Table A5-14
Baseline and Historical Data Analysis of Lipid Normalized Variance

| Statistic | Upper River 1 Carp |  |  | Upper River 2 Carp |  |  | Lower River Carp |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean <br> Square |
| SSc | 331468.6 | 1 | 331469 | 1662.9 | 1 | 1662.94 | No individual historical results. Can not calculate. |  |  |
| SSw | 35474118.1 | 20 | 1773706 | 1518256.1 | 16 | 94891.01 |  |  |  |
| SSt | 35805586.7 | 21 |  | 1516593.2 | 17 |  |  |  |  |
| F Statistic | 0.2 |  |  | 0.02 |  |  |  |  |  |
| $\mathrm{F}_{0.05,1, S \mathrm{~S}_{w} \mathrm{DF}}$ | 4.35 |  |  | 4.49 |  |  |  |  |  |
| Significant Diffference | No |  |  | No |  |  |  |  |  |

Values in Red exceed the
$F$ Value and the data sets
are significantly
different.

Table A5-14
Baseline and Historical Data Analysis of Lipid Normalized Variance

| Statistic | Upper River 1 Small Mouth Bass |  |  | Upper River 2 Small Mouth Bass |  |  | Lower River Small Mouth Bass |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean Square |
| SSc | 15422004.4 | 1 | 15422004 | 1782151.3 | 1 | 1782151 | 60.2 | 1 | 60.17 |
| SSw | 24722887.9 | 17 | 1454288 | 2440220.2 | 12 | 203352 | 68.6 | 11 | 6.24 |
| SSt | 40144892.3 | 18 |  | 4222371.5 | 13 |  | 128.8 | 12 |  |
| F Statistic | 10.6 |  |  | 8.8 |  |  | 9.6 |  |  |
| $\mathrm{F}_{0.05,1, \mathrm{SSmDF}}$ | 4.45 |  |  | 4.75 |  |  | 4.84 |  |  |
| Significant Diffference | Yes |  |  | Yes |  |  | Yes |  |  |

Values in Red exceed the
F Value and the data sets
are significantly
different.

Table A5-14
Baseline and Historical Data Analysis of Lipid Normalized Variance

| Statistic | Upper River 1 Rock Bass |  |  | Upper River 2 Rock Bass |  |  | Lower River Catfish |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SS | DF | Mean Square | SS | DF | Mean Square | SS | DF | Mean Square |
| SSc | 314511.6 | 1 | 314512 | No individual historical results. Can not calculate. |  |  | 195072.0 | 1 | 195072 |
| SSw | 2950858.5 | 9 | 327873 |  |  |  | 520628.3 | 9 | 57848 |
| SSt | 3265370.1 | 10 |  |  |  |  | 325556.2 | 10 |  |
| F Statistic | 1.0 |  |  |  |  |  | 3.4 |  |  |
| $\mathrm{F}_{0.05,5, \mathrm{SSwDF}}$ | 5.12 |  |  |  |  |  | 10.1 |  |  |
| Significant Diffference | No |  |  |  |  |  | No |  |  |

Values in Red exceed the
F Value and the data sets
are significantly
different.

## Box and Whisker Plots

## Example of Box and Whisker Plots



Observed Data is PCB concentration in fish tissue in $(\mathrm{mg} / \mathrm{Kg})$

UR1 and UR2 - Upper River site 1 and Upper River site 2 MR1 and MR2 LR and IH .

Middle River site 1 and Middle River site 2 Lower River and Inner Harbor


Observed Data is PCB concentration in fish tissue in ( $\mathrm{mg} / \mathrm{Kg}$ ).
There was only one carp collected in Middle River 2.
For Middle River 1, minimum and 1st quartile values too close to differentiate on plot. Maximum and 4th quartile values were equal.

UR1 and UR2 -
MR1 and MR2 -
LR and IH -

Upper River site 1 and Upper River site 2 Middle River site 1 and Middle River site 2 Lower River and Inner Harbor


Observed Data is PCB concentration in fish tissue in ( $\mathrm{mg} / \mathrm{Kg}$ ).
No adult suckers collected in Inner Harbor.
Only two adult suckers collected in Lower River. No minimum and maximum shown on box plot.

UR1 and UR2 MR1 and MR2 -
LR -

Upper River site 1 and Upper River site 2 Middle River site 1 and Middle River site 2 Lower River

## Juvenile Sucker Box and Whisker Plots



Observed Data is PCB concentration in fish tissue in ( $\mathrm{mg} / \mathrm{Kg}$ ). No juvenile suckers collected in Middle River 1 and Inner Harbor.

UR1 and UR2 -
MR2 -
LR -

Upper River site 1 and Upper River site 2 Middle River site 2
Lower River

## Smallmouth Bass Box and Whisker Plots



Observed Data is PCB concentration in fish tissue in ( $\mathrm{mg} / \mathrm{Kg}$ ).

- Outlier identified during box and whisker plotting not identified during previous outlier analysis.

UR1 and UR2 -
MR1 and MR2 -
LR and IH -

Upper River site 1 and Upper River site 2 Middle River site 1 and Middle River site 2 Lower River and Inner Harbor

## Rock Bass Box and Whisker Plots



Observed Data is PCB concentration in fish tissue in ( $\mathrm{mg} / \mathrm{Kg}$ ).
No rock bass collected in Inner Harbor.
There was only one rock bass collected in Middle River 1.

UR1 and UR2 MR1 and MR2 -
LR -

Upper River site 1 and Upper River site 2
Middle River site 1 and Middle River site 2
Lower River

## Longnose Dace Box and Whisker Plots



Observed Data is PCB concentration in fish tissue in ( $\mathrm{mg} / \mathrm{Kg}$ ).
No dace collected in Upper River 2, Lower River, and Inner Harbor.
For Middle River 1, the minimum and 1st quartile values were the same.

## UR1 -

MR1 and MR2 -

Upper River site 1
Middle River site 1 and Middle River site 2

## Catfish Box and Whisker Plots



Observed Data is PCB concentration in fish tissue in ( $\mathrm{mg} / \mathrm{Kg}$ ). No catfish collected in Upper River sites.
There was only one catfish collected in Inner Harbor.
Only three Lower River catfish plotted. Minimum and 1st quartile values too close to differentiate on plot. Maximum and 4th quartile values were equal.

MR1 and MR2 -
LR and IH -

Middle River site 1 and Middle River site 2 Lower River and Inner Harbor

## Walleye Box and Whisker Plots



Observed Data is PCB concentration in fish tissue in $(\mathrm{mg} / \mathrm{Kg})$.
No adult walleyes collected in Inner Harbor.
Only three walleye collected in Inner Harbor. Minimum and 1st quartile values too close to differentiate on plot. Maximum and 4th quartile values were equal.

MRI -
H

Middle River site 1
Inner Harbor

## Appendix 6

## Phase 1 Sampling Requirements

Fish Needs Statistical Analysis Lower Fox River Method

## (Upper River)

| Upper River 1 Species/Location | $\boldsymbol{\alpha}$ | $\boldsymbol{\beta}$ | MDRD | $\mathbf{C} / \mathbf{V}$ | $\mathbf{N}$ | \# |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Smallmouth Bass | 0.1 | 0.2 | 0.5 | 0.56 | 6 | 8 |
| Adult Carp | 0.1 | 0.2 | 0.5 | 0.83 | 12 | 12 |
| Juvenile Carp | 0.1 | 0.2 | 0.5 | N/A | N/A | 16 |
| Adult Suckers | 0.1 | 0.2 | 0.5 | 0.4 | 3 | 8 |
| Juvenile Suckers | 0.1 | 0.2 | 0.5 | 0.47 | 4 | 8 |
| Rock Bass | 0.1 | 0.2 | 0.5 | 0.72 | 9 | 8 |
| Longnose Dace | 0.1 | 0.2 | 0.5 | 0.89 | 14 | 8 |
| Walleye | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |
| Catfish | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |


| Upper River 2 Specles/Location | $\boldsymbol{\alpha}$ | $\boldsymbol{\beta}$ | MDRD | $\mathbf{C} / \mathbf{V}$ | $\mathbf{N}$ | \# |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Smallmouth Bass | 0.1 | 0.2 | 0.5 | 0.77 | 11 | 8 |
| Adult Carp | 0.1 | 0.2 | 0.5 | 1.02 | 19 | 16 |
| Juvenile Carp | 0.1 | 0.2 | 0.5 | $\mathrm{~N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ | 16 |
| Adult Suckers | 0.1 | 0.2 | 0.5 | 0.47 | 4 | 8 |
| Juvenile Suckers | 0.1 | 0.2 | 0.5 | 0.43 | 3 | 8 |
| Rock Bass | 0.1 | 0.2 | 0.5 | 0.69 | 9 | 8 |
| Longnose Dace | 0.1 | 0.2 | 0.5 | $\mathrm{~N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ | 8 |
| Walleye | 0.1 | 0.2 | 0.5 | $\mathrm{~N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ | 8 |
| Catfish | 0.1 | 0.2 | 0.5 | $\mathrm{~N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ | 8 |

Fish Needs Statistical Analysis Lower Fox River Method (Middle River)

| Middle River 1 Species/Location | $\boldsymbol{\alpha}$ | $\boldsymbol{\beta}$ | MDRD | $\mathbf{C} / \mathbf{V}$ | $\mathbf{N}$ | \# |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Smallmouth Bass | 0.1 | 0.2 | 0.5 | 0.56 | 6 | 8 |
| Adult Carp | 0.1 | 0.2 | 0.5 | 1.67 | 50 | 8 |
| Juvenile Carp | 0.1 | 0.2 | 0.5 | $\mathrm{~N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ | 8 |
| Adult Suckers | 0.1 | 0.2 | 0.5 | 0.669 | 8 | 8 |
| Juvenile Suckers | 0.1 | 0.2 | 0.5 | 0.47 | 4 | 8 |
| Rock Bass | 0.1 | 0.2 | 0.5 | $\mathrm{~N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ | 8 |
| Longnose Dace | 0.1 | 0.2 | 0.5 | 0.438 | 3 | 8 |
| Walleye | 0.1 | 0.2 | 0.5 | 0.42 | 3 | 8 |
| Catfish | 0.1 | 0.2 | 0.5 | 0.56 | 6 | 8 |


| Middle River 2 Species/Location | $\mathbf{a}$ | $\beta$ | MDRD | C/V | $\mathbf{N}$ | \# |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Smallmouth Bass | 0.1 | 0.2 | 0.5 | 0.37 | 2 | 8 |
| Adult Carp | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |
| Juvenile Carp | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |
| Adult Suckers | 0.1 | 0.2 | 0.5 | 0.51 | 5 | 8 |
| Juvenile Suckers | 0.1 | 0.2 | 0.5 | 0.28 | 1 | 8 |
| Rock Bass | 0.1 | 0.2 | 0.5 | 0.32 | 2 | 8 |
| Longnose Dace | 0.1 | 0.2 | 0.5 | 0.26 | N/A | 8 |
| Walleye | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |
| Catfish | 0.1 | 0.2 | 0.5 | 0.81 | N/A | 8 |

Fish Needs Statistical Analysis Lower Fox River Method
(Lower River Inner Harbor)

| Lower River Species/Location | $\boldsymbol{a}$ | $\beta$ | MDRD | $\mathbf{C} / \mathbf{V}$ | $\mathbf{N}$ | \# |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Smallmouth Bass | 0.1 | 0.2 | 0.5 | 0.53 | 5 | 8 |
| Adult Carp | 0.1 | 0.2 | 0.5 | 1.35 | 33 | 8 |
| Juvenile Carp | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |
| Adult Suckers | 0.1 | 0.2 | 0.5 | 0.22 | 1 | 8 |
| Juvenile Suckers | 0.1 | 0.2 | 0.5 | 0.41 | 3 | 8 |
| Rock Bass | 0.1 | 0.2 | 0.5 | 0.43 | 3 | 8 |
| Longnose Dace | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |
| Walleye | 0.1 | 0.2 | 0.5 | N/A | N/A | 9 |
| Catfish | 0.1 | 0.2 | 0.5 | 0.73 | 10 | 8 |


| Inner Harbor Species/Location | $\boldsymbol{a}$ | $\boldsymbol{\beta}$ | MDRD | $\mathbf{C} / \mathbf{V}$ | $\mathbf{N}$ | \# |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Smallmouth Bass | 0.1 | 0.2 | 0.5 | 0.31 | 2 | 8 |
| Adult Carp | 0.1 | 0.2 | 0.5 | 0.89 | 14 | 8 |
| Juvenile Carp | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |
| Adult Suckers | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |
| Juvenile Suckers | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |
| Rock Bass | 0.1 | 0.2 | 0.5 | N/A | N/A | 9 |
| Longnose Dace | 0.1 | 0.2 | 0.5 | N/A | N/A | 8 |
| Walleye | 0.1 | 0.2 | 0.5 | 0.42 | 3 | 8 |
| Catfish | 0.1 | 0.2 | 0.5 | 0.3 | 2 | 8 |

## Appendix 7

## Foth Multiple Regression Analysis

# Foth Infrastructure \& Environment, LLC Memorandum 

November 11, 2008

TO: Keith Egan, PRS; Ken Aukerman, PRS
CC: Steve Laszewski, Foth
FR: Steve Lehrke, Foth
RE: Analysis of Sheboygan River Fish Tissue Covariates

## Background

Fish tissue PCB sample results collected during August and September of 2008 were received by Foth Infrastructure \& Environment, LLC (Foth) from PRS for the purpose of conducting a multiple regression analysis. This data is included as Attachment 1. The analysis was performed to develop preliminary conclusions on the effectiveness of including covariates in future statistical tests for determining trends in fish tissue PCB levels. The covariates under consideration are fish length and percent lipids. Including these covariates in future statistical tests could potentially remove additional variation (or noise) from the data and allow a clearer determination to be made of fish tissue PCB concentration trends.

## Covariate Approach (Future Analysis)

A statistical method of determining significant changes in fish tissue PCB concentrations between baseline and post-remediation results is to utilize multiple regression analysis. In a multiple regression model, the covariates of fish length and percent lipids could be included to more easily detect changes between the baseline and post-remediation concentration levels. Possible models include a linear model of the form:

$$
\begin{equation*}
P C B=B_{0}+B_{1} \text { Length }+B_{2} \text { Lipids }+B_{3} \text { Remediation } \tag{Equation1}
\end{equation*}
$$

and an exponential model of the form:

$$
\begin{equation*}
P C B=e^{B 0+\text { B1Length }+ \text { B2Lipids }+ \text { B3Remediation }} . \tag{Equation2}
\end{equation*}
$$

In both models Remediation is an indicator variable taking on a value of 0 for baseline data and 1 for post-remediation data. A test of the effect of remediation on average PCB concentrations could then be constructed as

$$
H_{0}: B_{3} \geq 0 \text { v.s. } H_{A}: B_{3}<0 .
$$

If the test is significant, that is the coefficient $B_{3}$ is significantly less than 0 , the conclusion is made that remediation on average has reduced PCB concentration levels.

## Results of Current Data

The data included in Attachment 1 was utilized in various multiple regression analyses to verify if there was potential use in including fish length and percent lipids as covariates. Data sets were included for two sites in the Upper River, two sites in the Middle River, one site in the Lower River and one site in the Inner Harbor. Fish Types include adult carp, adult suckers, juvenile suckers, smallmouth bass, rock bass, longnose dace, walleye and catfish.

To do this, linear multiple regression models were developed in the form of:

$$
\begin{equation*}
P C B=B_{0}+B_{1} \text { Length }+B_{2} \text { Lipids } \tag{Equation3}
\end{equation*}
$$

and exponential models in the form of:

$$
\begin{equation*}
P C B=e^{B 0+B 1 L e n g t h+B 2 L i p i d s} . \tag{Equation4}
\end{equation*}
$$

In order for the covariates to be useful in removing noise in the PCB data, they need to "explain" a significant amount of variation in the PCB concentrations. The results of the multiple regression models in equations 3 and 4 provide metrics which are useful in determining how much variation is being explained by the covariates. These metrics include:

- $\mathrm{R}^{2}$ :

Provides a measure of how much variation in the PCB data is being explained by the entire model. Values fall between 0 and 1 , with a value of 0 implying no variation is explained and a value of 1 implying all the variation is explained.

- Coefficients $B_{1}$ and $B_{2}$ :

In the linear model, estimates of these indicate the proportional change in PCB concentrations for a unit change in the corresponding covariate. In the exponential model the estimates indicate the proportional change in the logarithm of the PCB concentrations.

- Standard Errors of $B_{1}$ and $B_{2}$ estimates:

The standard errors indicate how much variability can be expected in the estimates of $B_{1}$ and $B_{2}$.

- p-Level:

The corresponding p-level indicates the probability of a coefficient ( $B_{1}$ or $B_{2}$ ) being equal to zero. If a coefficient is significantly different from zero, the corresponding factor (Length or Lipids) has a significant impact on PCB concentrations. A p-level of less than 0.1 indicates the coefficient is significantly different from zero at a $10 \%$ error rate, and a p-level of less than 0.05 indicates significance at a $5 \%$ error rate.

The results of these metrics are given in Attachment 2 (Table 4). For each data set the model (linear or exponential) was chosen which gave the highest $\mathrm{R}^{2}$ value. There are several data sets for which the coefficients corresponding to length and lipids are significantly different from zero, which indicates these factors significantly affect tissue PCB concentrations. The data sets illustrating significance with these factors are as follows:

Upper River 1:

- Adult Carp (length only)
- Adult Suckers (lipids only)
- Longnose Dace (length only)


## Upper River 2:

- Adult Carp (length and lipids)
- Adult Suckers (length and lipids)
- Smallmouth Bass (lipids)
- Rock Bass (length and lipids)


## Middle River 1:

- Adult Carp (length)
- Walleye (lipids)


## Middle River 2:

- Rock Bass (lipids)
- Longnose Dace (length and lipids)

Lower River

- Adult Carp (lipids)
- Juvenile Suckers (length)
- Smallmouth Bass (lipids)
- Rock Bass (lipids)

Inner Harbor

- Adult Carp (lipids)
- Smallmouth Bass (lipids)

Since the coefficients for either length, lipids or both length and lipids were significantly different from zero in the above data sets, these are likely good factors to include in the covariate approach described above.

Note that in the above data sets several coefficients for length were negative (Attachment 2). This was the case for adult suckers and rock bass in the Upper River 2 data set, and smallmouth bass in the Inner Harbor data set. In these three cases length had an inverse effect on tissue PCB concentrations, meaning that the larger fish had lower concentrations.

## Conclusions

In summary, fish tissue PCB sample results collected during August and September of 2008 (Attachment 1) were analyzed by multiple regression techniques to determine the usefulness of including covariate measures of fish length and percent lipids in future analysis. The future analysis would use these covariates to reduce additional variation in the data so that conclusions concerning PCB concentration trends can be more readily made. Based on the results, the inclusion of fish length and percent lipids significantly reduced variation noise in several of the data sets as listed above.

## Attachment 1

## Fish Tissue PCB Data Sets

Upper River Fith Tissue Results - Site 1

| Adutt Carp Length | $\begin{aligned} & \text { Adult Corp } \\ & \text { \% Lipid } \end{aligned}$ | Adult Carp PCB | Ln Adult Carp PCE | Adult Whtte Sucker Length | Adult White Sus.ker \% Lipid | Adult Whtte Suckar PCB | In Adult White Surkar PCB | Juvanlle White Suckar Length | Juvenille White Sucker * <br> LIpid | Jivenila Whita Sucker $P C B$ | Ln Juvenilla White Sucker PCB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 24.0 | 4.60\% | 37.0 | 3.6 | 16.0 | 1.40\% | 15.9 | 2.8 | 6.00 | 0.151\% | 9.71 | 2.3 |
| 21.0 | 1.33\% | 73.1 | 4.3 | 14.0 | 1.33x | 16.6 | 2.8 | 6.00 | 0.367\% | 0.93 | 2.2 |
| 18.0 | 4.84\% | 1.63 | 0.5 | 13.0 | 0.555\% | 10.3 | 2.3 | 5.00 | 0.467\% | 6.08 | 1.8 |
| 19.0 | 4.45\% | 7.44 | 2.0 | 12.0 | 1.52\% | 20.6 | 3.0 | 6.00 | 0.249x | 4.85 | 1.6 |
| 15.0 | 2.19\% | 4.77 | 1.6 | 14.0 | 0.855\% | 10.6 | 2.4 | 7.00 | 0.330\% | 776 | 2.0 |
| 16.0 | 0.625\% | 14.0 | 2.6 | 12.0 | 0.495\% | 5.74 | 1.7 | 6.00 | 0.638\% | 6.51 | 1.9 |
| 20.0 | 2.50\% | 17.6 | 2.9 | 14.0 | 0.330x | 7.34 | 2.0 | 6.50 | 0.281\% | 2.28 | 0.1 |
| 19.5 | 0.390\% | 2.08 | 0.7 | 11.5 | 0.760x | 12.3 | 2.5 | 6.00 | 0.275\% | 1.99 | 0.7 |
| 25.0 | 7.49\% | 53.9 | 4.0 |  |  |  |  |  |  |  |  |
| 24.0 | 7.55\% | 28.4 | 3.3 |  |  |  |  |  |  |  |  |
| 21.0 | 3.44\% | 9.48 | 2.2 |  |  |  |  |  |  |  |  |
| 23.0 | 3.02\% | 29.4 | 3.4 |  |  |  |  |  |  |  |  |
| 25.0 | 13.7\% | 33.3 | 3.5 |  |  |  |  |  |  |  |  |
| 25.0 | 1.01\% | 9.55 | 2.3 |  |  |  |  |  |  |  |  |
| 22.5 | 8.70\% | 55.5 | 4.0 |  |  |  |  |  |  |  |  |
| 23.0 | 7.03\% | 36.9 | 3.6 |  |  |  |  |  |  |  |  |

Upper River Fish Tissua Results - Site 1

| Smallimouth Bass length | 5malimouth Base \% Lipid | 5mallmouth Bass PCB | In Smallmouth Bass PCB | Rock Bass ${ }^{4}$ length | rock fans \% Lipid | Rock Basa PCB | In Rock Buas PCB | Longmose Dice Length | $\begin{gathered} \hline \text { Longnose Dace } \\ \text { \% Lipid } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Longnose Daca } \\ \text { PCB } \\ \hline \end{gathered}$ | Longnose Dace PCB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 13.0 | 0.625\% | 18.5 | 2.9 | 8.50 | 0.415\% | 6.53 | 1.9 | 3.00 | 2.77\% | 17.6 | 2.9 |
| 10.0 | 0.400\% | 21.5 | 3.1 | B.00 | 0.590\% | 5.82 | 1.8 | 2.50 | 1.24\% | 3.20 | 1.2 |
| 15.0 | 1.43\% | 15.2 | 2.7 | 5.50 | 0.775\% | 16.8 | 2.8 | 2.00 | 1.14\% | 1.72 | 0.5 |
| 10.0 | 0.490\% | 22.2 | 3.1 | 6.00 | 102\% | 10.4 | 2.3 | 2.50 | 2.30\% | 3.29 | 1.2 |
| 10.0 | 0.695\% | 7.33 | 2.0 | 6.00 | 0.581\% | 7.91 | 2.1 | 3.50 | 4.00\% | 15.1 | 2.7 |
| 11.0 | 0.765\% | 6.14 | 1.8 | 7.00 | 0.325\% | 1.22 | 0.2 | 2.50 | 4.40\% | 5.11 | 1.6 |
| 14.0 | 117\% | 8,59 | 2.2 | 8.00 | 0.485* | 1.57 | 0.5 |  |  |  |  |
| 10.0 | 0.430\% | 4.09 | 1.4 | 5.50 | 0.619\% | 5,30 | 1.7 |  |  |  |  |

## Upper River Fish Tissue Results - Site 2

| Aduht Carp length | $\begin{aligned} & \text { Adult Curp } \\ & \times \text { Lipid } \end{aligned}$ | Adutt Carp PCB | In Adult Carp PCI | Adult White Surkar Length | Adult White Sucker \% Lipid | Adult White Sucker PCP | In Adult White Sucker PCB | Juwanile Whte Sucker Length | Juvanile White Sucker \% Lipid | duvenlle White 5ucker Pre | in Juventle White Sucker PCD |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 21.0 | 7.39\% | 34.5 | 3.5 | 11.0 | 0.960\% | 10.8 | 2.4 | 6,00 | 0.510\% | 4.39 | 1.5 |
| 23.0 | 2.05x | 5.14 | 1.6 | 13.0 | 1.37\% | 12.0 | 2.5 | 7.00 | 0.450\% | 11.5 | 2.4 |
| 18.0 | 3.99\% | 3.18 | 1.2 | 14.0 | 1.14x | 5.04 | 1.6 | 6.00 | 0.580\% | 5.71 | 1.7 |
| 15.0 | 4.64\% | 7.84 | 2.1 | 9.00 | 0.715x | 9.44 | 2.2 | 5.00 | 0.440\% | 5.96 | 1.8 |
| 18.0 | 1.26\% | 3.73 | 1.3 | 10.0 | 0.355\% | 3.95 | 1.4 | 5,00 | 0.490\% | 9.32 | 2.2 |
| 23.5 | 3.25\% | 30.2 | 3.4 | 13.5 | 1.28\% | 16.6 | 2.8 | 7.00 | 0.410\% | 4.17 | 1.4 |
| 21.5 | 0.975\% | 9.23 | 2.2 | 14.0 | 1.12\% | 5.95 | 1,8 | 8.00 | 0.595\% | 3.73 | 1.3 |
| 22.5 | 3.16\% | 22.7 | 3.1 | 13.0 | 0.840\% | 7.52 | 2.0 | 7.00 | 0.510\% | 9.78 | 2.3 |
| 18.0 | 0.955\% | 3,55 | 1.3 |  |  |  |  |  |  |  |  |
| 15.0 | 0.315\% | 1.71 | 0.5 |  |  |  |  |  |  |  |  |
| 25.0 | 10.0\% | 47.7 | 3.9 |  |  |  |  |  |  |  |  |
| 20.5 | 1.06\% | 10.5 | 2.4 |  |  |  |  |  |  |  |  |
| 20.0 | 0.290\% | 1.02 | 0.0 |  |  |  |  |  |  |  |  |
| 23.0 | 2.06\% | 15.8 | 2.8 |  |  |  |  |  |  |  |  |
| 17.5 | 0.405\% | 1.39 | 0.3 |  |  |  |  |  |  |  |  |
| 24.5 | 7.55\% | 37.3 | 3.6 |  |  |  |  |  |  |  |  |

Upper Rivar Fish Tissue Results - Site 2

| Smatimouth Bass Length | $\begin{aligned} & \text { Smallmouth Bess } \\ & \times \text { Lipid } \end{aligned}$ | $\begin{gathered} \text { Smailmouth Bats } \\ \text { PCB } \end{gathered}$ | Ln Smalmouth Bass PCB | Rock Bass Length | Rock Baes \% tipid | Rock Buns PCB | In Hect Bass PCD |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 11.0 | 1.78\% | 28.9 | 3.4 | 9.00 | 0.405\% | 1.04 | 0.0 |
| 13.0 | 0.775\% | 5.34 | 1.7 | 8.00 | 0.670\% | 4.24 | 1.4 |
| 11.0 | 1.16\% | 14.9 | 2.7 | 6.00 | 0.980\% | 8.25 | 2.1 |
| 12.0 | 1.67\% | 33.5 | 3.5 | 7.00 | 1.20\% | 8.72 | 2.2 |
| 13.0 | 1.26\% | 3.12 | 1.1 | 8.00 | 0.470\% | 4.32 | 1.5 |
| 10.0 | 0.970\% | 6.41 | 1.9 | 8.00 | 0.705\% | 3.78 | 1.3 |
| 10.0 | 1,69\% | 13.5 | 2.6 | 8.00 | 0.580\% | 3.04 | 1.1 |
| 10.0 | 1.29\% | 10.5 | 2.4 | 8.00 | 0.240\% | 0.739 | -0.3 |

## Middle River Fish Tissue Results - Site 1

| Adult Carp Length | Adutt Carp * Lipid | Adult Carp PCB | In Adult Carp PCB | Adult White Sucker Length | Adult White Sucker \% Lipid | Adult White Sucker PCB | In Adult Whte Sucker PCB | $\begin{gathered} \text { Smallmouth Bass } \\ \text { Length } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Smallmouth Bass } \\ \% \text { Lipid } \end{gathered}$ | 5malimouth Bass PCB | In Smallmouth Bass <br> PCB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 16.0 | 1.22\% | 2.06 | 0.7 | 16.0 | 0.870\% | 3.72 | 1.3 | 13.0 | 1.37\% | 14.1 | 2.6 |
| 16.0 | 0.770\% | 1.71 | 0.5 | 15.0 | 1.30\% | 11.8 | 2.5 | 15.0 | 2.27\% | 6.04 | 1.8 |
| 17.0 | 0.390\% | 1.33 | 0.3 | 10.0 | 0.740x | 3.24 | 1.2 | 14.0 | 1.09\% | 5.77 | 1.8 |
| 17.0 | 3.21\% | 2.51 | 0.9 | 16.0 | 0.795\% | 19.9 | 3.0 | 14.0 | 0.815\% | 4.20 | 1.4 |
| 15.5 | 0.845\% | 1.62 | 0.5 | 16.0 | 1.50\% | 8.79 | 2.2 | 14.5 | 0.765\% | 7.46 | 2.0 |
| 16.0 | 1.17\% | 1.28 | 0.2 | 14.0 | 0.705\% | 4.68 | 1.5 | 12.0 | 0.680\% | 9.29 | 2.2 |
| 17.5 | 1.14\% | 2.21 | 0.8 | 16.0 | 1.01\% | 9.23 | 2.2 | 15.0 | 1.30\% | 18.2 | 2.9 |
| 20.5 | 3.16\% | 22.8 | 3.1 |  |  |  |  | 11.0 | 0.830\% | 4.97 | 1.6 |

Middle River Fish Tissue Results - Site 1

| Rock Bass Length | Rack Bass \% Lipid | $\begin{aligned} & \text { Rock Bass } \\ & \text { PCB } \end{aligned}$ | In Rock Bass PCB | Longnose Dace length | Longnose Dace * Lipid | $\begin{gathered} \hline \text { Longnose Dace } \\ \text { PCB } \\ \hline \hline \end{gathered}$ | In longnose Dace PCB | Channel Catilish Length | Channel Catfigh \% lipid | Channel Cattish PCB | Ln Channel Catfish PCB | Walleye Length | Wallieye \% Lipid | Walleye PCB | Walleye PCB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7.00 | 0.810\% | 2.79 | 1.0 | 4.00 | 5.82\% | 17.8 | 2.9 | 21.0 | 4.02\% | 15.9 | 2.8 | 21.0 | 2.33\% | 16.8 | 2.8 |
|  |  |  |  | 3.50 | 2.08\% | 8.35 | 2.1 | 22.0 | 12.6\% | 49.2 | 3.9 | 19.5 | 2.11\% | 16.3 | 2.8 |
|  |  |  |  | 2.00 | 3.64\% | 8.92 | 2.2 | 19.0 | 6.34\% | 29.8 | 3.4 | 12.5 | 0.595\% | 5.58 | 1.7 |
|  |  |  |  | 2.50 | 4.84\% | 7.08 | 2.0 | 20.0 | 5.27\% | 16.6 | 2.8 | 16.0 | 1.52\% | 13.7 | 2.6 |
|  |  |  |  | 2.00 | 2.70\% | 7.10 | 2.0 |  |  |  |  | 16.0 | 0.695\% | 7.93 | 2.1 |
|  |  |  |  | 2.00 | 3.09\% | 7.56 | 2.0 |  |  |  |  | 17.5 | 1.61\% | 14.3 | 2.7 |
|  |  |  |  |  |  |  |  |  |  |  |  | 13.0 | 0.465\% | 6.03 | 1.8 |
|  |  |  |  |  |  |  |  |  |  |  |  | 15.5 | 1.00\% | 8.41 | 2.1 |


| Adult Carp Length | Adult Carp * Llipid | Adult Carp PCB | Adult White Sucker Length | Adult White Suckar \% Lipid | Adult Whita Sucker PCP | In Adult White Sucker PCB | Juvanlle White Suckar Length | Juvenile White Sucker \% Lipid | Jovenile White Surker <br> PCB | In Juwanila Whita Suckar PCB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19.0 | 0.730\% | 1.27 | 14.5 | 0.200\% | 3.24 | 1.2 | 8.00 | 0.480\% | 2.03 | 0.7 |
|  |  |  | 14.5 | 0.170\% | 2.37 | 0.9 | 8.00 | 0.400\% | 1.20 | 0.2 |
|  |  |  | 14.0 | 0.520\% | 3.51 | 1.3 | 8.00 | 0.740\% | 1.76 | 0.6 |
|  |  |  | 16.0 | 0.715\% | 3.48 | 1.2 | B.00 | 0.575\% | 1.13 | 0.1 |
|  |  |  | 14.0 | 0.150\% | 0.925 | -0.1 | B. 00 | 0.270\% | 0.980 | 0.0 |
|  |  |  | 16.0 | 1.23\% | 6.36 | 19 | 8.00 | 0.557\% | 1.08 | 0.1 |
|  |  |  | 15.0 | 0.585\% | 5.98 | 1.9 | 6.00 | 0.455\% | 1.40 | 0.3 |
|  |  |  | 13.5 | 1.36\% | 4.83 | 1.6 |  |  |  |  |

Middie Rwer Fish Tissue Results - She 2

| Smallmouth Bans Lengrh | Smallmouth Bass * Lipid | $\begin{gathered} \text { Smallmouth Basi } \\ \hline \end{gathered}$ | In smalifmouth Bass PCB | RoEk Bass Length | $\begin{gathered} \text { Rock Bass } \% \\ \text { Lipidd } \end{gathered}$ | Rock Bass PC: | In Rock Bass PCB | Longnose Dice length | $\begin{gathered} \text { Longnosa Dace } \\ \text { \% Lipld } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Longnose Dace } \\ \hline \text { PCB } \\ \hline \end{gathered}$ | In Lonemose Dace PCB | Channal Catilish Length | $\begin{gathered} \hline \text { Channal Catfish } \\ \times \text { Lipid } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Ohannel Cartish } \\ \text { PCB } \\ \hline \end{gathered}$ | Channel Catish |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 17.0 | 0.875\% | 3.53 | 1.3 | 7.00 | 0.480\% | 1.42 | 0.4 | 3.50 | 2.84\% | 6.20 | 1.8 | 19.0 | 4.21\% | 6.90 | 1.9 |
| 14.5 | 1.09\% | 7.65 | 2.0 | 7.00 | 0.593\% | 2.09 | 0.7 | 3.50 | 5.02\% | 9.60 | 2.3 | 22.0 | 6.01* | 8.69 | 2.2 |
| 12.0 | 2.00\% | 5.54 | 1.7 | 7.00 | 1.24\% | 1.88 | 0.6 | 3.50 | 6.09\% | 10.9 | 2.4 | 22.0 | 3,45\% | 16.6 | 2.8 |
| 11.0 | 1.06\% | 254 | 1.0 | 6.50 | 1.80\% | 3.47 | 1.2 | 4.00 | 5.50\% | 11.0 | 2.4 | 17.0 | 3.49* | 0.532 | -0.6 |
| 11.5 | 1.12\% | 3.65 | 1.3 | 5.50 | 1.02\% | 286 | 1.1 | 2.00 | 2.33\% | 4.86 | 1.6 |  |  |  |  |
| 11.0 | 1.09\% | 3.08 | 1.1 | 6.00 | 1.30\% | 3.70 | 1.3 | 2.50 | 5.09\% | 7.17 | 2.0 |  |  |  |  |
| 10.0 | 1.30\% | 4.28 | 15 | 6,00 | 0.583\% | 2.27 | 0.8 | 3.50 | 4.13\% | 9.86 | 2.3 |  |  |  |  |
| 12.0 | 1.26\% | 4.05 | 1.4 | 8.00 | 0.495\% | 2.20 | 0.8 | 3.00 | 5.74\% | 8.47 | 2.1 |  |  |  |  |

## Lower Rivar Fish Tissue Resulte

| Adult Corp Length | Adult Carp \% lipid | Adult Carp PCB | Ln Adult Carp PCB | Adult White Sucker Length | Adult White Sucker * Lipid | Adult White Sucker PCB | Jivenile White Sucker length | Juvarile White Sucker \% Lipid | Juvenlite White Sucker PCB | In Juvenile White Sucker PCB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 17.5 | 2.455\% | 2.52 | 0.9 | 12.5 | 1.025\% | 4.96 | 7.00 | 0.140\% | 1,27 | 0.2 |
| 24.5 | 2.69\% | 15.7 | 2.8 | 13.5 | 0.705\% | 3.65 | 8.00 | 0.205\% | 1.64 | 0.5 |
| 21.0 | 5.51\% | 0.458 | -0.8 |  |  |  | 6.50 | 0.245\% | 0.713 | -0.3 |
| 17.5 | 9.03\% | 44.9 | 3,8 |  |  |  | 5.00 | 0.094\% | 0.587 | -0.5 |
| 24.0 | 5.40\% | 18.4 | 2.9 |  |  |  | 7.00 | 0.405\% | 0.967 | 0.0 |
| 24.0 | 3.63\% | 4.46 | 1.5 |  |  |  |  |  |  |  |
| 18.0 | 0.825\% | 1.97 | 0.7 |  |  |  |  |  |  |  |
| 19.5 | 1.07\% | 1.89 | 0.6 |  |  |  |  |  |  |  |

Lower River Flyh Tlssue Resuits

| Smallmouth Buss Length | Smallmouth Bass \% Lipid | Smallmouth Bass PCB | In Smallmouth Bana PCE | Rock Buess Length | Rock Banis $\%$ Lupid | Rock Bass PCB | Ln Rock Bass PCB | Channel Catfish Length | Channel Catfish \% Lipid | $\begin{gathered} \text { Channal Catfilh } \\ \text { PCB } \end{gathered}$ | Channel Cortish |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10.0 | 1.19\% | 8.17 | 2.1 | 7.00 | 0.510\% | 1.76 | 0.6 | 19.0 | 4.11\% | 8.49 | 2.1 |
| 10.5 | 0.380\% | 5.14 | 1.6 | 6.50 | 0.410\% | 1.95 | 0.7 | 21.0 | 4.34\% | 11.7 | 2.5 |
| 13.0 | 0.650\% | 2.02 | 0.7 | 5,50 | 0.283\% | 1.40 | 0.3 | 20.0 | 4.98\% | 6.37 | 1.9 |
| 10.0 | 0.685\% | 1.78 | 0.6 | 5.00 | 0.982\% | 4.11 | 1.4 | 17.0 | 7.81\% | 28.4 | 3.3 |
| 12.0 | 1.50\% | 7.01 | 1.9 | 6.50 | 0.980\% | 3,33 | 1.2 |  |  |  |  |
| 11.0 | 0.915\% | 4.84 | 1.6 | 6.50 | 0.445\% | 1.84 | 0.6 |  |  |  |  |
| 12.0 | 2.13\% | 10.9 | 2.4 | 6.00 | 0.393\% | 1.63 | 0.5 |  |  |  |  |
| 10.5 | 1.05\% | 6.30 | 1.8 | 7.00 | 0.915\% | 4.27 | 1.5 |  |  |  |  |
|  |  |  |  | 5.50 | 0.300\% | 3.07 | 1.1 |  |  |  |  |


| Adult Carp Length | Adult Carp <br> * Lipid | $\begin{aligned} & \text { Adult Carp } \\ & \text { PCB } \\ & \hline \end{aligned}$ | In Adult Carp PCB | Smallmouth Bass Length | Smallmouth Rass \% Lipid | Smallmouth Bass PCB | In Smallmouth Bass PCB | Channel Catifish length | Channel Catfish \% Lipid | Channel Catrish PCB | Walleye Length | Walleye * Lipid | Walleye PCB | Walleye PCB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 21.0 | 3.83\% | 9.14 | 2.2 | 15.0 | 0.680\% | 1.44 | 0.4 | 20.5 | 12.2\% | 19.4 | 21.0 | 3.71\% | 3.00 | 1.1 |
| 23.0 | 1.91\% | 3.21 | 1.2 | 14.0 | 0.855\% | 2.70 | 1.0 |  |  |  | 21.0 | 2.71\% | 1.36 | 0.3 |
| 16.5 | 2.52\% | 2.46 | 0.9 | 12.0 | 0.935\% | 4.43 | 1.5 |  |  |  | 22.0 | 1.72\% | 1.74 | 0.6 |
| 17.0 | 3.03\% | 5.02 | 1.6 | 13.0 | 1.00\% | 3.10 | 1.1 |  |  |  |  |  |  |  |
| 18.5 | 4.04\% | 2.30 | 0.8 | 11.5 | 0.980\% | 4.18 | 1.4 |  |  |  |  |  |  |  |
| 16.5 | 4.06\% | 2.05 | 0.7 | 11.0 | 1.13\% | 4.31 | 1.5 |  |  |  |  |  |  |  |
| 18.5 | 1.29\% | 0.890 | -0.1 | 14.0 | 1.58\% | 3.91 | 1.4 |  |  |  |  |  |  |  |
| 19.0 | 0.630\% | 0.243 | -1.4 | 17.0 | 1.77\% | 2.83 | 1.0 |  |  |  |  |  |  |  |

## Attachment 2

## Regression Analysis Summary

Table 4
Two-Varlable Regreasion Results

| Reach | Statistic | Adult Carp |  | Adult Suckers |  | Juvenile Suckeri |  | Smallmouth Bass |  | Rock Bas! |  | Longnose Dace |  | Walleye |  | Catfish |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Length | Luplds | Length | Lplds | Length | Lupldı | Length | Lupids | Length | Liplds | Length | Lplda | Length | Lupldt | Length | Lpids |
| Upper River 1 | N | 16 |  | 8 |  | 8 |  | 8 |  | 8 |  | 6 |  | 0 |  | 0 |  |
|  | $\mathrm{R}^{2}$ | 0.39 |  | 0.91 |  | 0.02 |  | 0.20 |  | 0.53 |  | 0.85 |  | - |  | - |  |
|  | Coefficient | 0.195 | 4.59 | -0.485 | 1083.66 | -0.036 | 57.48 | 3.137 | -1911.86 | -0.876 | 1344.79 | 1.503 | 8.67 | - | - | - | - |
|  | Standard Error | 0.093 | 8.24 | 0.458 | 150.04 | 0.490 | 184.79 | 3.008 | 1690.26 | 1.548 | 870.67 | 0.507 | 19.21 | - | - | - | - |
|  | $p$ (2-Lail) | 0.056 | 0.587 | 0.338 | 0.001 | 0.944 | 0.768 | 0.345 | 0.309 | 0.596 | 0.183 | 0.059 | 0.682 | - | - | - | - |
|  | Model | Exponential |  | Linear |  | Exponential |  | Linear |  | Linear |  | Exponential |  | - |  | - |  |
| Upper River 2 | N | 16 |  | 8 |  | 8 |  | 8 |  | 8 |  | 0 |  | 0 |  | 0 |  |
|  | $\mathrm{R}^{2}$ | 0.88 |  | 0.69 |  | 0.09 |  | 0.59 |  | 0.95 |  | - |  | - |  | - |  |
|  | Coefficient | 1.925 | 341.85 | -0.224 | 191.41 | -0.072 | -136.27 | 1.033 | 2442.65 | -1.153 | 645.16 | - | - | - | - | - | - |
|  | Slandard Error | 0.517 | 55.57 | 0.096 | 57.33 | 0.180 | 291.98 | 2.564 | 910.63 | 0.514 | 146.26 | - | - | - | - | - | - |
|  | p (2-tail) | 0.003 | 0.000 | 0.067 | 0.021 | 0.707 | 0.660 | 0.704 | 0.044 | 0075 | 0.007 | - | . | - | - | - | - |
|  | Model | Linear |  | Exponential |  | Exponential |  | Linear |  | Linear |  | - |  | - |  | - |  |
| Middle River 1 | N | 8 |  | 7 |  | 0 |  | 8 |  | 1 |  | 6 |  | 8 |  | 4 |  |
|  | $\mathrm{R}^{2}$ | 0.88 |  | 0.37 |  | - |  | 0.05 |  | - |  | 0.77 |  | 0.96 |  | 0.96 |  |
|  | Coefficient | 0.445 | 20.95 | 0.159 | 33.50 | - | - | 0.655 | 50.73 | - | - | 2.716 | 140.90 | 0.017 | 635.96 | -1.850 | 436.13 |
|  | Slandard Error | 0.123 | 18.07 | 0.131 | 96.16 | - | - | 1.731 | 481.56 | - | - | 1.428 | 89.31 | 0.399 | 165.11 | 3.011 | 102.35 |
|  | p (2-(ail) | 0.015 | 0.299 | 0.294 | 0.745 | - | - | 0.721 | 0.920 | - | - | 0.153 | 0.213 | 0.967 | 0.012 | 0.649 | 0.147 |
|  | Model | Exponential |  | Exponential |  | - |  | Linear |  | - |  | Linear |  | Linear |  | Linear |  |
| Middle River 2 | N | 1 |  | 8 |  | 7 |  | 8 |  | 8 |  | 8 |  | 0 |  | 4 |  |
|  | $\mathrm{R}^{2}$ | - |  | 0.53 |  | 0.25 |  | 0.37 |  | 0.62 |  | 0.91 |  | . |  | 0.97 |  |
|  | Coefficient | - | - | 0.615 | 265.92 | -0.054 | 89.09 | 0.072 | 57.54 | -0.299 | 102.52 | 0.238 | 12.56 | - | . | 3.040 | -298.64 |
|  | Standard Error | - | - | 0.680 | 134.30 | 0.155 | 79.19 | 0.056 | 37.53 | 0.297 | 48.80 | 0.068 | 3.23 | - | - | 0.491 | 100.24 |
|  | p(2-tail) | - | - | 0.407 | 0.105 | 0.746 | 0.324 | 0.255 | 0.186 | 0.360 | 0.090 | 0.017 | 0.012 | - | - | 0.102 | 0.206 |
|  | Model | - |  | Linear |  | Exponential |  | Exponential |  | Linear |  | Exponential |  | - |  | Linear |  |
| Lower River | N | 8 |  | 2 |  | 5 |  | 8 |  | 9 |  | 0 |  | 0 |  | 4 |  |
|  | $\mathrm{R}^{2}$ | 0.64 |  | - |  | 0.93 |  | 0.76 |  | 0.67 |  | - |  | - |  | 0.86 |  |
|  | Coefficient | -0.639 | 425.64 | - | - | 0.414 | -145.70 | -0.710 | 506.71 | 0.022 | 311.03 | - | - | - | - | -0.508 | 499.45 |
|  | SLandard Error | 1.341 | 144.76 | - | - | 0.083 | 75.92 | 0.635 | 125.95 | 0.392 | 89.06 | - | - | - | - | 4.001 | 400.02 |
|  | p (2-tail) | 0.654 | 0.032 | - | - | 0.038 | 0.195 | 0.315 | 0.010 | 0.957 | 0.013 | - | - | - | - | 0.920 | 0.430 |
|  | Model | Linear |  | - |  | Linear |  | Linear |  | Linear |  | - |  | - |  | Linear |  |
| Inner Harbor | N | 8 |  | 0 |  | 0 |  | 8 |  | 0 |  | 0 |  | 3 |  | 1 |  |
|  | $\mathrm{R}^{2}$ | 0.64 |  | . |  | - |  | 0.90 |  | - |  | - |  | - |  | - |  |
|  | Coefficient | 0.165 | 67.67 | - | - | - | - | -0.533 | 196.02 | - | - | - | - | - | - | - | - |
|  | Standard Error | 0.133 | 23.19 | - | - | - | - | 0.081 | 43.76 | - | - | - | - | - | - | - | - |
|  | p (2-tail) | 0.269 | 0.033 | - | - | - | - | 0.001 | 0.007 | - | - | - | - | - | - | - | - |
|  | Model | Exponential |  | - |  | - |  | Linear |  | - |  | - |  | - |  | - |  |

Note: p-Level indicates the probability of the coeffiecient being equal to zero. Lower value of p indicate higher probabilities
that the factors of length or percent lipids significantly affeet fish tissue PCB concentrations.



[^0]:    ${ }^{1}$ The Upper River has already been remediated. The first annual event will be used as the baseline event.

[^1]:    ${ }^{2}$ The Near-Shore sediments are defined as sediment segments that may be found in the bank or river bed adjacent to the shoreline of the former Tecumseh plant, along the north side of the Sheboygan River as described in the External Source Assessment (ESA). Armored Areas were portions of the river bed that had been covered with a geotextile fabric, a one-foot layer of run-of-bank material, another layer of geotextile fabric, gabions (cages filled with larger stone pieces or cobbles) along the sediment periphery, and cobbles to fill in any gaps between the gabions and atop the fabric (i.e. armoring) to stabilize the river bed and prevent a release of contaminated sediments into the river.
    ) Soft Sediments are defined as the sediment found on the river bed as a result of the river deposited suspended material where sediment was measured greater than 1 foot thick during the 2004 pre-design investigation.

[^2]:    ${ }^{3}$ Much of the information presented in this section was obtained from the ROD.

[^3]:    ${ }^{4}$ Most recent WNDR data available was used. This ranged from 1990 (Inner Harbor) 2000 to 2004 (others), depending on species and reach.
    ${ }^{5}$ Conducted in 1994, 1995, 1996, 1998, 1999, 2000, 2001, and 2002.

[^4]:    ${ }^{6}$ This is a migratory fish species and most PCB burden is from Lake Michigan.
    ${ }^{7}$ There could be a lag period as older fish may have PCB concentrations reflective of when the sediment was more impacted.

[^5]:    ${ }^{8}$ Species was determined by SOP \#10, Fish Identification, and with assistance from CH2MHill.

[^6]:    ${ }^{9}$ Historical results were provided by the USEPA and WDNR. These included the BBL Interim Monitoring Program data and WDNR fish advisory studies. The data was provided in the Post Remedial Monitoring Plan.
    ${ }^{10}$ The significance level is the risk of a false rejection.

[^7]:    ${ }^{11}$ Historical results were provided by the USEPA and WDNR. These included the BBL Interim Monitoring Program data and WDNR fish advisory studies. The data was provided in the Post Remedial Monitoring Plan.
    ${ }^{12}$ The mean results were used.
    ${ }^{13}$ Fish were not monitored.

[^8]:    ${ }^{14}$ Each data set represented one fish species in one site or reach.

[^9]:    ${ }^{15}$ Where not provided in Fishes of Wisconsin, lifespan were obtained from various internet sources.

[^10]:    ${ }^{16}$ Units and other information provided on last page of table.

